

Uranium series isotopes concentration in sediments at San Marcos and Luis L. Leon reservoirs, Chihuahua, Mexico

C. Méndez-García, M. Renteria-Villalobos, R. García-Tenorio, and M. E. Montero-Cabrera

Citation: AIP Conference Proceedings **1607**, 83 (2014); doi: 10.1063/1.4890707 View online: http://dx.doi.org/10.1063/1.4890707 View Table of Contents: http://scitation.aip.org/content/aip/proceeding/aipcp/1607?ver=pdfcov Published by the AIP Publishing

Articles you may be interested in Spatial analysis techniques applied to uranium prospecting in Chihuahua State, Mexico AIP Conf. Proc. **1607**, 116 (2014); 10.1063/1.4890711

Potential human health risk by 234,238 U and 210 Po due to consumption of fish from the "Luis L. Leon" reservoir (Northern Mexico) AIP Conf. Proc. **1607**, 70 (2014); 10.1063/1.4890705

Thorium isotopes in colloidal fraction of water from San Marcos Dam, Chihuahua, Mexico AIP Conf. Proc. **1544**, 25 (2013); 10.1063/1.4813456

Leon M. Lederman and the high energy physics in Mexico AIP Conf. Proc. **531**, 250 (2000); 10.1063/1.1315041

The San Luis project: An attempt to decentralize physics in Mexico Am. J. Phys. **44**, 1034 (1976); 10.1119/1.10580

Uranium series isotopes concentration in sediments at San Marcos and Luis L. Leon reservoirs, Chihuahua, Mexico

C. Méndez-García^{1*}, M. Renteria-Villalobos², R. García-Tenorio³, M. E. Montero-Cabrera^{1a)}

 ¹ Centro de Investigación en Materiales Avanzados, CIMAV, Miguel de Cervantes 120, 31109, Chihuahua, Chihuahua, Mexico.
²Facultad de Zootecnia y Ecología Universidad Autónoma de Chihuahua, Periferico Francisco R. Almada Km 1, 31410, Chihuahua, Mexico.
³Applied Nuclear Physics Group, University of Seville, ETS Arquitectura, Avda. Reina Mercedes s/n, 41012 Seville, Spain.

* PhD Graduate Student at CIMAV

^{a)}Corresponding author: <u>elena.montero@cimav.edu.mx</u>

Abstract. Spatial and temporal distribution of the radioisotopes concentrations were determined in sediments near the surface and core samples extracted from two reservoirs located in an arid region close to Chihuahua City, Mexico. At San Marcos reservoir one core was studied, while from Luis L. Leon reservoir one core from the entrance and another one close to the wall were investigated. ²³²Th-series, ²³⁸U-series, ⁴⁰K and ¹³⁷Cs activity concentrations (AC, Bq kg⁻¹) were determined by gamma spectrometry with a high purity Ge detector. ²³⁸U and ²³⁴U ACs were obtained by liquid scintillation and alpha spectrometry with a surface barrier detector. Dating of core sediments was performed applying CRS method to ²¹⁰Pb activities. Results were verified by ¹³⁷Cs AC. Resulting activity concentrations were compared among corresponding surface and core sediments. High ²³⁸U-series AC values were found in sediments from San Marcos reservoir, because this site is located close to the Victorino uranium deposit. Low AC values found in Luis L. Leon reservoir suggest that the uranium present in the source of the Sacramento – Chuviscar Rivers is not transported up to the Conchos River. Activity ratios (AR) ²³⁴U/²³⁸U and ²³⁸U/²²⁶Ra in sediments have values between 0.9-1.2, showing a behavior close to radioactive equilibrium in the entire basin. ²³²Th/²³⁸U, ²²⁸Ra/²²⁶Ra ARs are witnesses of the different geological origin of sediments from San Marcos and Luis L. Leon reservoirs.

Keywords: Uranium, ²¹⁰Pb-¹³⁷Cs dating, fresh water sediments, radioactivity analysis, Chihuahua.

PACS: 91.67.Pq; 91.67.Qr; 91.67.Ty.

