

Soil-plant system and potential human health risk of chinese cabbage and oregano growing in soils from Mn- and Fe-abandoned mines: Microcosm assay

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1 **Abstract**

2 In Portugal many abandoned mines are often close to agricultural areas and might be
3 used for plant food cultivation. Soils in the vicinity of two Mn- and Fe- abandoned
4 mines (Ferragudo and Rosalgar, SW of Portugal) were collected to cultivate two
5 different food species (*Brassica rapa* subsp. *pekinensis* (Lour.) Hanelt and *Origanum*
6 *vulgare* L.). Chemical characterization of the soil-plant system and potential risk of
7 adverse effects for human health posed by plants associated with soil contamination,
8 based on the estimation of hazard quotient (HQ), were assessed in a microcosm assay
9 under greenhouse conditions. In both soils, the average total concentrations of Fe and
10 Mn were above the normal values for soils in the region and their concentration in
11 shoots of both species was very high. *Brassica rapa* subsp. *pekinensis* grew better in
12 Ferragudo than in Rosalgar soils and it behaved as an excluder of Cu, Mn, Fe, S and Zn
13 in both soils. The HQ for Cu, Fe, Mn and Zn in the studied species grown on both soils
14 was lower than unit indicating that its consumption is safe. The high Mn tolerance
15 found in both species might be due in part to the high contents of Fe in the soil available
16 fraction that might contribute to an antagonism effect in the uptake and translocation of
17 Mn. The obtained results emphasize the need of further studies with different food crops
18 before cultivation in the studied soils to assess health risks associated with high metal
19 intake.

20 **Keywords:** Mining, Human health risks, Chinese cabbage, Oregano

21 **Introduction**

22 Mining activities can be responsible for the release of several potentially toxic
23 elements in soil, water and air. Nonetheless, the abandoned mines often cause more
24 environmental impacts than active mines due to the inexistence of a legislation

25 requiring rehabilitation of mining areas after closure and/or Government inspection
26 (Kim et al., 2012; Santos et al., 2018).

27 Soils adjacent to the mining areas and/or developed on contaminated materials
28 (*e.g.* mine wastes) can be a source of potentially toxic elements for plants (Kabata-
29 Pendias 2004; Abreu et al. 2014a). Although plants only take up elements from the
30 available soil fraction, which corresponds to the amount of the chemical elements in the
31 soil solution and associated to the exchangeable complex of inorganic and organic soil
32 colloids, several characteristics related to plant species, climate and soils can affect this
33 fraction (Adriano 2001; Kabata-Pendias 2004). After uptake, the distribution and
34 accumulation of the elements in the different plant parts at different concentrations can
35 vary, for instance, with the element and plant species (Abreu et al. 2014a). The elements
36 accumulation in the edible part of plants can represent the principal entry for food chain
37 (van Rijn et al. 2002). Therefore, the soil enrichment with elements due to mining
38 activities, namely in kitchen gardens, increases the possibilities of these elements
39 getting into the food chain at levels that may be dangerous to human health (Abreu et al.
40 2014a; Hough et al. 2004; Zhuang et al. 2009). In fact, the consumption of
41 contaminated vegetables can be an important route to metal exposure (Hough et al.
42 2004; Zhuang et al. 2009; Abreu et al. 2014a). Consumption of food with high
43 concentrations of potentially toxic elements is a major contributor (more than 90%) to
44 human exposure to elements toxicity (Loutfy et al. 2006). This is the reason why
45 research on environmental health risks has increased in recent years (Árvay et al. 2017;
46 Chen et al. 2019) especially in the areas affected by mining activities.

