

Viewing emerging human infectious epidemics through the lens of invasion biology

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42

43 **Abstract**

44 Invasion biology examines species originated elsewhere and moved with the help of humans, and
45 their impacts on biodiversity, ecosystem services, and human well-being. In a globalized world, the
46 emergence and spread of many human infectious pathogens are quintessential biological invasion
47 events. Some macroscopic invasive species themselves contribute to the emergence and
48 transmission of human infectious agents. We review conceptual parallels and differences between
49 human epidemics and biological invasions by animals and plants. Fundamental concepts in
50 invasion biology regarding the interplay of propagule pressure, species traits, biotic interactions,
51 eco-evolutionary experience, and ecosystem disturbances can help to explain transitions between
52 stages of epidemic spread. As a result, many forecasting and management tools used to address
53 epidemics could be applied to biological invasions and *vice versa*. Thus, we advocate for increasing
54 cross-fertilization between both disciplines to improve prediction, prevention, treatment, and
55 mitigation of invasive species and infectious disease outbreaks, including pandemics.

56

57 **Keywords:** biosecurity, immunology, introduced species, One Health, SARS-CoV-2

58

59 Introduction

60 Invasive species – i.e. non-native (alien, exotic) species that have been introduced to new regions
61 by humans, form self-sustaining populations and spread rapidly from the sites of introduction
62 (Blackburn et al. 2011, Essl et al. 2018) – can have enormous impacts on the environment, the
63 economy and human well-being (Vilà and Hulme 2016, Pyšek et al. 2020). Invasion biology, a
64 discipline examining the ecological, evolutionary and anthropogenic processes involved in the
65 spread and impact of non-native species, has mostly focused on free-living, conspicuous
66 macroscopic species, which spread is observable and easy to track. In contrast, the invasion
67 dynamics of parasites and pathogens have received less attention, except for those causing
68 damage to agriculture, forestry and livestock (but see Mallon et al. 2015, Thakur et al. 2019, Pyšek
69 et al. 2020). More recently, the focus has expanded to include pathogens that affect wildlife
70 (Hatcher et al. 2012, Dunn and Hatcher 2015, Roy et al. 2017). The emergence and spread of
71 human infectious agents that rapidly increase in incidence and geographic area can also be viewed
72 as a biological invasion, but have rarely been treated as such (Hatcher et al. 2012, Nuñez et al.
73 2020) – although many studies have described the direct and indirect human health impacts of
74 biological invasions, including those involving the introduction of human pathogens (Hatcher et al.
75 2012, Rabitsch et al. 2017).

76 A human pathogen can spread beyond its historical range and become invasive, usually as a result
77 of the movement of infected human hosts. In addition to humans assisting the spread of invasive
78 animal and plant species, invasive species themselves can facilitate the large-scale propagation of
79 human pathogens and epidemics by acting as vectors or reservoir hosts of emerging human
80 pathogens, or by providing habitat for them (Fig. 1). Indeed, 16 % of the IUCN list of 100 of the
81 World's Worst Invasive Alien Species (Lowe et al. 2000) promote the spread and impact of human
82 pathogens (Table 1). Invasive insects are the most frequent vectors of pathogens causing human
83 diseases (Lounibos 2002). For example, the tiger mosquito (*Aedes albopictus*) has spread to all
84 inhabited continents through trade and is a vector of several infectious pathogens including those
85 causing dengue fever, yellow fever, West Nile Virus (WNV) and Chikungunya (Gratz 2004, Enserink
86 2008). Another group of invasive mosquitoes are some *Anopheles* spp., the most important
87 vectors of *Plasmodium* spp., the blood parasites that cause malaria (Lounibos 2002, Takken and
88 Lindsay 2019). Invasive vertebrates such as rodents are frequent reservoirs or intermediate hosts
89 of human pathogens (Hatcher et al. 2012, Hulme 2014a). Finally, invasive species, particularly
90 plants, can create habitat conditions conducive to local proliferation of vector or reservoir hosts
91 (Mack and Smith 2011, Rai and Singh 2020). For example, the invasive bush *Lantana camara*
92 attracts and provides refuge for tsetse flies away from river courses and close to villages,
93 promoting sleeping sickness epidemics (Syed and Guerin 2004). Similarly, water hyacinth,
94 *Eichhornia crassipes*, forms dense mats that provide breeding habitat for mosquitoes that transmit
95 *Plasmodium* (causative agent of malaria), *Filifilaria immitis* (filariasis) or *Flaviviruses* (dengue
96 fever) (Mack and Smith 2011). These cases exemplify the enormous diversity of combinations of
97 native-invasive pathogen, host and reservoir that are possible (Fig. 2), suggesting myriad potential
98 roles of invasive species in the ecology and global spread of pathogens (Rabitsch et al. 2017).

99 Both biological invasions and infectious diseases are becoming more prevalent and widespread
100 with globalization. Both phenomena share common drivers of introduction and spread (Mack et al.
101 2000, Jeschke et al. 2013). In biological invasions, there has been a substantial amount of research
102 on species traits conferring invasion potential (i.e. invasiveness), on the vulnerability of the
103 ecosystems to be invaded (i.e. invasibility), and on the role of environmental conditions facilitating
104 or preventing spread (Pyšek et al. 2012). Similarly, research on infectious diseases mainly focuses

105 on understanding factors influencing the ability to establish persistent infections and cause
106 disease (i.e. virulence) and on the transmission from host to host (i.e. transmission), why some
107 microorganisms and specific strains cause disease, which individuals and human populations are
108 more susceptible to infection, and how/which environmental conditions affect pathogen spread
109 (Horrocks et al. 2011). However, because research on invasions and epidemics are approached by
110 different disciplines, the bodies of literature and terminology are usually separated (Box 1). An
111 exchange and cross-fertilization between both research domains is needed to advance the
112 prevention, treatment and adaptation of their impacts (Conn 2009, Ogden et al. 2019, Hulme et al.
113 2020, Nuñez et al. 2020).

114 The introductions of invasive species and human pathogens have been described as co-occurring
115 phenomena caused by the transport of species, including people, during early European
116 colonization of the Americas, and some African and Asian territories during the XV-XVII centuries
117 (Crosby 2004, Spinage 2012). There are historical descriptions, for instance, of how these human
118 migration patterns led to disease outbreaks in the new territories (e.g. influenza, smallpox, and
119 measles). However, despite epidemiology having acknowledged the ecological aspects of
120 infectious diseases since its start, and invasion biology having some of its foundations in the
121 spread and impacts of pathogens – e.g. Elton (1958) highlighted several examples of plant, animal
122 and human pathogens as biological invasions, the formal interaction between both disciplines is
123 quite recent and currently limited: the number of publications bridging the two disciplines is
124 several orders of magnitude lower than in each field separately (Fig. 3).

125 Approaches such as One Health, EcoHealth, Planetary Health and One Biosecurity emphasizes the
126 links between human health, environmental health, and the health of plants and animals (Ogden
127 et al. 2019, Hulme 2021). Following this principle, there have been recent attempts to cross-
128 fertilize research on biological invasions and human infectious diseases both from conceptual and
129 methodological perspectives. While marked differences do exist in the ecology and evolution of
130 human pathogens and free-living macroscopic invasive species, including issues of host specificity,
131 immunity as well as the temporal and spatial scales of interactions, opportunities exist to bring
132 these disciplines together under a common framework (Lewis et al. 2016, Hulme et al. 2020).
133 Previous reviews have mostly focused on the stages of invasions and emerging infectious
134 pathogens, especially those that also affect wildlife (Hatcher et al. 2012, Jeschke et al. 2013, Dunn
135 and Hatcher 2015, Roy et al. 2017); on the role of invasive species as vectors and/or reservoirs of
136 pathogens worldwide (Hulme 2014a, Rabitsch et al. 2017); or on spatial dynamics (Hulme et al.
137 2020). Most of these interdisciplinary approaches have been on particular taxa, habitats or regions
138 (Crowl et al. 2008, Medlock et al. 2012, Conn). Yet, a detailed review of the parallels between
139 scientific approaches to invasions and human epidemics is still missing.

140
141 Given increasing rates of emerging infectious pathogens and biological invasions worldwide, and
142 the on-going global health crisis caused by the novel coronavirus SARS-CoV-2, the need for
143 integrative and interdisciplinary approaches to biosecurity has never been greater (Nuñez et al.
144 2020, Pyšek et al. 2020, Hulme 2021). Here, we provide a holistic review of key parallels in the
145 conceptual foundations in invasion biology and human infectious epidemics. Specifically, we (1)
146 describe approaches to the study of the pathways of introduction of invasive species and human
147 pathogens; (2) compare the stages and dynamics of the invasion process with those of epidemics;
148 (3) outline well-established hypotheses on the performance and impacts of invasive species, and
149 show their analogues in human pathogens; (4) summarize the usefulness and limitations of
150 forecasting tools; and finally (5) discuss the implications for biosecurity.

