

1 **Impact of hydrothermal treatment of FEBEX and MX80**
2 **bentonites in water, HNO₃ and Lu(NO₃)₃ media: Implications**
3 **for radioactive waste control**

4 **Francisco J. Osuna¹, Pablo Chain², Agustín Cota³,**
5 **Esperanza Pavón^{1,4}, María D. Alba^{1*}**

6
7 ¹Instituto Ciencia de los Materiales de Sevilla- Departamento de Química Inorgánica, CSIC-
8 Universidad de Sevilla. Avda. Américo Vespucio, 49. 41092, Sevilla. Spain

9 ²Consejería de Educación. Junta de Andalucía. 41092, Sevilla. Spain

10 ³Laboratorio de Rayos-X. Centro de Investigación, Tecnología e Innovación de la Universidad de
11 Sevilla (CITIUS), Universidad de Sevilla. Avda. Reina Mercedes, 4, 41012 Sevilla, Spain.

12 ⁴Center for the Development of Nanoscience and Nanotechnology, CEDENNA, 9170124 Santiago,
13 Chile

14
15 **ABSTRACT**

16 Engineered barriers of deep geological repositories (DGR) are commonly
17 constructed with bentonite. FEBEX and MX80 bentonites have been selected by different
18 countries as reference materials for the sealing of repositories; however, their chemical
19 reactivity with high-level long-lived radioactive wastes (HLRW) under subcritical
20 conditions had not been explored before. The hydrothermal stability in neutral and acid
21 media and chemical reactivity in contact with an actinide analogous compound were both
22 studied. The long-range and short-range structural changes were analyzed by X-ray
23 diffraction, nuclear magnetic resonance and scanning electron microscopy. Both bentonites

*Corresponding author. Tel.: +34 954489546
E-mail address: alba@icmse.csic.es

24 have exhibited a good stability in neutral and acid media and have generated a new phase
25 immobilizing the actinide analogous compound. The extent of the chemical reaction is
26 higher in MX80 bentonite than in FEBEX bentonite.

27

28 **Key Words**—Bentonite, Engineered barrier, rare-earth, hydrothermal treatment, subcritical
29 conditions, radioactive waste.

30

31 **1. Introduction**

32

33 The complexity of radioactive waste management is one of the most significant
34 challenges faced by all countries involved in nuclear power generation. It is also a concern
35 in many other countries using radioactive materials for medical, industrial, or research
36 purposes. Radioactive wastes have to be disposed in ways that minimize their negative
37 impact. In particular, the most hazardous and long-lived waste, such as spent nuclear fuel or
38 reprocessing waste arising from the operation and dismantling of nuclear reactors, must be
39 contained and isolated in order to ensure the long-term radiological protection of humans
40 and the environment. Disposal of such waste in deep underground-geological repositories
41 has become an effective and universally preferred option (Allègre, 1999; McCombie et al.,
42 2000; Savage et al. 2010).

43 A deep geological repository (DGR), designed for the long-term management of
44 high-level long-lived radioactive waste (HLRW), is a multi-barrier system consisting of
45 engineered and backfill barriers surrounded by an additional geosphere barrier (Astudillo,
46 2001; Chapman, 2006; McCombie et al., 2000). Engineered barriers are commonly
47 constructed with bentonite due to its low hydraulic conductivity and diffusivity and its high
48 cation sorption capacity and buffering properties (Landais, 2006; Meunier et al., 1998;
49 Pusch, 2006). In addition, bentonite retards the diffusion of radionuclides through sorption
50 processes, and it is able to react with radioactive leachates (Bradbury and Baeyens, 2005;
51 Coppin et al., 2002). The barrier, made of compacted bentonite (a highly swelling material),
52 will get saturated with groundwater while it is subjected to high temperatures due to the
53 radioactive decay of the wastes (Villar and Gómez-Espina 2008).

54 Under the conditions expected in HLRW repositories, alteration of a smectite (the
55 main mineral component of bentonites) is expected to increase its hydraulic conductivity
56 (permeability) and porosity and decrease its swelling pressure and cation exchange capacity
57 (CEC) (Bildstein et al., 2006; Carlson et al., 2007; Gaudin et al., 2009; Savage et al., 2010).
58 An alternative and effective immobilization mechanism that would overcome these
59 drawbacks could involve the chemical retention of actinides by the bentonite. Lutetium
60 disilicate ($\text{Lu}_2\text{Si}_2\text{O}_7$) is formed when a nitric acid solution of lutetium reacts with pure
61 smectite under subcritical pressures and temperatures as it is expected to occur in Deep
62 Geological Repositories (DGR), even after weathering of the smectite (Alba and Chain,
63 2005; Alba et al., 2009, Perdigón, 2002). Moreover, the persistence of $\text{Lu}_2\text{Si}_2\text{O}_7$ under
64 hydrothermal conditions covering a wide range of pH and different salt solutions was
65 demonstrated (Alba et al., 2009). Disilicate-phase synthesis was, therefore, proposed to be a
66 possible long-term retention solution for trivalent actinide cations. In this sense, Chain et al.
67 (2013) analyzed the structural properties of smectites that enhance their reactivity with Lu^{3+}
68 ions and lead to formation of the disilicate phase. However, the international HLRW
69 programs do not consider using purified smectite but raw bentonite as buffer in DGR for
70 both logistic and economic reasons. In this sense, FEBEX and MX80 bentonites were
71 selected by different countries as reference materials for the sealing of HLRW repositories
72 (Tripathy et al., 2004; Villar and Gómez-Espina, 2008), but their capacity for
73 immobilization of actinides by chemical reaction under subcritical condition need further
74 investigations.

75 Both the chemical reaction and interaction of bentonites with groundwater can,
76 however, modify their physical and chemical properties, thus decreasing their retention
77 capacity (Pusch et al., 2007). The hydrothermal stability of clay minerals in water, basic

78 and acidic media is of great importance, particularly where radioactive waste confinement
79 is concerned. Thus, a precise knowledge of the properties of clay minerals such as the long-
80 term integrity of the bentonite when it is in contact with either acidic tailings pore water (in
81 the case of slurry wall filter cake) or acidic stored mine water (in the case of acid mine
82 drainage collection ponds), is, therefore, of crucial importance for the safety of radioactive
83 waste disposal concepts.