47 El Hamiani et al. (2010, 2015) studied the effects on consumer's health after
48 eating food crops and/or vegetables produced in the vicinity of Southern Morocco
49 mines pointing out the risks, especially due to high concentrations of Mn, in vegetables

50 (*Lactuca sativa* L. and *Lolium multiflorum* L.) even if the consumption of the edible
51 part is relatively free of risks. In kitchen gardens closed to abandoned pyrite mines from
52 South of Portugal, the potential risk for human health depended on the plant species and
53 element. Thus, the accumulation of Cu and Zn in lettuce (*Lactuca sativa* L.), coriander
54 (*Coriandrum sativum* L.) and cabbage (*Brassica oleracea* L.) from Aljustrel, Lousal
55 and São Domingos did not represent risks for human health (Alvarenga et al. 2014).
56 However, coriander and cabbage cultivated in kitchen gardens near São Domingos mine
57 exceeded the maximum allowed values for Pb and As indicating a possible health risk
58 (Gonzalez-Fernandez et al. 2011). Studies carried out by Neves et al. (2012) in a
59 uranium contaminated area from North-Centre Portugal showed that the estimated level
60 of U exposure through the ingestion of several vegetable foodstuffs (*Lactuca sativa* L.,
61 *Solanum tuberosum* L., *Phaseolus vulgaris* L., *Daucus carota* L., *Brassica oleracea* L.,
62 *Malus domestica* Borkh and *Zea mays* L.) growing in the area was low, suggesting no
63 chemical health risk for local residents during their lifetime. Another study on the level
64 of toxic elements in foodstuffs from soils closed to abandoned mines in Korea
65 concluded that the estimated daily metal intake was below the provisional tolerable
66 daily intake limits (Ji et al. 2013) and in Uganda the consumption of vegetables growing
67 in farmer gardens around Kampala City was also safe (Nabulo et al. 2010). However,
68 residents in a village close to abandoned Cu mines (Goseong, Korea) denoted high Cd
69 concentrations in the blood and urine (Kim et al. 2008).

70 Although the micronutrients (like Mn) present low toxicity, the continuous
71 intake of edible plants with high concentrations of chemical elements can lead to an
72 excess of the elements in the organism and, consequently a significant risk for human
73 health. Manganese uptake by humans is primarily through foodstuffs and its toxic
74 effects occur in the respiratory tract and the brain (Tongesayi et al. 2013). It is also

75 responsible for Parkinson's disease (Powers et al. 2003; Lucchini et al. 2007; Sherzai et
76 al. 2016).

77 The aims of this work were to evaluate: i) the soil–plant system of two species of
78 food crops growing, in greenhouse and microcosm conditions, on soils from the
79 abandoned Mn and Fe mines (Rosalgar and Ferragudo), and ii) the concentration of
80 some potentially toxic elements in the same food crops growing in these soils and the
81 consequent potential human health risk through their consumption. Two different plant
82 species were selected, *Origanum vulgare* (oregano) and *Brassica rapa* subsp. *pekinensis*
83 (chinese cabbage), due to their short growth cycle. Moreover, oregano is an herb that
84 grows spontaneously in these mine areas, and its use as culinary and medical herbs in
85 South Portugal is very common, while the chinese cabbage is also cultivated in some
86 kitchen gardens.

87 **Material and Methods**

88 *Site description*

89 Two soils collected in the abandoned Fe-Mn mines located in SW of Portugal (Cercal-
90 Odemira region) were studied: Ferragudo (Beja district) and Rosalgar (Setubal district).
91 Both mine areas are included in the Portuguese sector of the Iberian Pyrite Belt (Matos
92 and Martins 2006). The orebody forming vein structures was composed of Fe and Mn
93 oxides in Rosalgar and Fe and Mn oxides as well as Mn carbonates in Ferragudo. The
94 mining activity in Rosalgar began 1867 but the intensive ore exploitation occurred
95 between 1959 and 2001. In Ferragudo, the ore exploitation occurred between 1875 and
96 1975 in both galleries and open pit, being intensively exploited in the forties and fifties
97 of the 20th century. The ending of the mining activity was mainly related to the closure
98 of the company where the ore was processed (Rosa et al. 2013).

99 In the Portuguese sector of the Iberian Pyrite Belt context, Rosalgar and Ferragudo are
100 considered with intermediate level of environmental hazard impact (level number two
101 and three, respectively, where level five is considered with extreme impact) due to the
102 small volumes of mine wastes with total concentrations of potentially toxic elements
103 relatively low, except for Mn (Matos et al. 2008).