INTRODUCTION

Disequilibrium in uranium and thorium decay series is an important and potential indicators for tracing migration of the these radionuclides in the environment [1]. They can show changes and disturbance in the lithology of a selected site. Usually, (dis)equilibrium is analyzed by means of the *activity ratios* (AR) between given radioisotopes of the same or different decay series.

 U_{nat} and Th concentration in the earth crust is around 1.7 and 8.5 mg kg⁻¹ for U_{nat} and Th respectively. They are associated to some oxides as uraninite, thorianite, which contents tend to have an increased value in igneous rocks.

The distribution of natural radionuclides in sediment cores and their AR has been studied by several authors [2-7] to know the effects of soil properties in biogeochemical mobilization of these radionuclides and understand their behavior in the environment.

The selected site for this study is located in the state of Chihuahua, Mexico's northern border. The Basin of Sacramento-Chuvíscar Rivers, which constitute the last major tributaries to the Rio Conchos, originates in the vicinity of a uranium deposit in San Marcos, Chihuahua, where a reservoir of the same name is located. The last Rio Conchos reservoir before joining the Rio Grande is created artificially by the Luis L. Leon "El Granero" dam [8].

Radiation Physics AIP Conf. Proc. 1607, 83-92 (2014); doi: 10.1063/1.4890707 © 2014 AIP Publishing LLC 978-0-7354-1243-9/\$30.00 In Chihuahua State, several radioanomalies have been found. Some of them are located in San Marcos basin [9]. Here, both surface and ground waters are important resources to the region. The surface water gathered in San Marcos dam reservoir is used for agricultural purposes, whereas groundwater from shallow and deep wells is used for both agricultural and domestic activities. Previous studies in the mentioned site show the presence of uranium in surface waters, fish, and some plants [10-13].

Data about reservoir's water radioisotopes content shows high uranium specific activities up to 7.7 $Bq \cdot L^{-1}$. The high uranium concentrations in the area are attributed to the lixiviation and/or erosion of the uraniferous deposits located in this area [14-16].

Natural uranium and thorium concentration has not been reported before in the Luis L. Leon area. The main purpose of this paper is to verify if U- and Th radioisotopes are transported from San Marcos area to Luis L. Leon reservoir.

STUDY AREA, MATERIALS AND METHODS

Two sites were selected for this stud, the first is San Marcos reservoir, located approximately 25 kilometers West-Northwest to Chihuahua City, Mexico. It was built in the beginning of the XX century. It has a capacity of 4.45 million m³ of water and is used in flood control and agriculture in the area. The lithology of the study area consists mainly of volcanic and volcanoclastic rocks of rhyolitic composition dikes and resurgent rhyolite domes and dikes of intermediate to basic composition [9]. It also has few outcrops of Cretaceous limestone in the northern boundary of the basin [16]. Uranium deposits at San Marcos, named as Victorino and San Marcos I, and their minerals have been recently described by Reyes-Cortés, et al. [17]. The mineral species found in the San Marcos deposits were: uranophane and metatyuyamunite at San Marcos I site; uranophane, uraninite, masuyite and becquerelite at Victorino site [17]. The other is the Luis L. Leon reservoir, located approximately 90 kilometers Northeast of Chihuahua City. It was built in 1968, having a water capacity of 854 million m³. The reservoir is used in the fish hatchery to control the species that there exist, farm, agricultural and domestic use. The rocks exposed in the region are limestone, igneous intrusive and volcanic, whose ages range is from the Upper Jurassic to the Recent. At least three abandoned mines are located near to this reservoir: Carrizalillo, Chorreras and La Verde [8].

Sampling

Three sediment cores were collected manually following a well-established protocol with a steel tube (7 cm diameter and 40 cm length). Figure 1 shows the geographic location of Chihuahua State, as well as of the two reservoirs studied. One sediment core was extracted at 20 m depth from San Marcos reservoir in September 2007 (CSM), at a place close to the dam's wall. The two cores from Luis L. Leon reservoir were taken at 10 m depth, one near to the reservoir's entrance (ECLLL) and the other at the northern zone (NCLLL) in April 2012. Also some surface sediment samples were taken, close to the sediment cores sampling sites; table 1 shows the details of the location of sediment cores and surface sediments samples.