84 In a previous study, Villa-Alfageme et al. (2014) studied the reactivity of bentonite
85 with Eu^{3+} . They observed that Eu^{3+} is immobilized in a higher amount than expected based
86 on the CEC of the bentonite and concluded that an additional chemical mechanism may
87 therefore participate to the immobilization. Since short-range study by NMR was not
88 carried out, the extension and nature of the involved chemical reaction(s) were not
89 evaluated. The main aim of this study is, therefore, to characterize the formation of
90 $\text{Lu}_2\text{Si}_2\text{O}_7$ in the bentonites FEBEX and MX80 upon treatment with a Lu^{3+} solution under
91 subcritical conditions, with the aim to study the structural changes occurred on the
92 bentonite and quantify the disilicate phase formation, and the hydrothermal stability of the
93 two bentonites in two media in which the reaction may take place, namely water and strong
94 acid. Although Lu^{3+} is not the best rare earth element (REE) analogous compound for
95 trivalent actinides most commonly found in high activity radioactive waste (Np, Am, and
96 Cm) (Astudillo, 2001), it was chosen because the formation of rare-earth disilicates is
97 common to all REE, with lutetium being the most reactive (Alba et al., 2009).

98

99 **2. Materials and method**

100

101 *2.1. Materials*

102

103 The two selected bentonites are the Spanish bentonite FEBEX and the American
104 bentonite MX80. Those bentonites were provided by ENRESA Company (the Spanish
105 Company in charge of radioactive wastes management) and CIEMAT, Spain.

106 Bentonite FEBEX is a Ca, Mg, Na-bentonite (predominantly Ca) extracted from the
107 Cortijo de Archidona deposit (Serrata de Nijar, Almería, Spain). It contains more than 90 %
108 w/w of montmorillonite and variable quantities of quartz, plagioclase, K-feldspar, calcite
109 and cristobalite (Caballero et al., 1983, 2005; Lloret, 2004; Villar 2002). The structural
110 formula of the montmorillonite component is $(Ca_{0.5}Na_{0.08}K_{0.11})(Si_{7.78}Al_{0.22})$
111 $(Al_{2.78}Fe^{3+}_{0.33}Fe^{2+}_{0.02}Mg_{0.81})O_{20}(OH)_4$ and the bulk cation exchange capacity (CEC) of the
112 bentonite is 158.2 meq/100 g (Galunin et al., 2011).

113 Bentonite MX80 is a Na-bentonite from Wyoming, USA. This bentonite consists
114 mainly of montmorillonite (70–85 % w/w), but also contains quartz, feldspars and smaller
115 quantities of cristobalite and calcite (Monts-H et al., 2005). The structural formula of the
116 montmorillonite component is $(Ca_{0.2}Na_{0.36})(Si_{7.96}Al_{0.04})(Al_{3.1}Fe^{3+}_{0.18}Fe^{2+}_{0.16}Mg_{0.56})O_{20}(OH)_4$
117 and the bulk CEC of the bentonite is 102.1 meq/100 g (Galunin et al., 2011).

118

119 *2.2. Hydrothermal treatments*

120

121 300 mg of powdered bentonite and 50 mL of $7.3 \cdot 10^{-2}$ M $Lu(NO_3)_3 \cdot 3.6H_2O$ solution,
122 distilled water or 0.01 M HNO_3 solution were transferred into a stainless steel T316SS
123 hydrothermal reactor and heated to 300° C for 48 hours under autogenous pressure.
124 Although geochemical waste degradation and waste/rock interaction processes in
125 hydrothermal environments remain predictable up to temperatures of about 200 °C, many

126 studies focused on simulating deep geological disposal conditions use temperatures up to
127 350 °C to increase the reaction rates (Alba and Chaín, 2007; Allen et al., 1988; Mather et
128 al., 1982; Savage and Chapman, 1982). The reactors were then cooled to room temperature
129 and the solids separated by filtration, washed repeatedly with distilled water, and allowed to
130 dry in air at 25 °C.

131

132 *2.3. Analytical methods*

133

134 X-ray powder diffraction patterns (XRD) were obtained using a Bruker D8I
135 instrument. This diffractometer is operated with a Ni-filtered Cu K α radiation, at 40 kV and
136 40 mA. Diffractograms were obtained from 3 to 70° 2 θ at a scan step of 0.05° 2 θ and with
137 a counting time of 3 s.

138 Scanning electron microscopy (SEM) observations of morphology and Energy
139 Dispersive X-Ray Spectroscopy (EDX) analyses of chemical composition were performed
140 at 20 kV with a JEOL JSM 5400 microscope , equipped with a LINK Pantafet probe
141 (Oxford Link ISIS) allowing chemical analyses using a Si/Li detector with Be window.

142 ^{29}Si and ^{27}Al solid state nuclear magnetic spectra (MAS NMR) were recorded with
143 a magnetic field of 9.36 T using a Bruker DRX400 spectrometer equipped with a
144 multinuclear probe. Powdered samples were packed in 4 mm zirconia rotors and spun at 12
145 kHz. ^{29}Si MAS NMR spectra were obtained at a frequency of 79.49 MHz using a $\pi/6$ pulse
146 width of 2.7 μs and a delay time of 60 s. ^{27}Al MAS NMR spectra were recorded at 104.26
147 MHz with a $\pi/20$ pulse width of 1.1 μs and delay time of 3 s. Chemical shifts are reported
148 in ppm with respect to tetramethylsilane for ^{29}Si and a 0.1 M solution of AlCl_3 for ^{27}Al . All

149 spectra were deconvoluted into single Gaussian-Lorentzian bands using a modified version
150 of the Bruker Winfit program (Massiot et al., 2002). The fitted parameters were intensity,
151 position, band-width, and Gaussian/Lorentzian ratio.

152

153 **3. Results and discussion**

154

155 *3.1. Hydrothermal stability of bentonites*

156

157 Structural changes in the long-range order of both bentonites after hydrothermal
158 treatments with water and 0.01M HNO₃ were analyzed by XRD (Fig. 1). The XRD patterns
159 of the raw bentonites (Fig. 1a and 1d) are consistent with the general (*hk*) and basal (*00l*)
160 reflections of a smectite, with minor reflections corresponding to impurities of quartz (PDF
161 04-007-0522), cristobalite (PDF 04-008-7742), and feldspar (PDF 04-017-1349). The *060*
162 reflections are observed at a distance between the reflection planes of 0.15 nm, typical of
163 dioctahedral 2:1 phyllosilicates (Alba et al. 2001a). The *00l* basal reflections corresponds
164 to a d_{001} value of 1.47 nm for bentonite FEBEX and 1.23 nm for bentonite MX80 and are
165 due to a water bilayer and monolayer, respectively, around the interlayer cations, mainly
166 Ca²⁺ in bentonite FEBEX and Na⁺ in bentonite MX80 (Alba et al. 2001a).