151 Pathways of introduction of invasive species and transmission of pathogens

152 With globalization, the numbers of invasive species and human pathogens has increased
153 exponentially in the 20th century, with no sign of saturation (Jones et al. 2008, Seebens et al.
154 2017). Invasive species including pathogens are rapidly transported by the same global networks
155 that move products and people to distant regions, where they are likely to encounter naïve
156 ecological and human communities that have not interacted with them before. For example,
157 dengue virus, the causative agent of dengue fever, is expanding its distribution range and it is now
158 reported in 128 countries. The main factor of its spread is related to climatic change that benefits
159 the *Aedes aegypti* mosquito, the main vector of the virus, and increased human movements
160 between populations; even sporadic indigenous virus transmissions have occurred in previously
161 dengue-free countries (Chomicz et al. 2016). Managing the pathways of introduction of invasive
162 species and infectious pathogens is a prerequisite to implementing effective surveillance, early
163 response and mitigation policies (Essl et al. 2015, Ogden et al. 2019).

164 The Convention on Biological Diversity (CBD) provides a global standard terminology for species
165 introduction pathways that can be classified by six mechanisms: release, escape, transported as
166 contaminant, transported as stowaway, corridors and unaided (Saul et al. 2017). These can be
167 further classified in 44 subcategories that identify their socioeconomic use and purpose of
168 introduction (e.g. horticulture, pet trade, fisheries, game, etc.). Recently, this classification has
169 been applied to thousands of non-native species introduced to Europe and worldwide (Pergl et al.
170 2020). Range-expansion of native species that track environmental changes is an ecological
171 phenomenon that gets often confounded with biological invasions. However, there are major
172 functional, phylogenetic, physiological, behavioural and phenology feature differences separating
173 range-expanding from non-native species (Essl et al. 2019); accordingly, both groups of species
174 deserve to be treated as distinct biogeographic entities (Essl et al. 2020). Range-expanding species
175 (i.e. neonatives) can also cause environmental and health impacts (Wallingford et al. 2020).
176 However, to not increase the complexity of our review, we do not include range-expanding species
177 in this study.

178 In human epidemiology, besides the dichotomy between active and passive introduction of
179 pathogens (Mallon et al. 2015), a classification of pathways to such detail as in biological invasions
180 is currently not available. The term 'pathways of introduction' refers to the movement of the
181 pathogen either as a free-living stages (environmental contamination), or via the original
182 (reservoir) host, the vector or by human hosts. Infected hosts that travel with their newly acquired
183 pathogens to distant places contribute to their geographical spread. Phylogenetic and genomic
184 analyses are important tools used to reconstruct epidemiological origin, history and links among
185 infectious hosts. Genomic surveillance is not routinely used in biological invasions to identify the
186 geographic origin and pathways of introduction of non-native macroorganisms (but see Hamelin
187 and Roe 2020).

188 Transmission of emerging infectious pathogens can also be classified as zoonotic or non-zoonotic.
189 A global analysis suggests that more than 60 % of human emerging infectious pathogens are
190 zoonotic, with 70 % of these originating in wildlife (Jones et al. 2008). The IUCN list 100 of the
191 World's Worst Invasive Alien Species contains twelve species that are reservoirs of pathogens that
192 infect humans (Table 1). The most well-known historical example is the house mouse (*Mus*
193 *musculus*) and the black rat (*Rattus rattus*) as hosts of *Yersinia pestis* causing bubonic plague.
194 Other invasive species include the small Indian mongoose, *Herpestes javanicus*, and the crab eating
195 macaque (*Macaca fascicularis*) as reservoirs for rabies. Zoonoses, by definition, involve pathogen

196 spillover from a vertebrate host to humans, although subsequent human-to-human transmission
197 is sometimes possible. These host-switching events from wildlife reservoir to human can be
198 preceded by an invasion event, e.g., when the reservoir host enters a previously unoccupied area
199 (e.g., wildlife transported to an urban market), or followed by an invasion event, e.g. when
200 infected people travel, with their newly acquired pathogens, to distant places. Zoonotic spillover is
201 seen for multiple pathogens including *Plasmodium* spp. (causative agent of malaria), *Trypanosoma*
202 *brucei* (trypanosomiasis) , *Leishmania* sp. (leishmaniasis), influenza A (flu), Human Immune
203 Deficiency Virus (AIDS), Ebolavirus (Ebola haemorrhagic disease) as well as the new coronavirus
204 related to MERS-CoV and SARS-CoV (Karesh et al. 2012).

205 In invasion biology, prevention requires an analysis of how the invasive species likely will arrive to
206 a new region (primary introduction) and how it spreads subsequently in the surrounding region
207 (secondary spread). This dual pathway classification has seldom been applied in emerging
208 infectious pathogens despite that it is well known that socioeconomic variables (e.g. behavior,
209 income, tourism, military deployment, trade, etc.) can highly influence transmission. An improved
210 understanding of mechanisms that link long- and short-distance pathogen spread with the
211 socioeconomic characteristics of the hosts is essential to prevent and manage epidemics.

212 **Stages and dynamics of invasions and epidemics**

213 There are several distinct terms used to describe processes of invasion and those of an epidemic;
214 but conceptually, the invasion of ecosystems and the infection process at the individual and
215 population level follow essentially the same basic series of stages, i.e. transport/exposure,
216 introduction/infection, establishment/transmission and spread/epidemics, respectively (Jeschke et
217 al. 2013, Dunn and Hatcher 2015, Plowright et al. 2017, Hulme et al. 2020, Nuñez et al. 2020). In
218 both cases, whether a particular invasive species or pathogen is able to pass on to the next stage
219 and has consequences for the receiving ecosystem or host depends on many filters and can be
220 substantially influenced by human interventions (Fig. 4). These stages have used different
221 terminology for invasions and infections, respectively, as indicated below.

222 **Transport/exposure.** International transport of the non-native species by human agency is the first
223 stage of the biological invasion process. Similarly, in emerging infectious pathogens, international
224 movement of hosts (e.g. planes or boats) represents the first contact (or exposure) of humans
225 with infected human hosts. The pathogen may originate in wildlife or domestic vertebrates and
226 spillover to humans either through a vector (e.g. insects) or through direct contact (i.e. zoonosis).

227 **Introduction/Infection.** Following transport, some non-native species are released directly into the
228 wild (e.g. for fishing or hunting purposes) escape from captivity (e.g. pets) or cultivation (e.g.
229 ornamental plants), or move unaided utilizing artificial corridors (e.g. waterways). A pathogen can
230 also be introduced through released and escaped reservoirs or move unaided through air (e.g. air-
231 conditioning) or water (e.g. sewage) infrastructures. For a pathogen, at the individual host level,
232 this is the infection stage where it enters the host body, circumventing behavioral, physical and
233 physiological barriers. Many human infectious pathogens such as Hendra virus, WNV or the strain
234 of Influenza A causing avian flu result from independent spillover from reservoirs with little
235 human-to-human transmission. These outbreaks tend to be short-lived, but nonetheless can have
236 high impact in humans (e.g. the case fatality rate for some avian flu is 60%, Greger 2007).

237 **Establishment/Transmission.** Establishment of an invasive species is the process by which a
238 founding non-native population reproduces, increases in size and becomes self-sustaining in the
239 new range. Invasive species introduced to a new region have to overcome several biotic and

240 environmental barriers imposed by the recipient region and its biota (Blackburn et al. 2011). For a
241 pathogen, at the level of the individual host, this is equivalent to overcoming immunological
242 barriers that allow within-host persistence, its multiplication and transmission to new hosts.
243 Widespread transmission and establishment within a new host population occurs when the basic
244 rate of reproduction (R_0 , the number of secondary cases resulting from each primary case)
245 exceeds 1. The likelihood of the pathogen evolving to become self-sustaining in the human
246 population increases with the spillover rate, the current R_0 and the mutation rate (Antia et al.
247 2003). For example, during the 2013-2016 Ebola virus outbreak, three adaptive mutations in the
248 virus genome occurred that affected the functional activity of various viral proteins increasing its
249 ability to enter human cells, grow and be transmitted (Urbanowicz et al. 2016).