FIGURE 1. Study area showing the location and the sampling points for a) San Marcos and b) Luis L. Leon reservoirs

Reservoir	Sample type	Sample ID	Location (lalitude, longitude)
San Marcos.	Core	CSM	28.74174, -106.35683
	Surface sediment	SM1	28.77962, -106.35839
	Surface sediment	SM2	28.78555, -106.36757
	Surface sediment	SM3	28.78999, -106.36956
	Surface sediment	SM4	28.79408, -106.36915
	Surface sediment	SM5	28.79525, -106.36228
Luis L. Leon.	Core	ECLLL	28.88401, -105.30953
	Core	NCLLL	28.95626, -105.29211
	Surface sediment	LLL1	28.94499, -105.27856
	Surface sediment	LLL2	28.91572, -105.27756
	Surface sediment	LLL3	28.92618, -105.29260
	Surface sediment	LLL4	28.90981, -105.30293
	Surface sediment	LLL5	28.86896, -105.31165
	Surface sediment	LLL6	28.94045, -105.31855
	Surface sediment	LL7	28.96748, -105.30624

TABLE 1. Sampling details of sediment cores and surface sediments collected at San Marcos and Luis L Leon reservoirs' areas.

SM represents San Marcos area.

LLL represents Luis L. Leon area.

Laboratory analysis

The cores were cut into 2 cm thick sections. Sections were dried at 55 °C to constant weight, ground to a powder with an agate mortar pestle, homogenized by sieving to <63 µm grain size and packed for further analysis.

Organic matter (OM) content was estimated by loss on ignition (LOI) at 550°C according to [18].

 210 Pb activity was determined by alpha counting of 210 Po, assuming secular equilibrium between 210 Pb, ²¹⁰Bi and ²¹⁰Po in the sediment after the sampling. The powder was digested with aqua regia (1:2) and ²¹⁰Po extracted using the technique of liquid-liquid extraction with tributyl phosphate (TBP), autodeposited on cupper disc and measured using an Alpha Analyst TM (CANBERRA) spectrometer. Samples were spiked with ²⁰⁹Po as radiotracer yield [19-22]. ²³⁸U and ²³⁴U ACs were determined for all samples, except in surficial sediments from San Marcos

reservoir.

To obtain ²³⁸U and ²³⁴U ACs, all samples were spiked with ²³²U and put under the radioanalytical analysis procedure. Total sample dissolution was performed by atmospheric acid digestion using HFc and HNO₃c (10:10). The uranium extraction was carried out by chromatography using UTEVA resins. Then, uranium was electrodeposited on stainless steel disc. Radiochemical yield was determined by the ²³²U counting rate [23-26]. The ²³⁸U, ²³⁴U ACs were also determined by alpha activity measurements, using the same Alpha spectrometer.

All sections from the sediment cores and surface samples were analyzed for ²³⁸U- and ²³²Th-series isotopes (reported here as ²²⁶Ra and ²²⁸Ra), as well as ⁴⁰K and ¹³⁷Cs activities, by high resolution gamma spectrometry using a Canberra HPGe detector. Prior to the analysis, aliquots around 20 g were stored into petri dishes (15mm height per 75 diameter capacity), and hermetically sealed for at least 21 days in order to reach radioactive equilibrium between ²²⁶Ra and ²²²Rn, and their progenies. The detector was calibrated using RGU, RGTh and RGK IAEA certified reference materials in the same geometry as the measured samples. Counting time was 48 h; the analytical precision of the measurements was around 10%. Considering the appropriate corrections for laboratory background, the activity of ²²⁶Ra (from ²³⁸U-series) was determined from the 609 keV line of ²¹⁴Bi; ⁴⁰K activity from its 1461 keV line, ²²⁸Ra (from ²³²Th-series) activity was estimated from the 911 keV line of 228 Ac and 137 Cs from its 662 keV line.

To carry out the sediments' dating procedure, total ²¹⁰Pb_{tot} activities were estimated by measuring the activity of ²¹⁰Po, where secular equilibrium from ²²⁶Ra to ²¹⁰Pb, ²¹⁰Bi and ²¹⁰Po in the sediment *after packing* is assumed [27] The sediment accumulation rate was obtained by the activity in excess ²¹⁰Pb_{exc} as calculated from the difference of total and supported ²¹⁰Pb. Supported ²¹⁰Pb activity is equal to ²²⁶Ra activity, by the secular equilibrium with its progenies attained before the measurement.