167 The positions of the *hk* reflections of both bentonites do not change upon
168 hydrothermal treatments (Fig. 1b, 1c, 1e and 1f), thus indicating that the basic structure of
169 the silicates remained intact and that no other crystalline phase was formed. However, their
170 widths and intensities changed after the hydrothermal treatments, especially in acidic
171 media, indicating the occurrence of structural disorder, also evidenced by the more
172 pronounced pattern background.

173 The position and intensity of the *00l* reflection of bentonite FEBEX does not
174 change after the hydrothermal treatments in water and acidic media. The hydrothermal
175 treatments did however cause the swelling of the bentonite MX80 layers up to a d_{001}
176 spacing of 1.42 nm in water medium and 1.52 nm in acidic medium. This swelling may be
177 due to the presence of cations with a higher hydration state. Similar results were observed
178 by Alba et al. (2001a, 2001b) after hydrothermal treatment where a leaching of divalent and
179 trivalent cations from the octahedral sheet with the replacement of Na^+ in the interlayer
180 space was observed.

181 The SEM micrographies (Fig. 2a-2c and Fig. 3a-3c) evidence the remaining
182 lamellar structure of the both bentonite after the hydrothermal treatments. The untreated
183 bentonites show EDX spectra (Fig. 2d and Fig. 3d) with the emission lines characteristic of
184 the elements found in their chemical composition. The hydrothermal treatment in water
185 (Fig. 2e and Fig. 3e) does not cause any change in their chemical composition, however,
186 the acid treatment (Fig. 2f and Fig. 3f) results in a decrease of the K, Ca and Na $K_{\alpha 1}$ lines
187 due to the replacement of the initial interlayer cations by those leached from the
188 framework. Finally, this treatment provokes a relative decrease of the intensity of the Si $K_{\alpha 1}$
189 line, in agreement with the observed decrease of the XRD reflections intensities of the
190 impurities.

191 Therefore, both bentonites show a good stability under hydrothermal treatment in
192 water and acidic media and the observed structural changes are more pronounced in the
193 acidic medium. Short-range structural effect of the acid treatment was therefore analyzed
194 by MAS NMR spectroscopy.

195 ^{29}Si MAS NMR spectra of the untreated bentonites (Fig. 4a and 4c) show two set of
196 signals, the main one is in the range of -100 to -80 ppm and is due to $\text{Q}^3(\text{mAl})$, typical

197 environment of smectite, and the second one is in the range of -100 to -115 ppm and is due
198 to $Q^4(\text{mAl})$ of the impurities (Engelhardt and Michel, 1987). The deconvolution parameters
199 of the ^{29}Si MAS NMR spectra (Table 1) show that the smectite Si environments are
200 different in the two bentonites. For FEBEX bentonite, the smectite contribution to the ^{29}Si
201 MAS NMR spectrum is reaching 94.5 %, with two signals at -93.33 ppm, $Q^3(0\text{Al})$, and -
202 88.05 ppm, $Q^3(1\text{Al})$, which is compatible with an octahedral configuration and with an
203 isomorphical substitution of Si by Al in the tetrahedral sheet (Alba et al., 2001b). The ^{29}Si
204 MAS NMR spectrum of MX80 shows a symmetric signal at -93.80 ppm due to $Q^3(0\text{Al})$,
205 reaching 79.7 % of the total Si of the spectrum. The shifts of the ^{29}Si resonance at lower
206 frequency is due to the lower total layer charge (Weiss et al., 1987) and the absence of
207 $Q^3(1\text{Al})$ environment is due to the lack of substitution of Si by Al in the tetrahedral sheet.
208 In both bentonites, the ^{29}Si MAS NMR spectra are compatible with their mineralogical
209 composition (Caballero et al., 1983; Lloret et al., 2004; Montes-H et al., 2005).

210 After the hydrothermal treatment in acidic medium, the Q^4 component is less
211 intense in the ^{29}Si MAS NMR spectra of MX80 bentonite (Fig. 4d, left) revealing a
212 diminution in the amount of impurities. The $Q^3(0\text{Al})$ component of MX80 is similar to that
213 of the untreated bentonite. In the FEBEX sample, the $Q^3(1\text{Al})$ contribution is lost,
214 indicating the leaching of tetrahedral Al.

215 The ^{27}Al MAS NMR spectra of the untreated bentonites (Fig. 4a and c, right) show
216 a main resonance centered at ca. 0 ppm, due to the octahedral Al environment and
217 compatible with the dioctahedral character of both bentonites (Engelhardt and Michel,
218 1987). Additionally, a small signal between 50 and 79 ppm is observed, due to Al in
219 tetrahedral coordination (Engelhardt and Michel, 1987). In both samples, a signal at 40-70
220 ppm is made of two signals; the main one at ca. 57 ppm is due to q^4 environment of

221 impurities, and, the very small one at ca. 65 ppm is due to q^3 environment in smectite phase
222 (Engelhardt and Michel, 1987). After the hydrothermal treatments in acidic medium, the
223 spectra (Fig. 4b and d, right) are quite similar to those of the untreated bentonite, the only
224 difference being that the q^4 environment of the impurities is absent, further indicating their
225 dissolution.

226

227 *3.2. Immobilization of actinides analogous compounds by bentonites under*
228 *hydrothermal conditions*

229

230 Structural changes in the long-range order of the bentonites after hydrothermal
231 treatment with Lu^{3+} solution were analyzed by XRD (Fig. 5). In general, hk reflections of
232 the montmorillonite and quartz and feldspar impurities mostly disappeared after the
233 treatment, thus indicating disruption of their basic structure (or dissolution). Only a small
234 001 reflection is observed at a d_{001} of 1.49 nm which is caused by the partial swelling of the
235 layers, possibly due to the presence of trivalent cation Lu^{3+} in the interlayer space with a
236 higher hydration state than the original cations (Alba et al., 2001a). The intensity of the
237 basal reflections decreased, probably because of a lower-order stacking of the layers and
238 due to the partial transformation of the sample into kaolinite after the treatment.

239 The XRD patterns of the hydrothermally treated samples are dominated by the
240 kaolinite reflections (PDF 01-075-0938) and both bentonites present reflections
241 corresponding to $\text{Lu}_2\text{Si}_2\text{O}_7$ (PDF 00-035-0326), more intense for MX80 than for FEBEX.
242 Additionally, small reflections due to chrysotile (PDF 00-025-0645) are observed for
243 MX80.