250 **Spread.** Finally, spread is the process by which an invasive species expands its range in the
251 introduced region beyond the area or host population in which it was first established. This
252 matches with the definition of epidemics as the spread of the pathogen to many persons in a
253 locality during a short period. Such an expansion of a pathogen in a human population can occur
254 through increased animal-to-human contacts (spillover) or through human-to-human
255 transmission. For human infectious pathogens, spread can occur anywhere along a gradient from
256 transmission between individuals in a local population, to global transport of infections between
257 populations. Like biological invasions in general, the large scale spread of pathogens follows hub-
258 and-spoke network dynamics, and does not occur homogeneously but rather in discrete,
259 sometimes lengthy jumps, facilitated by human transportation systems such as air travel
260 (Strickland et al. 2015). The most serious outcome of an emerging pathogen is a pandemic – an
261 epidemic occurring worldwide, or over a very wide area, crossing international boundaries and
262 usually affecting a large number of people.

263 Unprecedented opportunities for pathogen spread and transmission are generated by (1)
264 technological advances and social activities driving human mobility, as evident in the movement of
265 millions of humans between continents on a daily basis (Tatem et al. 2006), and (2) with
266 increasingly crowded living conditions and inadequate access to water, sanitation, and health care,
267 in many areas of the world. For example, the first cases of Sars-CoV2 in many countries were
268 associated to business and tourism, whereas subsequent local spread was mainly related to
269 factors such as housing density and occupational exposure (Bassino and Ladmiral 2020). Owing to
270 global transportation networks, introduced organisms – both pathogens and free-living
271 macroscopic species – create satellite outbreaks in distant regions that contribute to exponential
272 rates of spatial expansion.

273 **Rate of spread.** There are temporal and spatial differences in the dynamics of epidemics and
274 invasions. In an epidemic, the speed by which the pathogen can spread is usually faster than the
275 invasion of a free-living macroscopic species (Peterson 2008). The spread of human epidemic
276 pathogens can be explosive. It is generally one to three orders of magnitude faster than for
277 invasive species and plant pathogens (Fig. 5). This is due to their short generation times, high
278 mutational rate and by orders of magnitude higher effective population sizes. Rates of spread of
279 terrestrial flora and fauna are typically in the range of 0.1-100 km/yr (Hulme 2014b, Horvitz et al.
280 2017) with mobile species such as many invertebrates (e.g. forest pest insects) being faster
281 (Roques et al. 2016). In contrast, human epidemic viruses such as Zika, Ebola and West Nile Virus,
282 can spread at rates of 10^3 – 10^4 km/year (Zinszer et al. 2015, 2017, Hadfield et al. 2019), a velocity
283 only reached in some pathogens of marine wildlife (McCallum et al. 2003).

284 These differences in spread velocity matter because they influence the response of the recipient
285 systems in many ways. For instance, rapid range expansion could render phenotypic or genotypic
286 adjustments in recipient populations and communities less likely. Moreover, success in the control
287 of invasive species and infectious pathogen spread is highly dependent on the spatial distribution
288 of introductions (Hulme et al. 2020). Scattered nascent foci of invasive species or infested hosts
289 have the potential to spread more rapidly than one large continuous focus (Moody and Mack
290 1988). The recommendation to detect, isolate and trace every contact of the SARS-CoV-2 infected
291 individual follows this principle (e.g. Pagliari 2020).

292 **Lag times.** This phenomenon has received a fair amount of attention in invasion biology to define
293 the duration between invasion stages, and also between the introduction and the onset of rapid
294 range expansion (Crooks 2005, Rouget et al. 2016, Spear et al. 2021). Lag times are particularly
295 evident in ornamental plant species that only start to spread after several decades of being
296 introduced (Kowarik 1995). Many populations of non-native plants are dependent on repeated
297 introductions and need a long residence time before they form self-sustaining, viable populations
298 (Dlugosch and Parker 2008). Small populations are very sensitive to environmental stochasticity
299 that might limit their survival, reproduction and dispersal during early stages of invasion (Mack
300 2000). There are many cases of non-native species that were unnoticed for a long time and only
301 became invasive as a response to environmental changes.

302 Lag times are also identified in emerging human pathogens, owing to the latency period between
303 infection and disease symptoms that can range from a few days (e.g. SARS-CoV) to years (e.g. HIV).
304 More precise time intervals than for invasions are defined for pathogens in terms of stages of the
305 pathogen life-cycle and disease symptoms (Bar-On et al. 2020). For example, in virus infections,
306 time lags within an individual host are decomposed into (1) the eclipse period as the time to make
307 intracellular virions; (2) the latent period as the time from cell entry until the appearance of the
308 first extracellular viruses; (3) the infectious period (from infection to transmission) and (4) the
309 incubation period (from infection to the emergence of symptoms). The length of these four
310 periods are of paramount importance to slow down and deter the transmission stage to an
311 epidemic spread by establishing quarantine and confinement periods.

312 Many invasive species that are vectors of human parasites are increasing their ranges induced by
313 global warming (Medlock and Leach 2015). Similarly, many infectious diseases are increasing with
314 climate change e.g. by speeding up the life cycle of the pathogens. For example, human and dog
315 infections by *Dirofilaria* nematodes are becoming more frequent in Northern Europe with
316 increasing summer warming that facilitates parasite incubation (Genchi et al. 2011). Recognition of
317 long lag times and the role of environmental changes in invader and parasite dynamics suggests
318 that we need to endorse the precautionary principle: one should assume that any invader and
319 pathogen has the potential for undesirable effects and that lengthy periods of seemingly
320 innocuous behaviour can be a poor predictor of how these organisms will behave in the future
321 (Crooks 2005).

322 **Hypotheses explaining biological invasions and analogues to epidemics**

324 Invasion biology has formulated and tested several hypotheses on why some non-native species
325 go through the stages of the invasion process, whereas others do not (e.g. Catford et al. 2009,
326 Jeschke and Heger 2018). Invasions are influenced by many factors, and these can be grouped into
327 five categories related to propagule pressure, organism traits, biotic interactions, eco-evolutionary
328 experience and recipient system characteristics (Enders et al. 2020). Each of these five categories

329 encapsulates several hypotheses reviewed by Jeschke et al. (2020) and provides a different
330 perspective on the causes of invasion. Here, we explore the potential parallels between biological
331 invasions and human epidemics across the five categories of hypotheses. A detailed dissection of
332 them is presented in the Supplementary Material.

333 **Propagule pressure.** Propagule pressure refers to the frequency and size (i.e. numbers of
334 propagules introduced) of introduction events (Lockwood et al. 2005). A non-native species is
335 more likely to become invasive in a given region if it is introduced multiple times and with higher
336 numbers of individuals. This hypothesis is also applicable to human pathogens both from an
337 individual and a population perspective and at all stages of the infection process. Pathogen
338 pressure is defined as the abundance of pathogens exposed to the human host at a given point in
339 space and time. With increasing pathogen pressure, there is an increasing likelihood that the
340 pathogen will establish and undergo exponential growth within an individual host, reflecting the
341 well-known dose-response curve (Horrocks et al. 2011). The same idea applies to the population
342 level; it is well known that the number of infected individuals entering a population can strongly
343 influence pathogen dynamics (Ostfeld et al. 2008), as can the heterogeneity of pathogen
344 transmission by individuals (Woolhouse et al. 1997) such as the presence of ‘superspreaders’
345 (Lloyd-Smith et al. 2005). That is, the greater the number of infectious (reservoir or human) hosts
346 to arrive in a given locality, the higher the likelihood that the pathogen will establish and spread in
347 the population (Correa-Martínez et al. 2020). This concept of pathogen pressure is also useful to
348 understand the spillover stage in zoonotic diseases. Pathogen pressure depends on the pathogen
349 dynamics in reservoir hosts, pathogen release from reservoir hosts, and pathogen survival or
350 dispersal outside of reservoir hosts (Plowright et al. 2017).

351 **Organism traits.** Some traits – mainly related to growth, reproduction and dispersal rates –
352 explain why some non-native species have higher invasiveness (i.e. intrinsic potential to become
353 invasive). For example, pine species with small seeds and short generation time have higher
354 potential to invade (Richardson and Rejmánek 2011). Likewise, animals such as rats and pigeons
355 are notorious invasive species worldwide, and have key characteristics that form the basis of their
356 establishment to new areas (e.g., they are generalists, have high plasticity to cope with different
357 environmental conditions, and have adapted to urban environments). Some invasive species that
358 are reservoirs or vectors of human parasites also share some of these traits: young age at
359 maturity, large and frequent broods, explosive rate of replication, tolerance to harsh
360 environmental conditions including disturbances, high mobility of at least one life stage and high
361 dispersal strategies (Ostfeld et al. 2014).