Dating

²¹⁰Pb_{exc} (unsupported) was calculated in all cores in order to know which model would be applied [19].

$${}^{210}Pb_{exc} = {}^{210}Pb_{total} - {}^{210}Pb_{sup} \qquad (1)$$

The Constant Rate of Supply model was selected; this model assumes that the supply of ${}^{210}Pb_{exc}$ to the sediment is constant through time [28], although the profile of this isotope may reflect the interaction of the sedimentation rate (SR) and radioactive decay. Then, ${}^{210}Pb_{exc}$ and SR can be variable through the time.

In this model the age t of each section at depth z is estimated by the following expression:

$$t = \frac{1}{\lambda_{Pb}} \ln \left(\frac{I(0)}{I(z)} \right)$$
(2)

Where I(0) (Bq m⁻²) is the total ²¹⁰Pb_{exc} inventory of the sediment core and I(z) is the total ²¹⁰Pb_{exc} in the sediment layer below depth z.

The sedimentation rate R (kg m⁻² y⁻¹) is obtained directly from:

$$\boldsymbol{R} = \frac{\lambda_{Pb}I(z)}{A(z)} \quad (3)$$

Estimated ages obtained by applying this method were confirmed by comparison of the value derived from the ¹³⁷Cs activity measurements. Fallout of ¹³⁷Cs introduce this radioisotope in the water, and then in sediments. ¹³⁷Cs AC in sediment profiles has a peak in the vicinity of 1962-64 years associated to nuclear bomb testing in the atmosphere [29, 30], and in northern hemisphere, it has another peak related with year 1986 from the radionuclide dispersion produced by the Chernobyl accident [31-33].

RESULTS AND DISCUSSION

Sampling site of sediments affects both lithology and the environmental factors that may alter the weathering of the elements; consequently, sediments' radionuclides concentrations will be affected accordingly. The San Marcos reservoir has a narrowing that divides it into two vessels (West and East), and this may influence the sediment transport downstream along the lake (Figure 1). The sediment core CSM was collected in the vessel nearer to the dam's wall. This location is close to a spring located at the bottom of the San Marcos reservoir. The San Marcos dam collects water that runs from the West at the birth of the creek of the same name, near the Victorino uranium deposit during the rainy seasons. During these periods turbid water occurs frequently and it shows high organic matter content. During periods of drought in the lake the water level drops and the water looks very clean one.

Total ²¹⁰Pb_{tot} activities ranged from 84 to 313 Bq kg⁻¹, and the supported ²¹⁰Pb_{sup} activity varied from 60 to 275 Bq kg⁻¹. ²¹⁰Pb_{exc} activities were obtained by eq. (1). The ²¹⁰Pb_{exc} inventory in CSM was estimated to be 207.5±4 Bq m⁻², the mean ²¹⁰Pb_{exc} AC was 34.8 Bq kg⁻¹ and its distribution did not show a monotonic behavior whereby the CRS model dating was applied (Figure 2). The ²¹⁰Pb_{exc} distribution is characterized by a near exponential decline, behavior typical of undisturbed soils. The model employed was corroborated by ¹³⁷Cs activities, where its maximum concentration were found at 36 cm depth, corresponding to 1963 weapons fallout peak, and at 16 cm depth, relevant to Chernobyl accident in 1986. The results obtained by

CRS method were accepted, as there is no discrepancy with the ¹³⁷Cs maker. The sedimentation rate calculated by eq.(3) shows high variability with time. These results were compared with data of rainfall in the region and a high correlation (r = 0.89) was observed. Consequently, low values in sedimentation rate belong to drought seasons and conversely high sedimentation rates are associated with periods of high rainfall contribution to the reservoir.



FIGURE 2. ²¹⁰Pb, ¹³⁷Cs concentration activities and sedimentation rate profiles (with uncertainty about 3 years) in sediment core in San Marcos reservoir (CSM).