244 Changes in short-range order produced by the hydrothermal treatment were studied
245 by MAS NMR spectroscopic analysis of the active nuclei in the silicate (Fig. 6). The
246 hydrothermal treatment of both bentonites resulted in ^{27}Al MAS NMR spectra (Fig. 6a and
247 c, right) with a unique aluminum symmetric resonance at ca. 0 ppm, due to six-coordinated
248 aluminum. The position at ca. 0 ppm and the low quadrupolar line shape of the band
249 (symmetry band) are compatible with aluminum in the octahedral sheet of kaolinite
250 (Engelhardt and Michel, 1987). The absence of both tetrahedral resonances is compatible
251 with the disruption of the smectite framework (or its dissolution) and with the dissolution
252 of the q^4 impurities of the initial samples.

253 After the treatment, a complete transformation of the ^{29}Si MAS NMR spectra is
254 observed (Fig. 6b and d, left). On the one hand, the signals at ca. -110 ppm, $Q^4(\text{mAl})$
255 environment, disappear in good agreement with the absence of quartz, cristobalite and
256 feldspar evidenced by XRD. On the other hand, the signal in the range of $Q^3(\text{mAl})$
257 environment shifted towards higher frequency, at ca. -91.7 ppm, and is the convolution of
258 two Si environment, the $Q^3(0\text{Al})$ of kaolinite (Engelhardt and Michel, 1987) and Q^1 of
259 $\text{Lu}_2\text{Si}_2\text{O}_7$ (Alba et al., 2001b). The similitude of both resonances does not allow a
260 quantification of the Si contribution of kaolinite versus $\text{Lu}_2\text{Si}_2\text{O}_7$. Therefore, an analysis of
261 the samples by SEM/EDX was carried out.

262 After the hydrothermal treatment in $\text{Lu}(\text{NO}_3)_3$ solution, the FEBEX sample shows a
263 homogeneous morphology composed by lamellar particles (Fig. 7a), similar to those of the
264 untreated bentonite (Fig. 2a). The chemical analyses by EDX (Fig. 7b-d) reveal that: (i) the
265 Mg, Na, K and Ca $K_{\alpha 1}$ lines disappear as a consequence of the smectite framework
266 disruption (or dissolution); (ii) the layers are mainly composed by Al and Si in a proportion
267 typical of kaolinite; and; (iii) in some particles, the EDX spectra contain also Lu M_{α} , L_{α} and

268 L_{β} lines. No other particle with different morphology and/or higher content of Lu was
269 detected. Therefore, the chemical reaction between FEBEX bentonite and Lu^{3+} is at its
270 initial stage.

271 After the hydrothermal treatment in $\text{Lu}(\text{NO}_3)_3$ solution, the MX80 sample shows
272 two types of particles: (i) most particles have a lamellar morphology (Fig. 8b) with an EDX
273 spectra (Fig. 8d) characterized by Si and Al $K_{\alpha 1}$ lines in a proportion typical of kaolinite;
274 and, (ii) other particles are isolated compact spheres (Fig. 8b) with an EDX spectra
275 constituted by Si $K_{\alpha 1}$ lines and Lu M_{α} , L_{α} and L_{β} lines which corresponds to $\text{Lu}_2\text{Si}_2\text{O}_7$
276 phases.

277 Therefore, the chemical reaction between the actinides analogous component and
278 bentonite is more extensive in MX80 than in FEBEX. This observation is in agreement
279 with a previous study which concluded that a lower isomorphic substitution and Na^+ as
280 interlayer cations favor the reaction of formation of $\text{Lu}_2\text{Si}_2\text{O}_7$ (Alba et al., 2009).

281

282 **4. Conclusions**

283

284 The lamellar structure of FEBEX and MX80 bentonites remains intact after
285 hydrothermal treatment in water and only minor structural changes are observed after the
286 hydrothermal treatment in acidic medium. The only observed effect is the swelling of the
287 interlayer space which is associated to the replacement of original monovalent and/or
288 divalent cations by trivalent cations leached from the framework and the dissolution of the
289 impurities (quartz and feldspar). However, longer experiments may lead to additional
290 textural, chemical and structural changes.

291 Both bentonites are able to generate $\text{Lu}_2\text{Si}_2\text{O}_7$ phase under hydrothermal treatment
292 at 300°C in $\text{Lu}(\text{NO}_3)_3$ solution simulating subcritical conditions. A large proportion of the
293 montmorillonite component of the bentonites is transformed into kaolinite, which is likely
294 to generate a substantial loss of CEC and may affect porosity and hydraulic conductivity of
295 the clay materials due to volume changes. The formation of the insoluble immobilization
296 phase $\text{Lu}_2\text{Si}_2\text{O}_7$ is however expected to minimize the impact of these possible changes by
297 providing an additional immobilization mechanism to the clay barrier. Among the two
298 bentonites, MX80 is the most efficient barrier material as a higher amount of insoluble
299 $\text{Lu}_2\text{Si}_2\text{O}_7$ phase was generated compared to FEBEX in spite of its smaller smectite content.
300 The lowest octahedral isomorphic substitution of the montmorillonite component of MX80
301 and the presence of Na^+ instead of Ca^{2+} in its interlayer space seem to favour the chemical
302 reaction of bentonite with actinide analogous compound.

303

304 **Acknowledgements**

305

306 We would like to thank the Junta de Andalucía (Spain) and FEDER (Proyecto de
307 Excelencia de la Junta de Andalucía, project P12-FQM-567) for providing financial
308 support. We also thank to ENRESA (contract nº 0079000237) and CIEMAT for providing
309 samples and by their support. Dr. Pavón also thanks her grant project, approved by the
310 Andalucía Talent Hub Program launched by the Andalusian Knowledge Agency, co-funded
311 by the European Union's Seventh Framework Program, Marie Skłodowska-Curie actions
312 (COFUND – Grant Agreement nº 291780) and the Ministry of Economy, Innovation,
313 Science and Employment of the Junta de Andalucía.