104 Nowadays, in Rosalgar mine a cork oak tree system is naturally colonized by several
105 autochthones plant species, some of them aromatic species consumed by local
106 population (like *Origanum vulgare* subsp. *virens* (Hoffmanns. & Link) Bonnier &
107 Layens, *Lavandula stoechas* subsp. *luisieri* (Rozeira) Rozeira, *Calamintha nepeta* (L.)
108 Savi subsp. *nepeta*) (Rossini-Oliva et al. 2019), while in Ferragudo mine a holm oak
109 woodland was implemented with small tree density and cultivated grass species. The
110 soil interventions led to the actual existence of soils developed on different mine
111 wastes and host rocks being classified as Spolic Technosol (IUSS Working Group WRB
112 2015).

113 ***Pot experiment***

114 Composite soils samples, up to 20 cm depth, were collected during spring 2017 in the
115 two abandoned Mn–Fe mines (Rosalgar: 37°46'25.0"N 8°41'59.7"W; Ferragudo:
116 37°38'01.0"N 8°03'50.4"W). These samples were air dried, sieved and the fraction
117 <5 mm was used to a pot experiment. Randomized microcosm experiment, using 1.5 kg
118 of each soil per pot, was carried out in greenhouse and under controlled conditions to
119 evaluate plant growth, nutrients content and their accumulation patterns in chinese
120 cabbage and oregano.

121 Ten commercial seeds of each plant species were sown in each pot and thinned to leave
122 2–3 plants per pot. Five replicates for each soil and species were conducted. All pots
123 were kept at 70% of the maximum water-holding capacity during plants growth.

124 Some parameters associated to the plant growth (Photochemical Reflectance Index
125 (PRI), Normalized Difference Vegetation Index (NDVI), plant height and plant fresh
126 weight) were measure in both species at harvest, and soils and plants were collected for
127 chemical analysis. The PRI and NDVI were measured using a PlantPen model PRI 200
128 and PlantPen model NDVI 300 (Photon Systems Instruments), respectively. The NDVI
129 is an important indicator of chlorophyll content in plants. The PRI is sensitive to
130 changes in the carotenoid pigments, and it is related to the photosynthetic light use
131 efficiency and the rate of carbon dioxide fixation, being used as a reliable stress index in
132 studies of vegetation productivity (Garbulsky et al. 2011).

133 The chinese cabbage plants were harvested after two months and the oregano plants
134 after four months (because its much slower growth rate). Roots and shoots (leaves +
135 stems) of both plant species were collected separately. For oregano, leaves were
136 separated from stems since the last one are not edible. Shoots/leaves were washed with
137 distilled water and roots were washed with tap water and then with distilled water using
138 an ultrasound-assisted bath for 20 minutes (to eliminate the fine soil particles more
139 strongly adhered). The plant material was dried at 60 °C during 48 h and milled to
140 powder homogenously. Plant samples were digested with pure HNO₃ in a DigiPrep
141 digestor and concentrations of Ca, Cu, Fe, K, Mg, Mn, Na, P, S and Zn were analyzed
142 by Inductively Coupled Plasma-Optical Emission Spectrometry (ICP-OES Thermo
143 Scientific Mod. CAP 7000 Series).

144 Initial soil samples (fraction <2 mm) were analyzed with standard soil methodologies
145 (Póvoas and Barral 1992): pH and electrical conductivity (EC) in soil:water suspension
146 (1/2.5, *m/V*); exchangeable cations and cation exchange capacity (CEC) at 1 mol L⁻¹
147 ammonium acetate (pH 7); extractable P and K (Egner-Riehm method) and total
148 nitrogen (Kjeldahl method). Total organic carbon concentration was determined using
149 an Analytikjena analyser (MultiEA 4000 model) after sample combustion at 550 °C.
150 Iron content in both non-crystalline Fe oxides and total Fe oxides (*i.e.* non-crystalline
151 and crystalline phases) was extracted by a single step extraction using Tamm reagent
152 (0.1 mol L⁻¹ oxalic acid and 0.175 mol L⁻¹ ammonium oxalate at pH 3.2) in dark
153 conditions (Schwertmann 1964) and under UV radiation (De Endredy 1963),
154 respectively, and determined by ICP-OES. Manganese in manganese oxides was also
155 determined by ICP-OES after extraction with acidified hydroxylamine hydrochloride at
156 0.1 mol L⁻¹ (Chao 1972).