All radionuclide concentration values were different if compare them in surface and core sediments (Table 2). Radionuclide concentrations were higher in surface sediments than in the sediment core, except for ¹³⁷Cs, because it corresponds to the two mentioned fallout events.

TABLE 2. Comparison of radionuclide concentrations among core and surface sediments at San Marcos reservoir. Different letter means that significant difference exists. Letter A means higher concentration.

	²²⁸ Ra	²³⁸ U	²²⁶ Ra	⁴⁰ K	¹³⁷ Cs
CSM	В	В	В	В	А
SMS	А	А	А	А	В

⁴⁰K, ²²⁸Ra (²³²Th), ²²⁶Ra and ²³⁸U AC mean values in surface sediments were 1164, 377, 856, and 893.5 Bq kg⁻¹respectively (table 3); these results are alike those found in parent rocks in a previous study of the San Marcos – Victorino area [34] and correspond to their sampling locations, as pointed out above. On the contrary, core sediments' ACs are lower. These concentrations suggest that adsorption in core sediments correspond to the characteristics of the water column, provided in part by the spring located at the bottom of the reservoir, near the point where the sediment core was extracted.

Activity concentrations in both surface and core sediments are higher than the reported by UNSCEAR (2000), where reference values for earth crust are 35, 35, 30 and 400 Bq kg⁻¹ for ²²⁶Ra, ²³⁸U, ²²⁸Ra and ⁴⁰K, respectively. In core sediments ²³⁴U and ²³⁸U content shows a quite constant distribution (coefficient of variation =10%).

Activity ratios (AR) of U-series isotopes help in analyzing (dis)equilibrium and they may serve to reveal the relationship between factors such as weathering and adsorption [35]. The 234 U/ 238 U AR in core sediments were between 1.1 – 1.4 with a mean of 1.2, values close to the equilibrium (Table 3). High AR values were found in layers close to the water interface. This observation is consistent with values reported for San Marcos reservoir water in [14], that were obtained in a sampling performed at 2005, when the water level in reservoir was very low and most water in the vessel probably was supplied by the above mentioned spring.

 40 K, 228 Ra (232 Th) and 226 Ra ACs show a coefficient of variation >20% in the sediment core profile (Figure 3); 40 K AC shows some increasing behavior with depth, finding the highest at the bottom of the core. 228 Ra (232 Th) shows variability in the profile due to high AC found in the first and last layers; if these values were considered as outliers, the behavior of the 228 Ra (232 Th) would be almost constant with depth, which could be explained by the low solubility of thorianite (ThO₂) and Th strong sorption onto the sediments. 226 Ra AC shows a non-monotonic behavior in the core profile.

The values of ²²⁶Ra ACs divide the layers into two main groups, one for the layers closest to the surface, with ACs between 83 and 105 Bq kg⁻¹, and the other group for the deepest layers, with ACs between 57 and 71 Bq kg⁻¹. This behavior suggests that in deep strata reducing conditions prevail where ²²⁶Ra is more soluble, while in shallow layers oxidation conditions could be found.

CSM	⁴⁰ K	²²⁶ Ra	²²⁸ Ra	²³⁸ U	²³⁴ U	²²⁸ Ac/ ²³⁸ U	²³⁸ U/ ²²⁶ Ra	²²⁶ Ra/ ²²⁸ Ra	²³⁴ U/ ²³⁸ U
Mean	971.9	80.4	147.6	92.7	112.7	1.6	1.2	0.6	1.2
S.D	181.9	16.4	30.4	9.6	10.5	0.3	0.3	0.2	0.1
Range	749.9 - 1257.6	57 - 104.8	106.9 - 210.6	78.7 - 106.7	99.5 - 125.9	1.1 -2-4	0.8 - 1.7	0.3 - 0.7	1.1 - 1.4
C.V (%)	18.7	20.4	20.6	10.4	9.3	20.8	24.0	30.3	6.5
SSM									
Mean	1163.7	856.1	376.8						
S.D	26.2	68.7	71.6						
Range	1123.8 - 1185.9	738.2 - 897.6	321.3 - 501.9						
C.V (%)	2.3	8.0	19.0						

TABLE 3. Statistics of radionuclide concentrations and AR in San Marcos area.