314

315 **References**

- 316 Alba, M.D., Becerro, A.I., Castro, M.A., Perdigón, A.C., 2001a. Hydrothermal reactivity of
317 Lu-saturated smectites: Part I. A long-range order study. *Am. Mineral.* 86, 115-123.
- 318 Alba, M.D., Becerro, A.I., Castro, M.A., Perdigón, A.C., 2001b. Hydrothermal reactivity
319 of Lu-saturated smectites: Part II. A short-range order study. *Am. Mineral.* 86, 124-
320 131.
- 321 Alba, M.D., Chain, P., 2005. Interaction between Lutetium Cations and 2:1
322 Aluminosilicates under Hydrothermal Treatment. *Clay Clay Miner.* 53, 39-46.
- 323 Alba, M.D., Chain, P., 2007. Persistence of lutetium disilicate. *App. Geoch.* 22, 192-201.
- 324 Alba, M.D., Chain, P., Orta, M.M., 2009. Chemical reactivity of argillaceous material in
325 engineered barrier Rare earth disilicate formation under subcritical conditions. *App.*
326 *Clay Sci.* 43, 369-375.
- 327 Allègre, M., 1999. Geological repository: an unavoidable and ethically correct solution. In:
328 Proceedings of the 24-th Annual International Symposium, London. Allen, C.C.,
329 Wood, M.I., 1988. Bentonite in Nuclear Waste Disposal: A Review of Research in
330 Support of the Basalt Waste Isolation Project. *App. Clay Sci.* 3, 11-30.
- 331 Astudillo, J., 2001. El almacenamiento geológico profundo de los residuos radiactivos de
332 alta actividad. Principios básicos y tecnología. *ENRESA*, Madrid.
- 333 Bildstein, O., Trotignon, L., Perronnet, M., Jullien, M., 2006. Modelling iron-clay
334 interactions in deep geological disposal conditions. *Phys. Chem. Earth* 31, 618-625.
- 335 Bradbury, M., Baeyens, B., 2005. Modelling the sorption of Mn(II), Co(II), Ni(II), Cd(II),
336 Eu(III), Am(III), Sn(IV), Th(IV), Np(V) and U(VI) on montmorillonites: linear free

337 energy relationships and estimates of surface binding constants for some selected
338 heavy metals and actinides. *Geochimica. Cosm. Acta* 69, 875–892.

339 Caballero, E., Fernandes Porto, M.J., Linares, J., Reyes, E., 1983. Mineralogy,
340 geochemistry and mineral genesis of bentonites from Serrata de Nijar (Almeria,
341 Spain), *Estudios Geológicos* 39, 121–140.

342 Caballero, E., Jiménez de Cisneros, C., Huertas, F.J., Huertas, F., Pozzuoli, A., Linares, J.,
343 2005. Bentonite from Cabo de Gata, Almería, Spain: a mineralogical and
344 geochemical overview. *Clay Min.* 40, 463–480.

345 Carlson, L., Karland, O., Oversby, V.M., Rance, A.P., Smart, N.R., Snellma, M.,
346 Vähänen, M., Werme, L.O., 2007. Experimental studies of the interactions between
347 anaerobically corroding iron and bentonite. *Phys. Chem. Earth* 32, 334–345.

348 Chain, P., Cota, A., El Mrabet, S., Pavón, E., Pazos, M.C., Alba, M.D., 2013. Evaluation
349 of rare earth on layered silicates under subcritical conditions: Effect of the
350 framework and interlayer space composition. *Chem. Geol.* 347, 208-216.

351 Chapman, N., 2006. Geological disposal of radioactive waste — concept, status and trends.
352 *J. Iber. Geol.* 32, 7–14.

353 Coppin, F., Berger, G., Bauer, A., Castet, S., Loubet, M., 2002. Sorption of lanthanides on
354 smectite and kaolinite. *Chem. Geol.* 182, 57–68.

355 Engelhardt, G., Michel, D., 1987. High-resolution Solid-State NMR of Silicates and
356 Zeolites. John Wiley & Sons, New York.

357 Galunin, E., Alba, M.D., Santos, M.J., Abrao, T., Vidal, M., 2011. Examination of
358 competitive lanthanide sorption onto smectites and its significance in the
359 management of radioactive waste. *J. Hazard. Mater.* 186, 1930-1941.

360 Gaudin, A., Gaboreau, S., Tinseau, E., Bartier, D., Petit, S., Grauby, O., Foct, F.,
361 Beaufort, D., 2009. Mineralogical reactions in the Tournemire argillite after in-situ
362 interaction with steels. *App. Clay Sci.* 43, 196–207.

363 Landais, P., 2006. Advances in geochemical research for the underground disposal of high-
364 level, long-lived radioactive waste in a clay formation. *J. Geochem. Expl.* 88, 32–
365 36.

366 Lloret, A., Romero, E., Villar, M.V., 2004. FEBEX II Project. Final report on thermo-
367 hydro-mechanical laboratory tests. *Publicación Técnica ENRESA 10/04*. Madrid, pp
368 180.

369 Massiot, D., Fayon, F., Capron, M., King, I., Le Calvé, S., Alonso, B., Durand, J.O., Bujoli,
370 B., Gan, Z., Hoatson, G., 2002. Modelling one- and two-dimensional solid-state
371 NMR spectra. *Mag. Reson. Chem.* 40, 70-76.

372 Mather, J.D., Chapman, N.A., Black, J.H., Lintern, B.C., 1982. The geological disposal of
373 high-level radioactive waste—a review of the Institute of Geological Sciences
374 Research Programme. *Nucl. Ener.* 21, 167-173.

375 McCombie, C., Pentz, D.L., Kurzeme, M., Miller, I., 2000. Deep geological repositories: a
376 safe and secure solution to disposal of nuclear wastes. *GeoEng2000 — An*
377 *International Conference on Geotechnical & Geological Engineering*, Melbourne,
378 Australia. Technomic, Lancaster.

379 Meunier, A., Velde, B., Griffault, L., 1998. The reactivity of bentonites: a review. An
380 application to clay barrier stability for nuclear waste storage. *Clay Clay Miner.* 33,
381 187–196.

382 Montes-H, G., Marty, N., Fritz, B., Clement, A., Michau, N., 2005. Modelling of long-term
383 diffusion-reaction in a bentonite barrier for radioactive waste confinement. *Appl.*
384 *Clay Sci.* 30, 181-198.

385 Perdigón, A.C., 2002. Estudio del sistema saponita/Lu(NO₃)₃/H₂O en condiciones
386 hidrotérmicas, Ph.D. thesis, University of Seville, Seville, Spain.

387 Pusch, R., 2006. Engineered barriers. In: Popov, V., Pusch, R. (Eds.), *Disposal of*
388 *Hazardous Waste in Underground Mines*. Wessex Institute of Technology, UK , pp.
389 35–40.

390 Pusch, R., Kasbohm, J., Pakovsky, J., Cechova, Z., 2007. Are all smectite clays suitable as
391 “buffers”. *Phys. Chem. Earth* 32, 116–122.

392 Savage, D., Chapman, N.A., 1982. Hydrothermal behaviour of simulated waste glass and
393 waste-rock interaction under repository conditions. *Chem. Geol.* 36, 59–86.

394 Savage, D., Watson, C., Benbow, S., Wilson, J., 2010. Modelling iron–bentonite
395 interaction. *App. Clay Sci.*, 47, 91–98.

396 Tripathy, S., Sriharan, A., Schanz, T., 2004. Swelling pressure of compacted bentonites
397 from diffuse double layer theory. *Can. Geot. J.*, 41, 437–450.