362 Similarly to those of invasive species, different life-history traits of human pathogens appear
363 related to their ability to establish persistent infections within individual hosts and their
364 transmission from host to host. Two key traits that affect pathogen fitness are virulence and
365 transmissibility. They are related, among others, to their capacity to invade cells by adhering to
366 specific receptors, the production of exoenzymes and toxins that allow them to colonize specific
367 tissues of the hosts, and their capability to evade the immune system by self-protecting from
368 phagocytosis, exploiting molecules produced by the host or by antigenic variation (Alcami and
369 Koszinowski 2000). Antigenic variation, the production of different variants of a protein implicated
370 in the interactions with the host cells (Palmer et al. 2016) is a similar strategy as the phenotypic
371 variation of invasive species to cope with different environmental conditions (Davidson et al.
372 2011). Host-specificity is another trait that influences pathogen fitness and epidemics. Generalist
373 pathogens, those that can survive in different hosts, are more likely to cause zoonotic spillover

374 (Woolhouse 2002). These pathogens tend to use cell receptors, which are conserved across
375 different host species (Parrish et al. 2008).

376 Rapid evolution can lead to increased invasiveness of non-native species and to higher virulence
377 and transmissibility of pathogens, either native or non-native. Evolutionary changes during the
378 time span of a few centuries can allow plant physiology to adapt to the new climatic conditions of
379 the introduced range (Maron et al. 2007). Similarly, evolved resistance to pesticides also explains
380 high infestation levels of weeds and pests in crops. In humans, the massive use of antibiotic
381 treatments is causing the emergence of novel, resistant bacteria strains. For example, antibiotic
382 resistance is increasing sexually transmitted diseases such as *Neisseria gonorrhoeae* and
383 *Haemophilus ducreyi*, causative agents of gonorrhoea and chancroid, respectively (Ison et al. 1998,
384 González-Candelas et al. 2019).

385 **Biotic interactions.** Interactions between non-native and native species are crucial for
386 understanding invasions. A key point here is that the same non-native species can establish in one
387 ecosystem and not in another, depending on local biotic interactions (Zenni and Nuñez 2013).
388 Probably the most popular example on how biotic interactions shape the invasion process is the
389 enemy release hypothesis, which posits that the absence of enemies in the introduced range is a
390 cause of invasion because introduced species left their pathogens, parasites and predators behind
391 when colonizing a new ecosystem (Maron and Vilà 2001, Keane and Crawley 2002). The natural
392 enemies for pathogens are virophages and bacteriophages of the human microbiota (Dalmasso et
393 al. 2014). Most probably, in zoonoses, when pathogens jump from their original animal host to a
394 human host, virophages and bacteriophages in humans do not identify and act against the new
395 pathogen.

396 Mutualistic interactions between invasive and native species – e.g. animal mediated pollination,
397 seed dispersal and symbioses between plant roots and microbiota– can be disruptive for the
398 native species but highly beneficial to the integration of the invasive species in the recipient
399 ecosystem (Richardson et al. 2000). A similar situation in emerging pathogens is the case of co-
400 infections among pathogen or parasite species or strains/clones of the same species. A clear case
401 is HIV which makes the host susceptible to a range of other pathogens. The outcome of biotic
402 interactions can be antagonistic (competition and superparasitism), neutral but also mutualistic
403 (Griffiths et al. 2011). These interactions have significant epidemiological clinical and evolutionary
404 implications since they affect the susceptibility of the host to subsequent infections as well as
405 pathogen virulence and transmissibility. For example, given the tradeoff between Type 1 and Type
406 2 immune responses induced by micro and macroparasites, co-infection with endemic helminth
407 infections has been predicted to increase the severity of SARS-CoV-2 (Bradbury et al. 2020). Even if
408 pathogens do not interact, death of co-infected hosts can decrease the fitness of individual
409 pathogens (Hamelin et al. 2019).

410 **Eco-evolutionary experience.** A long-standing hypothesis explaining the impact of biological
411 invasions is that species introduced to ecosystems lacking functionally or phylogenetically similar
412 natives are more likely to disrupt communities, because these communities lack effective
413 physiological, morphological, or behavioral adaptations – that is, they are naïve to such invasive
414 species (Diamond 1986, Ricciardi and Atkinson 2004). Eco-evolutionary naïveté explains why
415 native prey populations typically suffer greater damage from introduced consumers than from
416 native consumers (Salo et al. 2007, Paolucci et al. 2013, Saul and Jeschke 2015, Anton et al. 2020).
417 The hypothesis also predicts heightened sensitivity of insular ecosystems, such as islands and
418 lakes, to the effects of invasions. For example, oceanic island endemisms have been devastated by
419 non-native mammalian predators and herbivores, largely because most island biota evolved in the

420 absence of such species (Russell et al. 2017). The eco-evolutionary experience hypothesis also
421 applies to sessile organisms such as plants (Mack 2003). A novel plant life form in a new range can
422 affect its invasiveness as well as the magnitude of its impact on native vegetation. For example,
423 pines originated in the Northern hemisphere, and their impacts are larger when introduced in the
424 Southern hemisphere where not only the taxon, but also the life form, is completely new in many
425 communities it invades. Differences in the mechanisms of pine impacts among regions are not well
426 known, but might be related to different biogeochemical effects on the soil to which the native
427 plants are not adapted (Davis et al. 2019).

428 Analogously, immunological naïveté to infectious agents contributes to a large public health toll.
429 Historical exposure and co-evolution between hosts and pathogens, typically lowers its severity
430 within a population or region. In the case of malaria, for example, human populations at higher
431 altitudes in the East African highlands are more susceptible to infection and suffer more severe
432 symptoms compared to populations in lower-latitude areas, where they have had greater and
433 longer exposure to the parasite (Pascual et al. 2008). Paralleling invader-community interactions,
434 the more experienced hosts within pathogen-host interactions offer resistance to infection and
435 experience less harm (Domínguez-Andrés and Netea 2019). Influenza pandemics, for example,
436 cause lower mortality in populations that have had some evolutionary exposure (immunological
437 memory) from previous pandemics (Horimoto and Kawaoka 2005). However, pandemics typically
438 involve novel viruses arising from antigenic shift or zoonotic spillover, which preclude human
439 populations from having immunity. For example, the emergence of swine flu in 2009 resulted from
440 recombination of segments of influenza A from pigs, birds and human hosts, creating a strain with
441 the ability to target human respiratory receptors, but with a novel antigenic profile (Smith et al.
442 2009). Within a human population, naïveté decreases as more people are infected. Once some
443 immunity develops within the host population, the R_{eff} (effective reproduction number) will
444 decline, a phenomenon that is exploited in the use of vaccination programs.

445 **Recipient system characteristics.** Pristine native ecosystems with high biodiversity often resist
446 invasion via a process termed biotic resistance (Levine and D'Antonio 1999). Similarly, ecosystems
447 with high animal and plant diversity has consistently been shown to reduce the transmission of
448 infectious pathogens due to reduced chances to encounter hosts (Keesing et al. 2010, Myers et al.
449 2013, Johnson et al. 2015). In the case of pathogens, the limitation in the establishment of a new
450 microorganism when the invaded community has high species diversity is rooted on the
451 microbiostasis concept (Mallon et al. 2015). Plant and microbe experiments using synthetic
452 communities from low to high diversity species assemblages have shown that invader
453 establishment and abundance increase in depauperate communities (Zavaleta and Hulvey 2004,
454 Eisenhauer et al. 2013). In humans, the microbiome is a barrier to pathogens (Penders et al. 2013).
455 The relationship between alterations of the microbiome composition and diversity with
456 antimicrobial resistance is a topic of major research interest in biomedicine.