²³⁸U-and ²³²Th-series isotopes commonly occur together in nature; the ²³²Th/²³⁸U mass ratio is around 3.5 in almost all natural systems [36]. AR mean value in earth crust is 1.2 [37]. The ²²⁸Ra (²³²Th) /²³⁸U AR mean value in the sediment core was higher than 1.2. ²³⁸U/²²⁶Ra ARs show disequilibria (AR \neq 1) in the ²³⁸U chain at the first layer and the deepest four in the sediment core, where were found AR values between 1.3 and 1.7. This behavior may be explained by their differential ion mobility [6]. ²²⁸Ra (²³²Th) /²²⁶Ra AR mean value in the core was 1.8, higher than AR=1.1 reported by (Evans 1997) in most environmental samples. Usually this AR is applied to assess the conservation of the proportion within ²³²Th and ²³⁸U decay series. Thereby, in this sediment core the expected proportion could not been observed.



FIGURE 3. Distribution of radionuclides in sediment core in San Marcos reservoir.

Luis L Leon reservoir collects water that comes from the Rio Conchos basin. Its elongated shape has a length of approximately 12 km. Its relatively recent construction from a rough relief area but with a steady stream of water, produces a reservoir bottom of abundant vegetation including trees, which are now submerged. The vegetation and the topography may play the role of filtering the solutions and suspended material in the water flowing along the reservoir.

Both entrance and northern sediment cores (ECLLL and NCLLL) extracted from Luis L. Leon reservoir show a non-monotonic distribution of ²¹⁰Pb_{exc} concentration with depth, but they have an exponential decreasing. The CRS model dating was applied for them. ²¹⁰Pb_{tot} activities ranged from 48 to 59 Bq kg⁻¹, and 41 to 55 Bq kg⁻¹ for ECLLL and NCLLL, respectively (Figure 4). The ²¹⁰Pb_{exc} inventory in ECLLL and NCLLL were estimated to be 194 ± 5 and 179 ± 3 Bq m⁻², respectively. The age calculated for the deepest layer of the ECLLL was down to the year 1976, whereas for the NCLLL it was 1971. The maximum ¹³⁷Cs activity concentration was found around 1989 in both cases, which suggest the acceptance of the dating model applied in this reservoir by the radioisotope concentration of the Chernobyl accident fallout. The sedimentation rates in ECLLL are correlated with rainfall (0.708 p<0.001) but not in the case of NCLLL. This could be related to the variability of the annual water discharge from the dam, producing changes in the sedimentation rates near dam wall.

U_{nat} concentrations were one order of magnitude higher in SMS and CSM than in sediments from the Luis L. Leon reservoir area. Activity concentrations of U_{nat} isotopes were higher in ECLLL than in NCLLL and SLLL (Table 4), suggesting that radionuclides are transported by the different streams reaching the reservoir. This feature suggests that radioisotopes are filtered from the water by the reservoir accidents, preventing their further transport downstream to the Rio Conchos. ²³²Th, ²³⁸U, ²²⁶Ra ACs did not show difference between NCLLL and SLLL, suggesting that the core taken

²³²Th, ²³⁸U, ²²⁶Ra ACs did not show difference between NCLLL and SLLL, suggesting that the core taken in this point could retain the features of the Luis L. Leon surface lithology; the difference in ⁴⁰K concentration in these samples is due to low content of feldspar in NCLLL.



FIGURE 4.²¹⁰Pb, ¹³⁷Cs concentration activities and sedimentation rate profiles in sediment core in Luis L. Leon reservoir. ECLLL means entrance and NCLLL Northern Core.

TABLE 4. Comparison of radionuclide concentrations among core and surface sediments at Luis L Leon reservoir area. Different letter means that significant difference exists. Letter A means higher concentration.

	²²⁸ Ra	²³⁸ U	²²⁶ Ra	⁴⁰ K	¹³⁷ Cs
ECLLL	А	А	А	А	А
NCLLL	В	В	А	В	А
SLLL	В	В	А	А	В

Only ACs values of 40 K were above the reference values established by [38] (400 Bq kg⁻¹) (Table 5). The high concentration of potassium corresponds to the feldspar appearing in the diffraction patterns.