157 Total soil (fraction <2 mm) concentration of the same elements than in plants were
158 determined using ICP-OES (Thermo Scientific Mod. CAP 7000 Series) after microwave
159 digestion with a two-acid mixture (HNO₃-HCl 1:3 *V/V*) + H₂O₂ addition. The same
160 elements were also analyzed in a soil extractable aqueous solution, composed of a
161 mixture of organic acids (acetic acid + lactic acid + citric acid + malic acid + formic
162 acid at 10 mmol L⁻¹) which simulates the rhizosphere conditions (Feng et al. 2005), and
163 therefore considered as the available fraction of the elements for plants in the soils. This
164 available fraction of the elements was determined in the initial soil and after the plant
165 growth (fraction <2 mm). These extractable soil solutions were analysed by ICP-OES.

166 Quality assurance of chemical analysis in plants was performed using analytical blanks
167 and certified reference material (NCSDC73348). The results obtained for certified
168 materials showed a recovery range from 86 to 103% while procedural blanks were

169 usually below the detection limit. All analyses performed were done in duplicate and all
170 results for plants and soils were calculated on a dry weight basis.

171 **Data analysis**

172 A statistical package (Statsoft package v6.12) was used for statistical data analysis. In
173 the pot assay, comparisons between plant and soils characteristics were performed with
174 the t-test. Relations between plant and soil variables were determined by Pearson
175 correlation analysis. In all cases, the values for $p < 0.05$ were considered as significant.

176 The translocation coefficient (TC, *i.e.* the ratio between element concentrations in aerial
177 part and roots) was calculated for both plant species and assays in order to evaluate the
178 plant capacity to translocate a chemical element from roots to aerial parts. The
179 bioaccumulation coefficient (BC, *i.e.* the ratio between element concentrations in
180 shoots/leaves and the available content in soil) was also calculated to determine the
181 ability of the studied species to uptake and accumulate elements in the aerial part.
182 According to Bu-Olayan and Thomas (2009), the plants may be considered
183 accumulators of an element if TF is >1 and BC >1 . In order to evaluate the potential risk
184 of adverse effects for human health posed by both plant species, the hazard quotient
185 (HQ) for Cu, Fe, Mn and Zn was calculated as follows (USEPA 2000):

$$186 \quad \text{HQ} = \text{ED}/\text{RfD}$$

187 ED (Exposure Dose) was calculated by the equation:

$$188 \quad \text{ED} = \text{MC} \times \text{D}_{\text{food}} \times \text{E}_d \times \text{E}_f / \text{B}_{\text{average weight}} \times \text{T}_e$$

189 where, MC is the measured concentration of metal in the sample (mg kg^{-1} of dry
190 weight), D_{food} is the mean mass of oregano and chinese cabbage being consumed daily
191 by an individual (kg/person/d FW), E_d is the exposure duration (in this study 70 years),
192 E_f is the exposure frequency (day for year), $\text{B}_{\text{average body weight}}$ is the body weight (65 kg),

193 T_e is the average exposure time for non cancer risk ($E_d \times 365$ days) and RfD is the oral
194 reference dose (mg/kg/day) defined as the maximum tolerable daily intake of a specific
195 metal that has not adverse affect The RfD of Cu, Fe, Mn and Zn was reported by
196 USEPA (2011, 2012) . The E_f for chinese cabbage was calculated considering that local
197 population consume this vegetable at a maximum of 150 day/years while for oregano
198 was 180 day/year. Values of $HQ > 1$ implies a potential risk to consumers (USEPA
199 2002).

200 It was assumed, according to the local population use, that the local adult inhabitants
201 consumed an average of 1 g oregano/day and an average of 28 g chinese cabbage/day.
202 The MC metal of plant was converted with a factor of 0.20, because both species
203 commonly contain approximately 80% water content in the aerial part. The risk to
204 human health for a metal was only associated with the direct ingestion of the edible part
205 of the species.