457 The diversity-invasion relationship can be uncoupled with increased availability of resources.
458 Disturbances offer windows of opportunity for invasive species by disrupting biotic resistance and
459 thus freeing resources (Hobbs and Huenneke 1992, Jeschke and Heger 2018). Disturbances can
460 also pre-adapt plants and animals for colonization of human-dominated ecosystems (Hufbauer et
461 al. 2012). The same appears to be true for epidemics. After natural disasters there are numerous
462 opportunities for pathogen outbreaks driven by people crowding, poor sanitation leading to
463 increased exposure to pathogens and malnutrition increasing susceptibility to disease (Watson et
464 al. 2007). Altered ecosystems by deforestation, agricultural expansion, harvesting of bush meat,
465 and other anthropogenic disturbances can facilitate the emergence of zoonotic pathogens

466 (Keesing et al. 2010) and create opportunities for spillover (Jones et al. 2013). For example, in
467 Australia and Asia, changes in land use and habitat loss have changed the ecology and behavior of
468 fruit bats that are natural reservoirs of Nipah and Hendra viruses increasing spillover chances to
469 humans (Kessler et al. 2018). At the level of the individual host, altered immunological or
470 physiological conditions affect susceptibility to infection and the severity of the disease (Plowright
471 et al. 2017). For example, certain medicines, immunosuppression caused by co-infections or
472 medical/surgical procedures, nutrition, and autoimmune diseases offer windows of opportunity
473 for infection.

474 In sum, the invasions and epidemics are driven by historical, intrinsic and extrinsic characteristics
475 of the species/pathogens such as the abundance of propagules, frequency of the introduction
476 events, attributes of interacting species/strains, and characteristics of the invaded or host system.
477 The interplay and importance of these factors are highly context-specific and highly dependent on
478 the spatial scale of analysis (von Holle and Simberloff 2005, DeVincenzo et al. 2010).

479 **Forecasting biological invasions and human epidemics**

480 Forecasting the occurrence and timing of future invasions is challenging owing to the high intrinsic
481 uncertainty associated with many potential origins, trends and pathways of introduction,
482 particularly for new invasive species that have not been previously recorded as problematic
483 (Seebens et al. 2018). Similar challenges apply to emerging human pathogens. The analysis of past
484 events has facilitated the identification of potential spatio-temporal patterns of invasion and
485 pathogen emergence, which allows prioritizing surveillance efforts on the most likely threats and
486 vulnerable areas. For instance, invasive species are dominated by plants (e.g., lantana, kudzu,
487 water hyacinth), are dispersed by human activities that involve transportation and commerce,
488 their global spread is largely driven by climate, land use and environmental degradation (Pyšek et
489 al. 2020). Likewise, most pandemics—e.g., HIV, severe acute respiratory syndrome, Covid19—
490 appear to have originated in animals, are caused by viruses, and their emergence is driven by
491 ecological, behavioral, or socioeconomic changes (Morse et al. 2012). For example, a study in 2013
492 reported the presence of a large reservoir of SARS- like coronaviruses in horseshoe bats which,
493 together with the custom of eating non-native mammals in southern China, was already alerting
494 epidemiologists to the risk of a human epidemic (Ge et al. 2013). Some of the differences and
495 common challenges shared between the study of biological invasions and emerging pathogens are
496 outlined below and summarized in Table 2.

497 **Data.** Problems of low data quality and uneven sampling effort are common for both fields. Data
498 on species occurrence, used in invasion studies, is strongly biased geographically and
499 taxonomically (Pyšek et al. 2008), with invasive pathogens being specially understudied (Roy et al.
500 2017). Similarly, in an epidemic the quality of data on the number of infections, deaths, tests, and
501 other factors needed for robust modelling is often limited by under-detection, reporting delays,
502 and poor documentation (Jewell et al. 2020). Recent methods for estimating occupancy dynamics
503 under imperfect detection are promising to reduce the uncertainty of predictions, particularly for
504 host-pathogen systems (Bailey et al. 2014). Both fields would benefit from common monitoring
505 systems and open data platforms to facilitate standardization and data sharing.

506 **Indicators.** The focus of invasive species forecasts is usually the likelihood of species
507 presence/absence, and therefore the total number of invasive species that could invade an area,
508 rather than their potential abundance or impacts. In contrast, the most important indicator used
509 to assess the spread rate of an epidemic is R_0 . The larger the value of R_0 , the harder it is to control
510 an epidemic. The demographic analogue for invasive species is λ , the population rate of

511 change (Caswell 2000). When applied to population dynamics, a value of $\lambda < 1$ will similarly lead to
512 population decline and ultimately extinction. In both cases, however, any value that is even only
513 slightly above 1 will lead to population growth of the invasive species or pathogen, until other
514 limiting factors set in. Calculating λ for invasive species is knowledge and data intensive and
515 becomes complicated because individuals can reproduce and disperse for many years, and survival
516 depends on multiple factors that can be deeply affected by environmental gradients (Krkosek and
517 Lewis 2010). This has limited the use of population models to rather few invasive species with
518 enough information, frequently plants and invertebrates (Buchadas et al. 2017). Considering the
519 close relationship between biological invasions and epidemics, the use of common spatio-
520 temporal indicators of risk would provide insights into their inter-relationship and common
521 underlying drivers (Allen et al. 2017, Hulme et al. 2020).

522 **Models.** Among the multiple modelling techniques employed in invasion studies, Species
523 Distribution Models (SDM) have become the gold standard method to identify the habitats or
524 geographical areas most prone to be invaded under current and future climate change scenarios
525 (e.g. Thuiller et al. 2005, Bradley 2010). In contrast, from the 174 infectious pathogens with
526 comprehensive geographical information, only 7 (4%) had been comprehensively mapped
527 including Dengue, Lassa, Mayaro, Monkey pox viruses, and the malaria parasites *Plasmodium*
528 *falciparum* and *P. vivax* (see Hay et al. 2013). This is likely because of the complex characteristics
529 of the host-pathogen system, which requires a re-evaluation of the traditional biogeography
530 framework (sensu “pathogeography”, Murray et al. 2018). In this sense, a key difference between
531 invasive species and epidemics originated by pathogens with complex life-cycles is that the
532 distribution of the pathogen is defined by the joint distributions of all species involved in its
533 transmission cycle as dictated by the suitable ecological conditions and dispersal limitations for
534 each. Consequently, models should integrate the large biogeographic factors that condition the
535 presence of vectors, hosts and reservoirs, with the microscale characteristics of hosts that allow
536 the survival, reproduction and transmission of pathogens (Johnson et al. 2019). Multi-species joint
537 distribution modelling (Pollock et al. 2014) could be thus interesting for infectious diseases,
538 particularly for multi-host pathogens or to investigate the interaction among pathogens.
539 Furthermore, a better understanding the global distribution of mammal zoonotic hosts could help
540 predict future hotspots of zoonotic pathogen emergence (Han et al. 2016).

541 However, not all pathogens are appropriate for SDM modelling depending on their life cycle,
542 host(s) and spread mode. Instead, dynamic models explicitly represent the key population groups
543 and central processes of epidemic spread. Dynamic models can be used to predict future trends of
544 pathogen spread, although the uncertainty of exponential processes such as epidemics is
545 considerable. Dynamic models have been increasingly used for invasive species since the late
546 1990s, mostly focused on plants such as the blue-leafed wattle (*Acacia saligna*), and invertebrates
547 like the zebra mussel (*Dreissena polymorpha*) (see Buchadas et al. 2017 for a review). Dynamic
548 models are especially useful to support local management of invasions and yet they are not
549 routinely implemented, probably because of the high data demand, complex model procedures
550 and detailed parameterization needed to understand, analyze and forecast biological invasions
551 (Gallien et al. 2010). Hybrid models that combine the low data requirements of statistical models
552 (such as SDMs) with the ability of dynamic models to describe underlying processes are promising
553 to improve the reliability of forecasts and facilitate the optimization of management and
554 governance (Gallien et al. 2010). In the fundamental susceptible infected-recovered (SIR) model,
555 groups of individuals within the host population are classified as “susceptible” to infection,
556 “infectious” and able to transmit the pathogen, or “recovered” and immune to reinfection (Lloyd-
557 Smith et al. 2009). Recently, the Epidemiological Framework for Biological Invasions (EFBI) has

558 adapted SIR compartment models to characterize biological invasions by treating ecosystems as
559 hosts and has allowed generalizations from epidemiology, such as the force of infection, the basic
560 reproductive ratio R_0 , super-spreaders, herd immunity, cordon sanitaire and ring vaccination, to
561 be discussed in the novel context of non-native species (Hulme et al. 2020).

562 **Factors.** Environmental conditions, including climate, set the minimum requirements necessary for
563 survival but rarely prevent the distribution of either invasive species or human pathogens (Ibáñez
564 et al. 2006). Beyond climate, invasive species modelling has demonstrated that accounting for
565 human related factors associated with the pathways of introduction and propagule pressure, such
566 as human population density, transportation networks and anthropogenic degradation, is critical
567 to increase the reliability of predictions (Gallardo et al. 2015). The same can be expected for the
568 modelling of infectious pathogens that use information on human population density and
569 movement to improve forecasts (e.g. Colizza et al. 2006, Tatem et al. 2006). Incorporating human
570 behavior, education and culture into models remains challenging for both disciplines, but could be
571 facilitated by non-traditional sources of information, such as mobile apps, news media, citizen
572 science, social media or syndromic surveillance.