Activity concentration of potassium corresponds to the refuspar appearing in the diffraction patterns. Activity concentrations were homogenous in both sediment cores, the lower ACs were found in NCLLL. 228 Ra(232 Th)/ 238 U AR in ENLLL, NCLLL and surface sediments were about one, as in the upper earth crust. 234 U/ 238 U ARs are alike in both sediment core (1.1 – 1.2) indicating equilibrium between isotopes. The same may be observed for 226 Ra/ 238 U AR.

The radionuclide concentrations in both sediment cores did not show high temporal variability (C.V <20%). 238 U and 234U showed almost the same behavior in both sediment cores, the concentrations of these radionuclides from certain depth to the bottom of the cores is almost constant (Figure 5).



FIGURE 5. Distribution of radioisotopes in sediment core in Luis L. Leon reservoir.

ECLLL	⁴⁰ K	²²⁶ Ra	²³² Th	²³⁸ U	²³⁴ U	²²⁸ Ac/ ²³⁸ U	²³⁸ U/ ²²⁶ Ra	²²⁶ Ra/ ²²⁸ Ra	²³⁴ U/ ²³⁸ U
Mean	526.4	29.3	41.2	40.5	47.6	1.0	1.0	0.9	1.2
S.D	68.1	4.6	8.6	1.8	2.1	0.1	0.1	0.2	0.03
Range	452.3-652.5	33.3-46.4	28.5-56	37.1-43.8	42.7-52.5	0.8-1.1	0.8-1.1	0.6-1.2	1.1-1.2
C.V (%)	12.9	15.7	20.9	4.4	4.4	10.0	10.0	20.0	10.0
NCLLL									
Mean	422.2	25.3	29.0	29.5	33.9	1.4	1.0	0.7	1.2
S.D	23.7	3.7	2.2	3.0	3.3	0.3	0.2	0.2	0.02
Range	383.2-461.7	18.9-32.1	23.9-32.94	23.4-32.9	27.6-37.8	0.8-2.0	0.8-1.4	0.4-1.3	1.1 1.2
C.V (%)	5.6	14.7	7.7	10.2	9.6	22.4	16.6	25.8	1.9
SCLLL									
Mean	517.4	26.6	29.0						
S.D	101.6	4.3	5.2						
Range	352.8-659.1	18.1-31.1	24.3-38.7						
C.V (%)	19.6	16.1	17.8						

TABLE 5. Statistics of radionuclide concentrations and AR in Luis L. Leon area.

CONCLUSIONS

Variations in radionuclide concentrations with depth in all sediment cores provided information about the behavior of these in three sediment cores from two different reservoirs, and the way that they are transported or mobilized in the environment.

 U_{nat} concentrations in San Marcos area sediments are affected by the erosion or lixiviation of the rocks from the natural deposits.

 U_{nat} concentrations were one order of magnitude higher in SMS and CSM than in sediments from the Luis L. Leon reservoir area. These results suggest that erosion and weathering products of rocks from Victorino uranium deposit in the San Marcos area, that are assumed to be responsible of observed uranium concentration values in groundwater of the Chihuahua-Sacramento Valley [16], we conclude that at least in recent times do not reach the low Rio Conchos subbasin.

The non-anthropogenic or natural origin from San Marcos area uranium contents suggest that the natural mean concentration of U (2.3 mg kg⁻¹) in Luis L. Leon area may be considered as the Natural Concentration Level in the Rio Conchos basin.

REFERENCES

- [1] M. Ivanovich, Radiochim. Acta 64, 81-94. (1994).
- [2] T.-C. \$Chu, and J.-J. Wang, Appl. Radiat. Isot. 48, 1149 (1997).
- [3] F. B. Ribeiro *et al.*, Appl. Radiat. Isot. **54**, 153 (2001).
- [4] H. R. von Gunten *et al.*, Chem. Geol. **160**, 225 (1999).
- [5] J. C. Lozano, P. Blanco Rodríguez, and F. Vera Tomé, J. Environ. Radioact. 63, 153 (2002).
- [6] A. Navas *et al.*, J. Soils Sed. **11**, 1082 (2011).
- [7] A. Navas, J. Soto, and J. Machín, Appl. Radiat. Isot. 57, 579 (2002).