398 Villa-Alfageme, M., Hurtado, S., Castro, M.A., El Mrabet, S., Orta, M.M., Pazos, M.C.,
399 Alba, M.D., 2014. Quantification and comparison of the reaction properties of
400 FEBEX and MX-80 clays with saponite: Europium immobiliser under subcritical
401 conditions. *Appl. Clay Sci.*, 101, 10-15.

402 Villar, M.V., 2002. Thermo-hydro-mechanical characterisation of a bentonite from Cabo de
403 Gata. A study applied to the use of bentonite as sealing material in high level
404 radioactive waste repositories. *Publicación Técnica ENRESA 01/2002*, pp 258.
405 Madrid.

406 Villar, M.V., Gómez-Espina, R., 2008. Effect of temperature on the water retention
407 capacity of FEBEX and MX-80 bentonites. *Unsaturated Solids: Advances in Geo-*
408 *engineering*, 257-262.

409 Weiss, C.A., Altaner, S.P., Kirpatrick, R.J., 1987. High-resolution ^{29}Si NMR spectroscopy
410 of 2:1 layer silicates: Correlation among chemical shift, structural distortions, and
411 chemical variations. *Am. Miner.* 72, 935-942.

412

413 .

414

415

416

Table 1. ^{29}Si chemical shift, linewidth (FWHM) and area under the curve of the different contribution obtained from the fitting of the ^{29}Si MAS NMR spectra of bentonite FEBEX and MX80: starting materials, and, after hydrothermal treatment at 300 °C for 48 h in 0.01 M HNO_3 solution.

sample	δ (ppm)	FWWH (KHz)	%	Phase
Bentonite FEBEX				
raw	-88.05	0.819	18.4	$\text{Q}^3(1\text{Al})$
	-93.33	0.325	76.3	$\text{Q}^3(0\text{Al})$
	-110.34	0.661	5.3	$\text{Q}^4(0\text{Al})$
treated in acid	-93.54	0.437	100	$\text{Q}^3(0\text{Al})$
Bentonite MX80				
raw	-93.80	0.758	79.7	$\text{Q}^3(0\text{Al})$
	-107.85	0.077	2.4	$\text{Q}^4(0\text{Al})$
	-109.52	0.296	17.9	$\text{Q}^4(0\text{Al})$
treated in acid	-93.41	0.536	89.0	$\text{Q}^3(0\text{Al})$
	-107.87	0.118	11.0	$\text{Q}^4(0\text{Al})$

417 **FIGURE CAPTION**

418

419 **Fig. 1.** XRD patterns of bentonite FEBEX (upper) and MX80 (down): (a, d) starting materials, and,
420 after hydrothermal treatment at 300 °C for 48 h in two different chemical media (b, e) water and (c,
421 f) 0.01 M HNO₃ solution. q=quartz (PDF 04-007-0522), c=cristobalite (PDF 04-008-7742), and,
422 f=feldspar (PDF 04-017-1349).

423 **Fig. 2.** SEM images and EDX spectra of bentonite FEBEX: (a,d) starting materials, and, after
424 hydrothermal treatment at 300 °C for 48 h in two different chemical media (b,e) water and (c,f) 0.01
425 M HNO₃ solution.

426 **Fig. 3.** SEM images and EDX spectra of bentonite MX80: (a,d) starting materials, and, after
427 hydrothermal treatment at 300 °C for 48 h in two different chemical media (b,e) water and (c,f) 0.01
428 M HNO₃ solution.

429 **Fig. 4.** ²⁹Si and ²⁷Al MAS NMR spectra of bentonite FEBEX (upper) and MX80 (down): (a, c)
430 starting materials, and, (b, d) after hydrothermal treatment at 300 °C for 48 h in 0.01 M HNO₃
431 solution.

432 **Fig. 5.** XRD patterns of bentonite FEBEX (upper) and MX80 (down): (a, c) starting materials, and,
433 (b, d) after hydrothermal treatment at 300 °C for 48 h in a $7.3 \cdot 10^{-2}$ M Lu(NO₃)₃ solution. q=quartz
434 (PDF 04-007-0522), c=cristobalite (PDF 04-008-7742), f=feldspar (PDF 04-017-1349),
435 cr=chrysotile (PDF 00-025-0645), k=kaolinite (PDF 01-075-0938), and, *=Lu₂Si₂O₇ (PDF 00-035-
436 0326).

437 **Fig. 6.** ²⁹Si and ²⁷Al MAS NMR spectra of bentonite FEBEX (upper) and MX80 (down): (a, c)
438 starting materials, and, (b, d) after hydrothermal treatment at 300 °C for 48 h in a $7.3 \cdot 10^{-2}$ M
439 Lu(NO₃)₃ solution.

440 **Fig. 7.** a) SEM images of bentonite FEBEX after hydrothermal treatment at 300 °C for 48 h in a
441 $7.3 \cdot 10^{-2}$ M Lu(NO₃)₃ solution. EDX spectra of bentonite FEBEX: (b) starting materials, and, (c,d)
442 after hydrothermal treatment at 300 °C for 48 h in a $7.3 \cdot 10^{-2}$ M Lu(NO₃)₃ solution.

443 **Fig. 8.** (a, b) SEM images of bentonite MX80 after hydrothermal treatment at 300 °C for 48 h in a
444 $7.3 \cdot 10^{-2}$ M $\text{Lu}(\text{NO}_3)_3$ solution. EDX spectra of bentonite FEBEX: (c) starting materials, (d) lamellar
445 particles of Fig. 8.a., and, (e) the ball particles of Fig. 8.b.
446