573 **Approaches.** Studies of biological invasions are often used to anticipate the number and spatial
574 coverage of invasions under current and future scenarios. In contrast, epidemiologic models are
575 frequently used to estimate the relative effect of medical (e.g. vaccination) and non-medical (e.g.
576 social distancing, use of masks) interventions in reducing risk. For instance, the University of
577 Oxford and Imperial College both provided intervention scenarios for Sars-CoV-2 pandemic that
578 allowed the calculation of the estimated effect of various combinations of COVID19
579 countermeasures on R_0 (<https://bit.ly/3ezKciZ>) (Ferguson et al. 2020). Intervention scenarios on
580 the impact of biological invasions are less developed (but see Lenzner et al. 2019, Roura-Pascual et
581 al. 2021) and could greatly benefit from this approach.

582 **Biosecurity**

583 Although based on quite different disciplines, the fields of public health and invasion biology share
584 similar goals in terms of having to deliver procedures and policies that lead to the exclusion,
585 eradication or effective management of biological risks. Biosecurity policies should, by definition,
586 encompass both the risk to human health and to the environment arising from the emergence of
587 pathogens and invasive species. However, in practice nation states and multilateral conventions
588 address these risk through quite different mechanisms (Hulme 2011). Nevertheless, many
589 biosecurity risks transcend the traditional boundaries of human health and the environment and
590 call for a unified framework to reduce these risks (Hulme 2020). For example, the two most
591 common invasive non-native rats worldwide are the black rat *Rattus rattus* and the brown rat *R.*
592 *norvegicus*. Rat-borne pathogens have claimed more human lives than all the wars in history
593 combined (Hulme 2014b). The omnivorous feeding habits of rats are also implicated in crop losses
594 as well as causing the decline of many small mammals, birds, reptiles and invertebrates. Their
595 effect has been particularly severe on islands where rats have had more impact on endemic
596 biodiversity than any other factor (Townes et al. 2006). Furthermore, the global drivers of future
597 risks to public health and the environment from emerging human pathogens and invasive species
598 share many parallels. For example, climate change is likely to facilitate the poleward expansion of
599 human pathogens and non-native species; greater urbanization will lead to new hotspots for novel
600 human pathogens and invasive species; the growth in international travel has been a major
601 pathway for infectious diseases and non-native species; and increased intensification of

602 agriculture has facilitated the emergence of zoonotic agents and the spread of non-native pests
603 (Hulme 2020).

604 Unfortunately, whereas some aspects of public health ensuing from the introduction of human
605 pathogens and vector mosquitoes are managed, others, including potential vertebrate hosts and
606 ectoparasites, are less effectively addressed. Thus, an integrated approach to biosecurity that
607 addresses both species invasions and emerging infectious pathogens appears necessary. The
608 research, stakeholder and policymaker communities are rapidly beginning to understand the need
609 for better integration between disciplines. This includes initiatives such as One Health, which has a
610 goal to achieve optimal public health outcomes by monitoring and managing the interactions
611 between humans, animals, and their environment. Likewise, the Planetary Health Alliance seeks to
612 determine the human health consequences of human-caused disruptions of Earth's natural
613 systems (Myers 2017). Nevertheless, neither One Health nor Planetary Health adequately captures
614 the underlying nature of invasions by human pathogens and their relationship with invasive non-
615 native species. A more robust framework can be provided by the concept of One Biosecurity that,
616 in addition to increasing the synergies between human health and invasion science, aims to
617 refocus discussions towards practical tools and policies for preventing, eradicating and containing
618 biosecurity risks (Hulme 2020). The possibility of implementing the One Biosecurity concept has
619 been further elaborated to highlight how international public health policy can be adapted to
620 address much wider biosecurity risks stemming from invasive non-native pathogens, plants and
621 animals through developing new risk assessment tools that look beyond national borders towards
622 biosecurity risks of international concern; a stronger regulatory instrument to address biosecurity
623 threats at a worldwide scale; and the establishment of an international biosecurity convention
624 responsible for biosecurity governance (Hulme 2021).

625 *Management actions.* Management actions against epidemics follow the same steps as in
626 invasions: prevention, early detection, containment, control and eradication, and long-term
627 management (Dunn and Hatcher 2015, Robertson et al. 2020). Many countries have in place early
628 detection and rapid response systems, but the administrations in charge are usually not the same,
629 with public health institutions to prevent epidemics, separated from environmental bodies to
630 avert invasions. Successful management prospects decrease with time elapsed since the onset of
631 the invasion or pathogen emergence (Fig. 4). Due to the rapid range expansion of many invasive
632 species and pathogens, the window of opportunity for early detection and response is often very
633 short. Control is usually the action that takes most of the time and effort. Eradication is difficult to
634 achieve except in small areas or remote areas and if actions start at early stages of invasion (Pluess
635 et al. 2012). Prompt detection and control of emerging pathogens requires proper tracing of
636 infected hosts independently of whether they are symptomatic or not. Eradication is very difficult
637 when infected hosts are widespread, and often requires vaccination of 50-90 % of the population
638 depending on how contagious the pathogen might be to achieve herd immunity. A major
639 difference between an epidemic and an invasion is that when an epidemic takes place at a given
640 locality, all of these management strategies might need to be set up simultaneously. That is, within
641 a human population, different groups of people need to take different precautions or treatment
642 measures, depending on their exposure to the pathogen. In a pandemic, all management practices
643 need to be scaled up at once, both within and among populations of different regions. Conversely,
644 since the rate of expansion of an invader follows a slower pace than that of a pathogen (Fig. 5), its
645 management is more aligned with the stage of invasion than in epidemics.

646
647 *Risk assessments.* To inform managers and policy makers, research on biological invasions

648 provides semi-quantitative risk assessment tools to identify and prioritize species likely to become
649 invasive and cause damage. Risk assessments also seek to identify the most susceptible habitats to
650 invasion by a particular, or several, invasive species, through consideration of both species traits
651 and recipient ecosystem characteristics. In human epidemics, the focus of the risk analysis is
652 primarily on a particular pathogen, albeit multiple hosts, and the risk of contagion and spread, is
653 based on the traits of the pathogen and the demographic characteristics (e.g. gender, age, activity)
654 of the receptive human host population. Spatially explicit risk assessments of invasion are very
655 common and mainly rely on land-use and climate correlates between the native and the
656 introduced area. These risk analyses have been implemented in vector-borne pathogens but could
657 also be conducted for emerging pathogens albeit human population density and movement
658 patterns seem to be better predictors of disease vulnerability than environmental characteristics
659 (Jones et al. 2008). Models such as EFBI, that view ecosystems as hosts that differ in exposure,
660 susceptibility, infectivity and rates of recovery could potentially be a basis for parallel risks
661 assessments for invasive species and human pathogens since they explicitly link the transmission
662 of invasive species between ecosystems and rather than derive an arbitrary score or probability on
663 invasion likelihood, risk assessment tools could be designed to estimate R_0 (Hulme et al. 2020).

664 The evaluation of the impacts caused by epidemics focuses on the rates of infected people and
665 fatalities, which are used to compare pathogens, regions, and management responses. However,
666 as in invasions, which consequences extend beyond environmental impacts, the consequences of
667 epidemics extend beyond health, both having socioeconomic impacts (Dobson et al. 2020).
668 Attempts to quantify socioeconomic impacts in monetary terms are unlikely to provide a useful
669 basis for evaluating and comparing impacts of invasive species and pathogens, because they are
670 extremely difficult to estimate and may neglect important aspects of human well-being. In
671 invasions, there are many standardized impact assessment protocols that allow objective and
672 transparent ways to rank and identify the worst invasive species. Notably, the Socio-Economic
673 Impact Classification of Alien Taxa (SEICAT, Bacher et al. 2017) classifies invasive species based on
674 the magnitude of their impacts on human well-being, based on the capability approach from
675 welfare economics (Robeyns 2011). In SEICAT, impacts are assigned to one of five levels – from
676 minimal concern to massive – according to semi-quantitative scenarios that describe the severity
677 of the impacts on security, material and non-material assets, health, freedom of choice and action,
678 and social, spiritual and cultural relations. All these impacts apply to any epidemic and thus SEICAT
679 could be used to summarize and compare their impacts at national, regional or global scales.