206

207 **Results and discussion**

208 *Soil properties and chemical composition*

209 The two soils properties are given in Tables 1 and 2. The results showed that the soils
210 were different regarding chemical characteristics except for organic C contents (Table
211 1) that, according to their medium texture, can be considered as very high for both soils
212 (INIA-LQARS 2000). Both soils had almost neutral pH, although Ferragudo had a
213 significantly higher value (6.9), and can be classified as non-saline considering the
214 conductivity values that inhibit seed germination and plant growth (4 dS m^{-1}) (Ye et al.
215 2002). The cation exchange capacity (CEC) also varied between soils (Table 1),
216 although both values can be considered low (INIA-LQARS 2000). However, the soil
217 from Rosalgar presented lower CEC than Ferragudo soil, which can be related to lower
218 Mn oxides concentrations (Table 1) and quantity and type of clay minerals, and

219 therefore a smallest capacity to supply nutrients. In fact, the soil from Rosalgar showed
220 a lower content of exchangeable cations than those from Ferragudo (Table 1).

221 Ferragudo soil presented higher extractable P and K contents (Table 1), compared to
222 Rosalgar soil, but the opposite was obtained for total N concentration. It is important to
223 underline the very low value of extractable P and medium values of extractable K (≤ 10
224 mg P kg⁻¹ and 41–85 mg K kg⁻¹; INIA-LQARS 2000) in the Rosalgar soil. Also, the
225 total and available concentration of nutrients, such as Ca, K, P and Mg (Tables 2 and 3)
226 was the lowest in Rosalgar soil. This variation at available nutrients level can be
227 associated to the intervention realized in the soil of each area (namely soil fertilization)
228 due to their use. In Rosalgar mine was established a cork oak tree system naturally
229 colonized by several autochthonous plant species while in Ferragudo mine was
230 implemented a silvopastoral system (especially with cultivated grass species) in the
231 holm oak woodland. In general, and considering the available concentrations of the
232 majority of nutrients (Tables 1 and 3), Ferragudo soil could be considered more fertile
233 than Rosalgar.

234 Iron in total Fe oxides was higher in Rosalgar soil than Ferragudo soil while the
235 opposite was obtained for Fe in non-crystalline Fe oxides. Moreover, Fe was mainly
236 associated to the crystalline Fe oxides fraction of both soils. Manganese concentrations
237 associated to Mn oxides were very high in both soils, especially in Ferragudo soil where
238 it reached ~ 31 g kg⁻¹ (Table 1). This Mn oxides concentration in Ferragudo can also
239 justify the highest CEC values obtained in this soil (Brady and Weil 2008).

240 In general, total concentrations of the studied elements in Ferragudo soil reached the
241 highest values (Table 2). Exceptions were obtained for Fe, S and Zn where Rosalgar
242 soil had higher concentrations than Ferragudo soil. As expected, both soils contained

243 high total content of Mn being higher in Ferragudo soil. The high Fe content obtained in
244 Rosalgar soil (Table 2) is a consequence of the highest richness in Fe in the Rosalgar
245 Fe–Mn mineralization (mainly composed of Fe and Mn oxides containing 43% of Fe
246 and 8% of Mn; Matos et al. 2013) than in Ferragudo where the mineralization rich in
247 carbonates and oxides of Mn produced 30% and 50% of Mn, respectively (Almeida and
248 Fernandes 1948; Matos and Rosa 2001).

249 The total concentrations of Cu, Mn, S and Zn in Rosalgar soil as well as Fe in
250 Ferragudo soil exceeded the European topsoils baseline values (Salminen et al. 2005),
251 as well as the average concentrations in non-contaminated soils from the region (Abreu
252 et al. 2012) indicating an enrichment of these elements in the soils by mining activity.
253 Considering agricultural purposes and total concentrations, both soils also exceeded the
254 Canadian Soil Quality Guideline values for Cu and Zn (values should be lower than 63
255 mg Cu kg⁻¹ and 200 mg Zn kg⁻¹; CCME, 2006) and the maximum allowable
256 concentration for Mn (1500–3000 mg kg⁻¹; Kabata-Pendias 2011).