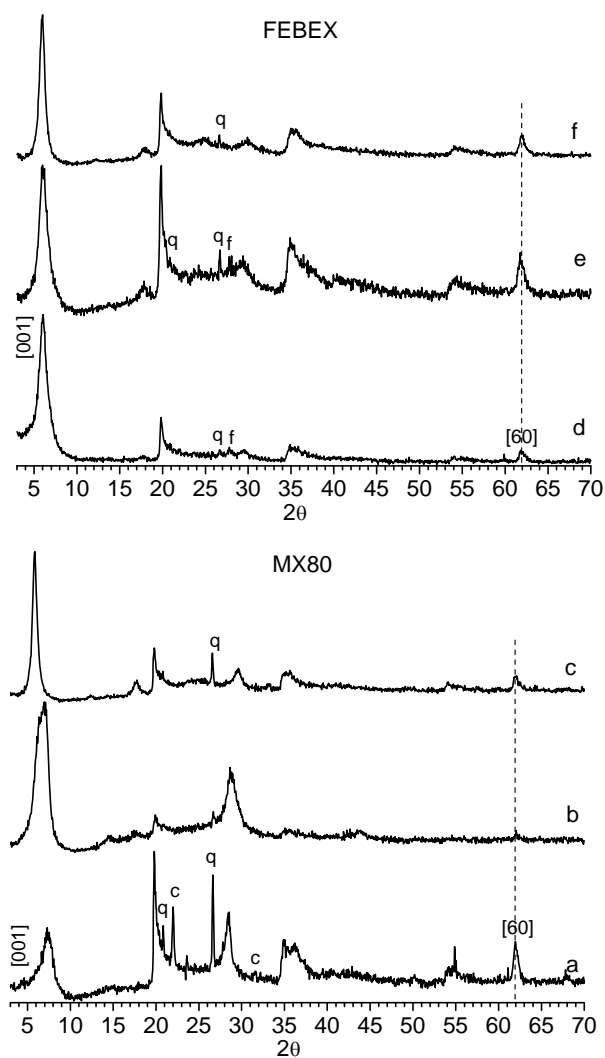
447

448

449

450

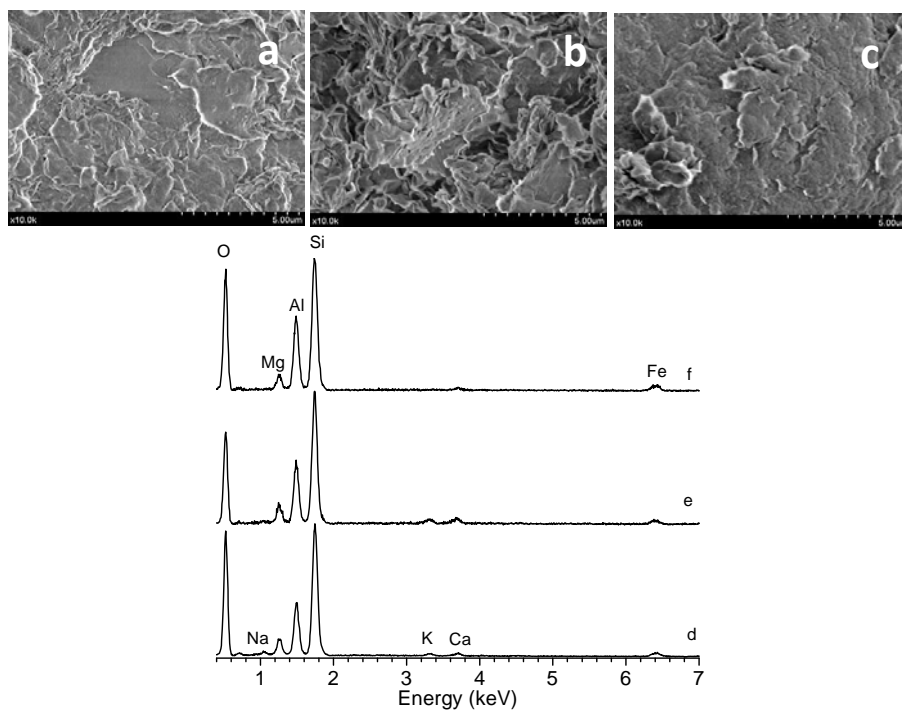
Fig. 1



451

Fig. 2

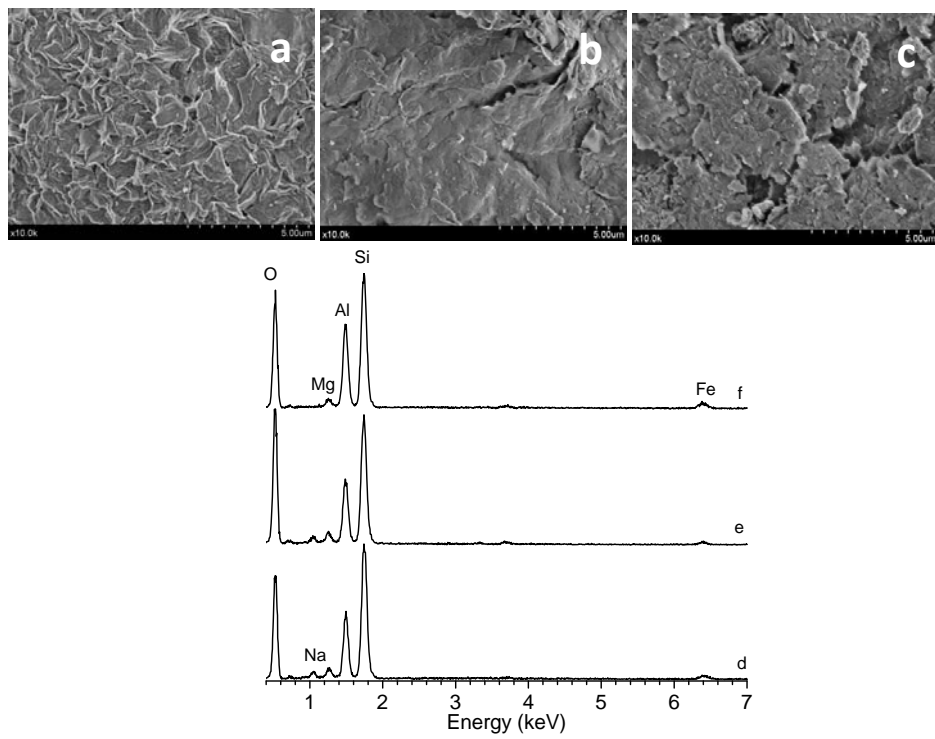
452



453

Fig. 3

454

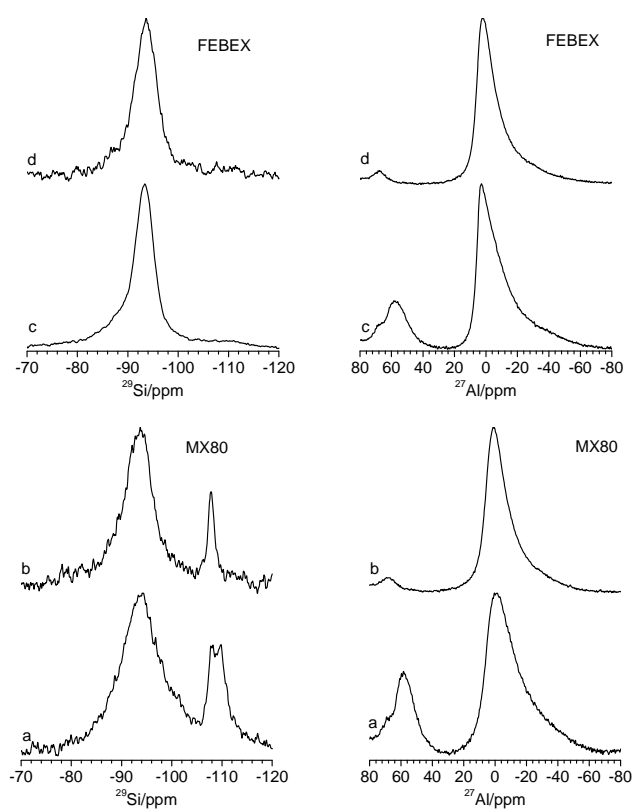


455