680 **Conclusions**

681 In recent decades, we have witnessed how human activities that are poorly regulated can drive
682 harmful invasive species and pathogen outbreaks (Perrings et al. 2002, Stein 2020). The
683 epidemiology of human pathogens and invasion biology share many of the same mechanisms,
684 phenomena and challenges, but also potential solutions (Table 3). Global trade and travel are
685 prime causes for the introduction of invasive species and pathogens, for invasive vertebrate
686 reservoirs and for invasive insect vectors. Even the patterns and dynamics of spread of re-
687 emerging “native” diseases, such as Ebola in West Africa and dengue in Southeast Asia, share
688 similarities to those of invasive species. Many of the pathogens that cause these diseases can
689 quickly become pandemics and then go through the same stages as invasive species. Much theory
690 and empirical insights gained in invasion biology can be extended to the study of emerging
691 pathogens; similarly, invasion biology can immensely benefit from insights gained on the study of
692 emerging human infectious pathogens. The amount and quality of the data collected on human

693 infectious pathogens is undoubtedly much more refined than that available for other invasive
694 species, as has been shown for SARS-CoV-2 (Bertelsmeier and Ollier 2020).

695 A cross-disciplinary perspective on infectious diseases and invasion biology could advance both
696 fields. We advocate for an One Biosecurity (sensu Hulme 2020, 2021) approach to: (1) develop a
697 unified frameworks for studying the pathways of introduction and the consequences of eco-
698 evolutionary novelty; (2) compile and harmonize databases and information systems on major
699 invasions and epidemics; (3) share predictive modelling skills of the spread and impacts of invasive
700 species based not only on species traits but also on environmental characteristics; and (4) discuss
701 institutional approaches and protocols in horizon scanning, risk assessments, systematic
702 surveillance and monitoring of invasions and epidemics.

703 Undoubtedly, globalization and the movement of organisms across biogeographic barriers is not
704 only threatening biodiversity but also directly affecting human well-being through an array of new
705 emerging infectious threats. Invasion biology has accumulated over recent decades many insights
706 that could help improve the way we deal with these pathogens and the diseases they cause, but
707 crossing this disciplinary bridge requires more tangible collaborations and concrete policy
708 initiatives. Scientists, governments and institutions should promote the cross-disciplinary
709 approach to further advance in understanding the increasing threats of these novel entities and
710 improve prevention and response measurements.

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1192 **Figure legends**

1193 *Figure 1. Human emerging diseases can be caused directly by invasive pathogens, by pathogens*
1194 *transported by invasive vectors or reservoirs, or facilitated by invasive species not directly involved*
1195 *in the life cycle or transportation of the pathogen, but rather promoting the presence and*
1196 *abundance of its vectors and reservoirs. See examples in Table 1.*

1197 *Figure 2. Interplay between biological invasions and human emerging infectious diseases.*
1198 *Pathogen transmission can be within invasive species (left), within native or livestock species (right)*
1199 *and across invasive and native species. Dashed arrows indicate pathogen transmission to humans*
1200 *within a population (small circle) or globally (large circle).*

1201 *Figure 3. Cumulative number of publications on biological invasions, human epidemics and the*
1202 *combination of both topics according to the Web of Science from 1800 until 2020. Notice that the*
1203 *y-axis is in log scale. The search term for human epidemics was “human epidemics” whereas for*
1204 *biological invasions, the search term was “ecological invasions”. This term was more specific to*
1205 *retrieve all studies on that topic, while excluding non-topic studies (e.g. cancer research,*
1206 *pharmacology and biomaterial science).*

1207 *Figure 4. Comparing the stages of biological invasions and human epidemic (adapted from*
1208 *Woolhouse and Gaunt 2007, Blackburn et al. 2011, Hatcher et al. 2012, Jeschke et al. 2013), and*
1209 *possible management actions at these stages (adapted from Dunn and Hatcher 2015, Robertson et*
1210 *al. 2020). Pathogens that emerge and cause an epidemic anywhere on the globe can be*
1211 *transported and spread globally leading to a pandemic in the worst case (dotted arrow). Bent*
1212 *arrows indicate potential positions of zoonotic pathogen interspecific spillover.*

1213 *Figure 5. Density plot showing the frequency of observed radial spread rates (log scale) for*
1214 *different pathogens and invasive taxonomic groups. The height of each density curve indicates the*
1215 *relative number of data points, normalized to 1. Numbers at the right indicate the median rate of*
1216 *spread for the group. Figure created with packages ggplot2 and ggridges in R v. 4.0.0. Raw data*
1217 *extracted from: (Smal and Fairley 1984, van den Bosch et al. 1992, Holmes 1993, Teangana et al.*
1218 *2000, McCallum et al. 2003, Phillips et al. 2007, Pioz et al. 2011, Fraser et al. 2015, Zinszer et al.*
1219 *2015, 2017, Evans 2016, Roques et al. 2016, Horvitz et al. 2017, Hadfield et al. 2019).*

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1221 *Box 1. Definitions of terms and concepts as used in this paper*

1222 **Emerging infectious disease:** an infectious disease that appears in a human population for the first
1223 time or has existed previously but is rapidly increasing in incidence, impact or geographic range
1224 (<http://www.emro.who.int/health-topics/emerging-diseases/index.html>).

1225 **Epidemic:** a disease event affecting many persons at the same time, and spreading from person to
1226 person in a locality or region during a specific period of time
1227 (https://www.who.int/csr/disease/swineflu/frequently_asked_questions/pandemic/en/).

1228 **Invasive species:** a non-native introduced species that form self-sustaining populations and spread
1229 rapidly from the sites of introduction (Blackburn et al. 2011).

1230 **Invasiveness:** intrinsic characteristics of a non-native species to invade outside its region of origin
1231 (Lonsdale 1999).

1232 **Invasibility:** susceptibility of an ecosystem to be invaded. It depends on the biotic and abiotic
1233 characteristics of the recipient ecosystem (Lonsdale 1999).

1234 **Non-native species:** an introduced species transported intentionally or unintentionally to a new
1235 region by humans (Blackburn et al. 2011).

1236 **One Biosecurity:** an interdisciplinary approach to biosecurity policy and research that builds on
1237 the interconnections between human, animal, plant, and environmental health to effectively
1238 prevent and mitigate the impacts of invasive alien species (Hulme 2021).

1239 **One Health:** cross-sectoral approach to achieve optimal public health outcomes by monitoring,
1240 managing and investigating the interactions between humans, animals, and their environments
1241 (Ogden et al. 2019).

1242 **Outbreak:** the occurrence of more infection cases than expected in a particular population, in a
1243 specific geographical area and in a specified period ([http://www.emro.who.int/health-](http://www.emro.who.int/health-topics/disease-outbreaks/index.html)
1244 [topics/disease-outbreaks/index.html](http://www.emro.who.int/health-topics/disease-outbreaks/index.html)).

1245 **Pandemic:** an epidemic occurring worldwide, or over a very wide area, crossing international
1246 boundaries and usually affecting a large number of people
1247 (https://www.who.int/csr/disease/swineflu/frequently_asked_questions/pandemic/en/)

1248 **Pathogen pressure:** amount of pathogen available to the human host at a given point in space and
1249 time (Plowright et al. 2017).

1250 **Reservoir:** an animal species that hosts a pathogen, typically without being harmed, and is the
1251 source of infection to other host species (Rabitsch et al. 2017).

1252 **Spillover:** transmission of a pathogen from a reservoir to a novel susceptible host (Rabitsch et al.
1253 2017).

1254 **Time lag:** period between the introduction of a non-native species and its establishment in the
1255 new range. In the broad sense, it can be applied to the time required to overcome any phase of
1256 the invasion process (Crooks 2005).

1257 **Vector:** a species, typically but not always an arthropod, that carries and transmits a pathogen to
1258 another species (Rabitsch et al. 2017).

1259 **Virulence:** ability of a microorganism to cause disease. It depends on characteristics of the
1260 pathogen and the host (Horrocks et al. 2011).

1261 **Zoonosis:** a disease causing pathogen that is transmitted between vertebrate animals (wildlife,
1262 livestock or domestic animals) and humans (Rabitsch et al. 2017).