257 The elements concentrations (except for Ca) in the available fraction of both soils were
258 small representing lower than 5% of the total concentrations (Table 3). Nonetheless, the
259 availability of Ca was significant representing 15 and 45% of the total concentration for
260 Ferragudo and Rosalgar soil, respectively. The available concentrations of Ca, P and
261 Mg were the highest in Ferragudo soil, agreeing with the tendency obtained for total
262 concentration of the same elements. However, higher concentration of Fe, K, Mn and
263 Zn in the available fraction was obtained in Rosalgar soil (Table 3). It should be noted
264 the great concentration of Mn in the available fraction of both soils (Table 3), being
265 higher in Rosalgar than Ferragudo, whereas the opposite was obtained for total Mn
266 concentrations. Total Mn contents in soils worldwide vary from 411 to 550 mg kg⁻¹
267 (Kabata-Pendias 2011), which are lower than the Mn concentration in the available

268 fraction of Rosalgar and Ferragudo soils (1053 mg kg⁻¹ and 873 mg kg⁻¹, respectively).
269 The ability of Mn to form anionic complexes and complexes with organic ligands may
270 contribute to increase Mn concentration in the soil solution (Kabata-Pendias 2011). The
271 available Fe content was very low in Ferragudo (Table 3, 6.88 mg kg⁻¹) but high in
272 Rosalgar (75.8 mg kg⁻¹).

273 *Concentration of the elements in soil available fraction before and after plant growth*
274 *under pot experiment*

275

276 In Table 3, the concentration of the elements in the available fraction of the soils before
277 and after the plant cultivation are presented. Each plant species affected differently the
278 pH and the final chemical composition of the soils. Oregano growth led to a
279 decrease of the pH in both soils, while chinese cabbage did not change the soil pH.

280 Nonetheless, elements availability varied with plant species, soil and element. In fact,
281 the available concentrations of most elements were the lowest after chinese cabbage
282 growth (Ferragudo: Ca, K, Mn, P and S; Rosalgar: Fe, K, Mn and Zn). This variation
283 can be associated to the rhizosphere conditions created by each plant species and/or
284 behaviour and nutritional needs of the plants (Dakora and Phillips 2002; Cesco et al.
285 2010). In general, oregano growth contributed to an increase of the availability of most
286 of the elements in both soils (Ferragudo and Rosalgar: Cu, K, Mg, P and S; Ferragudo:
287 Zn). Manganese contents decreased after both species growth and this pattern was also
288 observed for Fe in Rosalgar and for Ca in Ferragudo. It is interesting to underline the
289 pH change after oregano cultivation in both mine soils (Table 3). This species induced a
290 slightly decrease of soil pH, which can explain the increase of the availability of most of
291 the studied elements (Table 3). Changes in soil pH induced by plant roots may be
292 produced by differential cation/anion uptake, root respiration, H⁺ release (carboxylates)

293 or redox processes at the rhizosphere (Hinsinger et al. 2003) and, consequently, affect
294 the elements availability (Kabata-Pendias 2011). Plant root exudates, whose
295 composition varies with plant species or varieties, associated to microorganisms' action
296 and the soil conditions in which the plant grows can also contribute to the availability of
297 the elements to the plants (Kabata-Pendias 1993; Hinsinger et al. 2003).

298

299 *Soil-plant system, growth and elements accumulation in food crops under pot*
300 *experiment*

301 Oregano plants presented good development and without visual symptoms of elements
302 deficiency or toxicity in both soils, while chinese cabbage showed reddish coloured
303 leaves when grown in Rosalgar soils, that is a common symptom of P deficiency (Lee et
304 al. 1996) (Figure 1). Differences in the plant growth (measured through fresh biomass
305 weight) in the two studied soils were obtained only in chinese cabbage plants (Table 4).
306 Moreover, the cabbage plants growing in Ferragudo soil exhibited the best growth, in
307 terms of plant weight and height (Table 4). This fact can be associated to the better
308 fertility of Ferragudo soil than Rosalgar soil (Table 1) namely higher contents of
309 available cations, as Ca, Mg and P (Table 3).