456

457

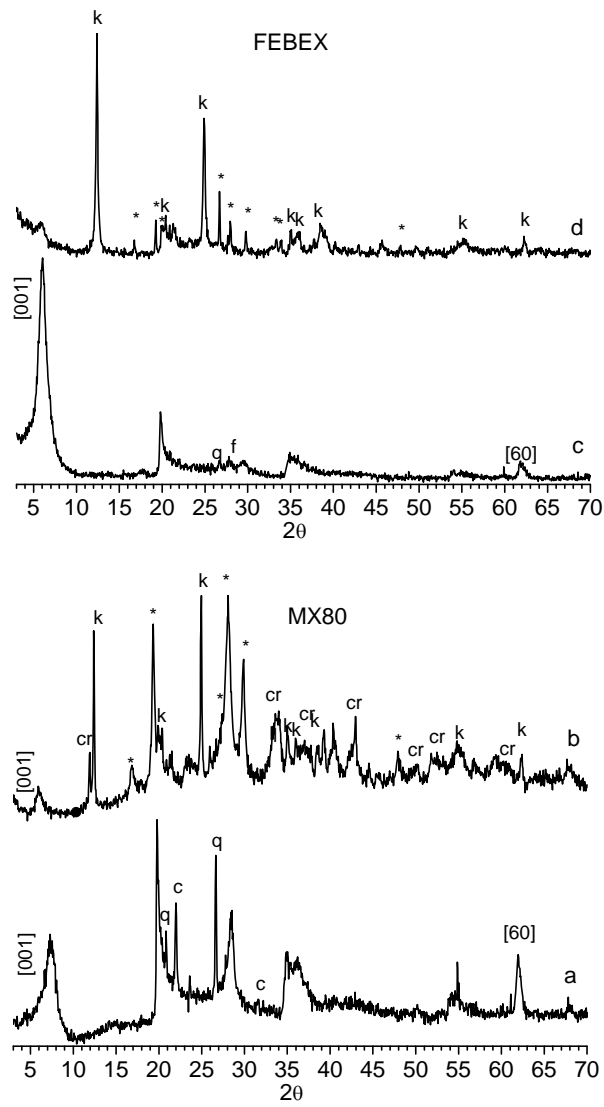
Fig. 4



458

Fig. 5

459



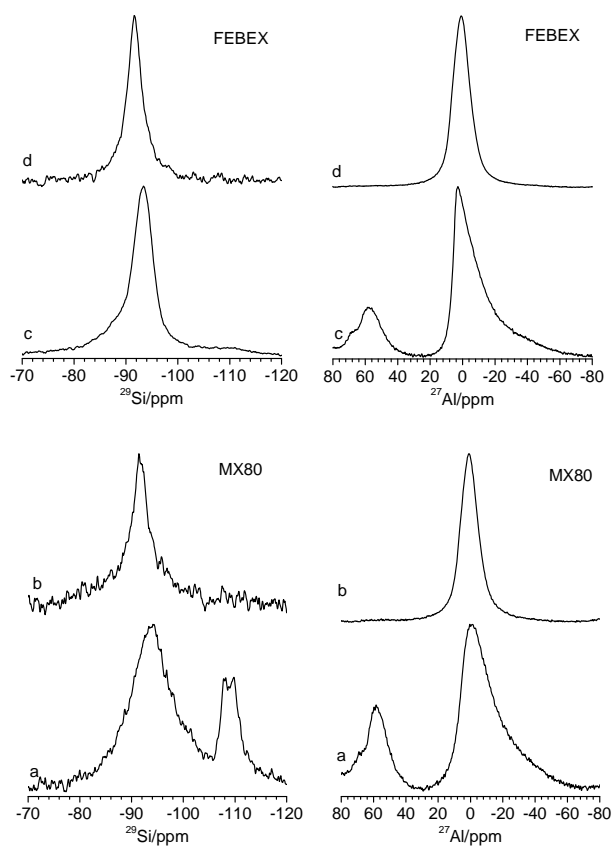
460

461

462

463

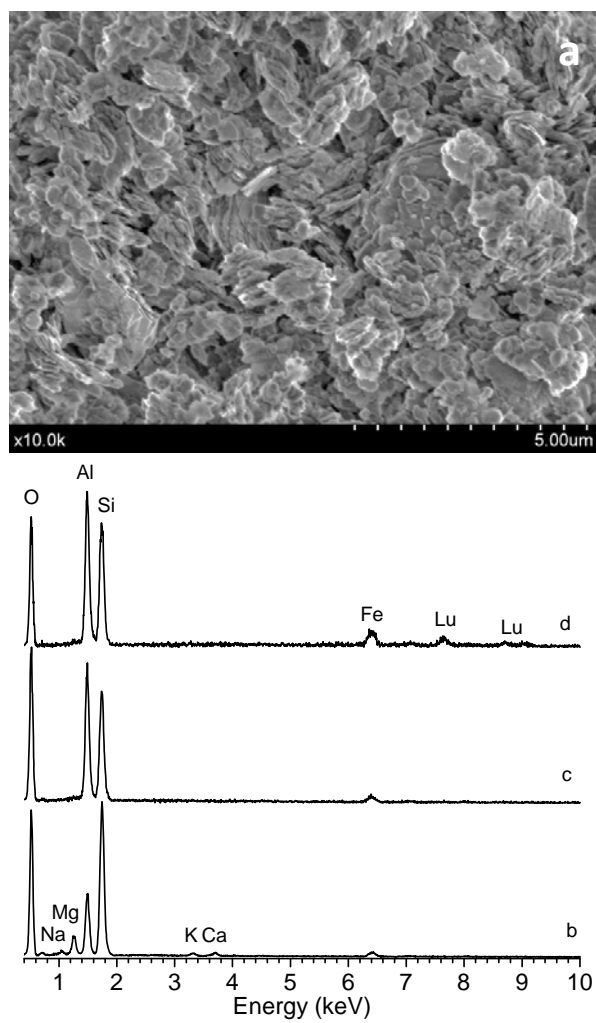
Fig. 6



464

Fig. 7

465



466

467

468

Fig. 8