- [8] INEGI, (Instituto Nacional de Estadística Geografía e Informática, Aguas Calientes, México, 1999), p. 222.
- [9] H. Ferríz, in Uranium deposits in volcanic rocks (IAEA El Paso, Texas, 1985), pp. 197.
- [10] M. E. Montero Cabrera et al., in VI Congreso Internacional y XII Nacional de Ciencias AmbientalesChihuahua, Mexico, 2007).
- [11] M. Rentería Villalobos *et al.*, in *2nd International Conference on Radioactivity in the Environment* (International Union of Radioecology, Nize, France, 2005).
- [12] M. Y. Luna Porres *et al.*, RMxF **57**, 40 (2011).
- [13] J. C. Burillo Montufar *et al.*, Appl. Geochem. **27**, 1111 (2012).
- [14] M. Rentería Villalobos *et al.*, RMxF **S53**, 16 (2007).
- [15] M. Reyes-Cortes *et al.*, RMxF **\$53**, 23 (2007).
- [16] M. Reyes-Cortés *et al.*, Revista Mexicana de Ciencias Geológicas **29**, 659 (2012).
- [17] M. Reyes-Cortés et al., MinPe 99, 121 (2010).
- [18] E. M. Bengtsson L, Handbook of Holocene palaeoecology and palaeohydrology. Chichester: John Wiley & Sons Ltd., 423 (1986).
- [19] R. García-Tenorio García-Balmaseda, in *Física Atómica, Molecular y Nuclear* (Universidad de Sevilla, Sevilla, 1986), p. 282.
- [20] J. Sanchez-Cabeza, P. Masqué, and I. Ani-Ragolta, J. Radioanal. Nucl. Chem. 227, 19 (1998).
- [21] M. A. A. Aslani *et al.*, Radiat. Measur. **39**, 129 (2005).
- [22] D. Desideri et al., J. Environ. Radioact. 94, 86 (2007).
- [23] L. Hallstadius, NIM Phys Res 223, 266 (1984).
- [24] N. A. Talvitie, Anal. Chem. 44, 280 (1972).
- [25] R. N. Dos Santos, L. S. Marques, and F. B. Ribeiro, Appl. Radiat. Isot. 56, 741 (2002).
- [26] C. Méndez et al., JMatS 45, 5061 (2010).
- [27] W. R. Schell, Nevissi, A., Technical Reports Series No. 91. IAEA, Vienna Chap. 11. (1983.).
- [28] J.-A. Sanchez-Cabeza, M. Díaz-Asencio, and A. C. Ruiz-Fernández, *Radiocronología de Sedimentos Costeros Utilizando*²¹⁰*Pb: Modelos, Validación y Aplicaciones* (IAEA, Vienna, Austria, 2012), p. 105.
- [29] D. E. Walling, and Q. He, CATENA 29, 263 (1997).
- [30] J. C. Ritchie, and J. R. McHenry, Journal of environmental quality 19, 215 (1990).
- [31] E. Garcia Agudo, in IAEA-TECDOC-1028 (IAEA, Vienna, 1998), p. 123.
- [32] C. M. Alonso-Hernandez et al., CSR 26, 153 (2006).
- [33] R. T. Lowson, and M. G. McIntyre, J. Environ. Radioact. 118, 157 (2013).
- [34] M. Reyes Cortés, in *Física de Materiales* (Centro de Investigación en Materiales Avanzados, Chihuahua, México, 2009), p. 156.
- [35] F. Chabaux, B. Bourdon, and J. Riotte, in *Radioactivity in the Environment*, edited by D. Porcelli, S. Krishnaswami, and J. K. Cochran (Elsevier, 2008), pp. 49.
- [36] M. Ivanovich, and R. S. Harmon, *Uranium-series disequilibrium: applications to earth, marine, and environmental sciences* (Clarendon Press, 1992).
- [37] D. Porcelli, and P. W. Swarzenski, Reviews in Mineralogy and Geochemistry 52, 317 (2003).
- [38] United Nations Scientific Committee on the Effects of Atomic Radiation, (UNSCEAR, New York, USA., 2000)