310 The photosynthetic performance indexes PRI and NDVI measured on cabbage
311 leaves were not statistically different between plants growing in the two soils (Table 4).
312 No correlations were found between PRI and NDVI, plant biomass or plant height.
313 Similar results were observed for oregano plants (Table 4). In both soils, the two species
314 showed a mean value of NDVI that can be considered as normal for healthy plants (0.2–
315 0.8; Rouse et al. 1974). The NDVI is a good estimator of the spatial variability of
316 carbon uptake (Garbulsky and Paruelo 2004) and PRI is a good index to estimate
317 photosynthetic efficiency (Gamon et al. 1992; Garbulsky et al. 2011). Thus, the present

318 results indicated that cabbage and oregano plants growing in both soils had similar
319 photosynthetic efficiency.

320 The plant analysis (Table 5) indicate that both species growing in metal enriched
321 soils have high element concentrations, mainly Fe, Mn, S and Zn, in the leaves even if
322 elements uptake and translocation change according to plant species and ecotype (Abreu
323 et al. 2014a; He et al. 2015). Regardless the high concentrations of those elements in
324 above ground part of both species, their levels were different. Significant differences in
325 metal accumulation (Cd, Hg, Pb, Zn) have also been found in different vegetable
326 gardens (e.g. tomato, been, carrot) near an abandoned lead/zinc mine (Sipter et al.
327 2008). Manganese concentrations in the leaves of oregano and shoots of chinese
328 cabbage grown in both soils (Table 5) were greater than the normal range for mature
329 leaf tissues (30–300 mg kg⁻¹; Kabata-Pendias 2011) but non-toxic for cattle and other
330 domestic animals (400-2000 mg kg⁻¹, Chaney 1989; National Research Council 2005).
331 Generally, Mn is rapidly taken up, when it occurs in soils in available forms, and
332 translocated within plants (Kabata-Pendias 2011). Nonetheless, the highest Mn
333 concentrations in the available fraction in Rosalgar soil (Table 3) were not reflected in
334 the Mn concentration of the plants shoots. In fact, elements concentrations in plants
335 cannot be the direct result of the available soil fraction of the element since other
336 processes can occur. These processes can be related to the interaction between elements
337 (antagonism and/or synergism) from available soil fraction and/or potential uptake
338 restriction and low extent uptake of elements by plants when their concentrations in soil
339 are high (Abreu et al. 2014a).

340 Iron concentration in the aerial part of both species (Table 5) was higher in
341 plants growing in Rosalgar soil than in Ferragudo soil, and it was elevated considering
342 the critical value referred by Markert (1996; 500 mg Fe kg⁻¹), reflecting the higher total

343 and available Fe content in Rosalgar soil (Tables 2 and 3). The Fe concentration
344 measured in aerial part of both species grown in Rosalgar soil was much higher than the
345 maximum tolerable value for some domestic animals like cattle, sheep and poultry (500
346 mg kg⁻¹, National Research Council 2005) and higher than the range considered normal
347 for plants (50–250 mg Fe kg⁻¹, Srivastava and Gupta 1996). Nonetheless, both species
348 tolerated well this high Fe concentration and visual toxicity symptoms were not
349 displayed. Plants tolerate Fe excess by different mechanisms, which include the
350 oxidation and immobilization and/or exclusion of soluble Fe by roots (Kabata-Pendias
351 2011).

352 The P concentration in cabbage shoots grown in Rosalgar soil was significantly
353 lower than when grown in Ferragudo soils explaining the visual deficiency symptoms
354 already referred (Fig. 1). The low P content in the shoots of cabbage (Table 5) might be
355 due to the high level of Fe, element that may interfere with P uptake (Zheng et al. 2009;
356 Kabata-Pendias 2011; Rai et al. 2015), and/or the low contents of available P in
357 Rosalgar soil (Table 1). The symptoms of P deficiency were not observed in oregano
358 probably because its roots were able to acidify the rhizosphere soil (Table 3), increasing
359 de availability of P in the soil, which can be taken up by the plants.

360 In the cabbage a positive correlation was found between plant P and fresh
361 biomass ($r=0.93$) while a negative one was found between fresh biomass with plant Fe,
362 S and Zn concentrations in shoots ($r=-0.84$; $r=-0.95$; $r=-0.93$ respectively).

363 Shoots of cabbage and leaves of oregano showed normal Cu concentrations
364 (Kabata-Pendias 2011). Cabbage plants grown in Rosalgar soil showed higher Zn and S
365 concentrations (Table 5) than normal values in edible part of food plants (Kabata-
366 Pendias 2011) and in the reference plant for S (Markert, 1996), but not reaching Zn

367 toxicity levels for cattle, sheep and chicken (300-1000 mg kg⁻¹, Chaney 1989) or other
368 domestic species (swine and poultry) (500-1000 mg kg⁻¹, National Research Council
369 2005). Plants growing in Rosalgar soil showed higher Zn concentration than plants
370 growing in Ferragudo, which is in agreement with the highest concentration of this
371 element in the total and available fraction of this soil (Tables 2 and 3).

372 Potentially toxic elements may be immobilized in the roots or translocated into
373 the shoots (Abreu et al. 2014a; Baker 1981). Root uptake and transport into shoots are
374 influenced by many factors such as the total and available contents of potentially toxic
375 elements, composition of the rhizosphere exudates, environmental conditions and
376 particular plant traits (Kalis et al. 2008; Soriano-Disla et al. 2014). Moreover, some
377 plants have developed different tolerance mechanisms to exclude toxic elements and
378 avoid harmful intracellular levels (Baker et al. 2010; Krämer, 2010; Abreu et al. 2014a;
379 Rossini-Oliva et al. 2018). The studied species stored in their roots most of the chemical
380 elements as the calculated TF was <1 (McGrath and Zhao 2003) (Fig. 2a-b), except for
381 S in cabbage growing in the Rosalgar soil. These species were able to immobilize
382 potentially toxic elements in their roots and limit their translocation to the leaves
383 (*excluders*, according to Baker 1981), this is an important feature since the edible part of
384 the plants is the aerial part in both species.

385 The bioaccumulation coefficient (BC) was higher than unit for Cu, Fe, S and Zn
386 in both species, except for Mn (Figure 3). Although no significant correlations were
387 found between elements concentrations in the available fraction of the soils and in
388 plants, the BC values seem to indicate that both species were able to uptake elements
389 from the soils when they occur in the available form, except for Mn.

390 ***Risk of adverse health effects from human consumption of cabbage and oregano***

391 The transfer of potentially toxic elements from soils to plants is a great concern
392 when the soils are used for cultivation purposes (Chojnacka et al. 2005). The studied
393 species showed high concentrations of Fe and Mn in aerial part (Table 5) and,
394 consequently, their consumption might represent a risk for humans. According to
395 Kabata-Pendias (2011), the Mn content in plant foodstuffs is variable being the highest
396 values obtained in beet roots (36–113 mg kg⁻¹), whereas in edible parts of vegetables
397 the Fe concentration reach mean values from 29 to 130 mg kg⁻¹. The Mn and Fe
398 concentrations found in oregano and cabbage shoots/leaves were much higher than
399 those reported above (Table 5).

400 According to the calculated values of hazard quotients (Table 6), the
401 contribution of the chinese cabbage and oregano to the daily intake of Cu, Fe, Mn and
402 Zn would not represent risk of adverse effects to consumers health since they were
403 below the unit. Similar results were found in vegetables produced in kitchen gardens
404 located in the vicinity of abandoned pyrite mines from the Portuguese sector of the
405 Iberian Pyrite Belt (Alvarenga et al. 2014) as well as for *Arbutus unedo* leaves collected
406 in Panasqueira mine area of north Portugal (Abreu et al., 2014b).

407

408 **Conclusions**

409 The past mining activities in Rosalgar and Ferragudo caused soils contamination
410 by Fe and Mn, reaching the total concentration of Mn values up to 122 g kg⁻¹ of soil.
411 Soils from both abandoned mines are suitable for the cultivation of cabbage and
412 oregano, since the calculated hazard quotient of Cu, Fe, Mn and Zn were lower than unit
413 and, therefore, its consumption can be considered safe for alimentary uses.

414 Since the soils are still quite enriched with Mn with high levels in the available
415 fraction, which can be easily uptake by plant food species, attention must be paid by the

416 inhabitants of the region before use these soils for plant food cultivation. More studies
417 using other edible species should be undertaken in both areas. With this information, the
418 best food plant species considering the lower health risk associated to metals excess
419 might be suggested to the local populations in order to produce food plants with lower
420 health risk.

421

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