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1265 *Table 1. Species from the IUCN list “100 of the World’s Worst Invasive Alien Species” (Lowe et al.*
 1266 *2000) that can transmit pathogens to humans or are themselves pathogens. The introduction*
 1267 *pathways (according to the Convention of Biological Diversity) and impact types (A: damage*
 1268 *human activities such as to agriculture, forestry, livestock or infrastructures; B: biodiversity; H:*
 1269 *human health) are indicated.*

Invasive species	Pathogens (diseases)	Transmission	Pathways	Impacts
<i>Acridotheres tristis</i> , common myna	<i>Ornithonyssus bursa</i> and <i>Dermanyssus gallinae</i> (dermatitis, skin inflammation, severe irritation and rashes, asthma) Their droppings can spread psittacosis, ornithosis, salmonellosis and arboviruses.	Reservoir	Intentional/ Escape from confinement: Zoo, Pet trade Intentional/ Release in nature: Fauna “improvement”	A, B, H
<i>Aedes albopictus</i> , tiger mosquito	<i>Flavivirus</i> spp. (e.g. West Nile, dengue fever), <i>Diofilaria immitis</i> (filariasis)	Vector	Unintentional/ Transport-stowaway: Vehicles	H
<i>Achatina fulica</i> , Giant African land snail	<i>Metastrongylus</i> spp., <i>Angiostrongylus cantonensis</i> and <i>A. costaricensis</i> (pulmonary metastrongylosis and eosinophilic meningoencephalitis)	Reservoir	Intentional/Escape from confinement: Pet, Aquarium and terrarium species, Research, Horticulture, Live food	H, A
<i>Anopheles quadrimaculatu</i> , mosquito	<i>Plasmodium</i> spp. (malaria), West Nile virus (meningoencephalitis)	Vector	Unintentional/ Transport-stowaway: Vehicles	H
<i>Eichhornia crassipes</i> , water hyacinth	<i>Plasmodium</i> spp. (malaria) transmitted by Anopheline mosquitoes	Invasive facilitator (habitat for vector)	Intentional/ Escape from confinement: Aquarium species	A, B, H
<i>Eriocheir sinensis</i> , Chinese mitten crab	<i>Paragonimus westermanii</i> (human lung fluke parasite),	Reservoir	Intentional/ Escape from confinement: Aquaculture, Aquarium species.	A, B, H

			Unintentional/ Transport- stowaway: Ship- boat ballast water, Ship-boat hull fouling	
<i>Euglandina rosea</i> , rosy wolf snail	<i>Angiostrongylus cantonensis</i> (pulmonary metastrongylosis and eosinophilic meningoencephalitis)	Reservoir	Intentional/ Release in nature: Biological control	B, H
<i>Herpestes javanicus</i> , small Indian mongoose	<i>Leptospira interrogans</i> (Weil's disease), Lyssavirus (rabies)	Reservoir	Intentional/ Release in nature: Biological control	B, H
<i>Lantana camara</i> , lantana shrub	<i>Tripanosoma</i> spp. (sleeping sickness) transmitted by <i>Glossina</i> spp., tse tse fly	Invasive facilitator (habitat for vector)	Intentional/ Escape from confinement: Horticulture	A, B, H
<i>Macaca fascicularis</i> , crab-eating macaca	Macacine herpesvirus 1 (herpes B), Lyssavirus (rabies)	Reservoir	Intentional/ Escape from confinement: Live food, Research	A, B, H
<i>Mus musculus</i> , house mouse	<i>Yersinia pestis</i> (bubonic plague), <i>Salmonella</i> spp. (salmonellosis)	Reservoir	Unintentional/ Transport- stowaway: Container, bulk	A, B, H
<i>Rattus rattus</i> , black rat	<i>Leptospira interrogans</i> (Weil's disease), <i>Yersinia pestis</i> (bubonic plague)	Reservoir	Unintentional/ Transport- stowaway: Container, bulk	A, B, H
<i>Sturnus vulgaris</i> , starling	<i>Chlamydophila psittaci</i> (psittacosis)	Reservoir	Intentional/ Release in nature: Biological control, Hunting, Fauna "improvement"	A, H

<i>Sus scrofa</i> , feral pig	<i>Leptospira interrogans</i> (Weil's disease)	Reservoir	Intentional/ Release in nature: Hunting	A, B, H
<i>Trachemys scripta elegans</i> , red eared slider turtle	<i>Salmonella</i> spp. (salmonellosis)	Reservoir	Intentional/ Escape from confinement: Aquarium and terrarium species	A, B, H
<i>Vulpes vulpes</i> , red fox	Possible role in Lyssavirus (rabies) transmission	Reservoir	Intentional/ Release in nature: Hunting	A, B, H

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Table 2. Differences and common challenges associated with the forecasting of biological invasions and human epidemics with indications of the potential for collaboration and cross-fertilization across disciplines.

	Biological invasions	Human epidemics	Potential cross-fertilization across disciplines
Data used	Geo-referenced species occurrence Rarely, abundance data	Number of infected individuals Information rarely geo-referenced	Common monitoring systems and data platforms
Indicators (developed to follow an outbreak)	Likelihood of species presence (suitability) Number of non-native species	R_0 , likelihood of exponential spread	Correlation between disease and invasion indicators
Models	Mostly spatially, niche-based, e.g. Species Distribution Models (SDMs)	Dynamic, biology-based e.g. Susceptible Immune Recovered (SIR)	Sharing modelling tools and advances to reduce uncertainty
Scales	Regional to global Years/decades	Local to regional Rarely global Weeks/months	Automatically updated platforms to follow an outbreak
Critical factors (ordered)	Climate Environmental conditions Human activities (e.g. transport, land-use) Biological (e.g. dispersal)	Biological (e.g. transmissibility) Human activities (e.g. transport) Human behavior (e.g. sociability) Management (e.g. medical and non-medical actions)	Share environmental and human data for modelling New sources of human-related data (e.g. mobile phones, trade flows)
Approaches	Exploratory Climate change scenarios Management scenarios	Intervention scenarios	Common scenario frameworks and workflows
Common challenges	Data quality and quantity Modelling of complex systems under imperfect detection Incorporating human activities and behaviors Anticipating alternative policy and management scenarios High intrinsic uncertainty associated to exponential processes Traceability of origin and expansion of pathogen/invader Lag phases (e.g. between introduction and impact, between management and effective mitigation)		

	Anticipating the next biological threat based on transmissibility/spread and potential impacts
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Table 3. Comparison of main features and established concepts of biological invasions with human epidemics.

Feature	Biological invasions	Human epidemics	References
Biogeographic and evolutionary origin	Non-native species from a region where they could not be dispersed without human agency	Non-native pathogens dispersed directly or indirectly by humans or emerging native pathogens. Crossing a species barrier rather than a biogeographic barrier	(Jones et al. 2008, Pyšek et al. 2017)
Routes of dispersal	<p>Pathways</p> <p>Intentional: release and escape</p> <p>Unintentional: contaminant, stowaway, corridor and unaided</p>	<p>Routes of infection</p> <p>Unintentional: vector borne, zoonotic, human contact, indirect contact by ingestion or the environment</p> <p>Also intentional: historical cases during colonization of new territories, bioterrorism and anthrax mailing</p>	(Wolfe et al. 2007, Hulme et al. 2008, Saul et al. 2017)
Founder populations	Repeated introductions from several populations, genetically diverse (admixtures)	Few introductions from a single or few populations	
Stages	Transport, introduction, establishment, spread	Exposure, infection, transmission, epidemic spread; zoonotic spillover	(Woolhouse and Gaunt 2007, Blackburn et al. 2011, Jeschke et al. 2013)
Spread rates and time lags	<p>0.1-10² km/yr</p> <p>Years-decades</p>	<p>10³-10⁴ km/yr</p> <p>Days-decades</p>	<p>(Kowarik 1995, McCallum et al. 2003)</p> <p>See Figure 4</p>

Main studied causes of non-native species performance and impact	Traits of the organism (invasiveness), biotic and abiotic characteristics of the recipient ecosystem (invasibility) and the intensity and frequency of introduced individuals (propagule pressure)	Traits of the organism (pathogenicity), host age, genetics, physiology, immunity and people behavior	(Lonsdale 1999, Mack et al. 2000, Enders et al. 2020)
Forecasting models' focus and explanatory variables	On the invasive species. Environmental and proxies for propagule pressure as explanatory variables	On infected people (not the pathogen). Human demographics including movement and pathogen transmission as explanatory variables	See Table 2
Traditional impact focus	Biodiversity, environment, agriculture and farming	Medical, public health	(Jeschke 2014, Vilà and Hulme 2016)
Traditionally involved management sectors	Environment, agriculture and farming, veterinary, water resources, trading	Public health, food, foreign affairs, traveling, veterinary, water resources	(Ogden et al. 2019)

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Figure 1

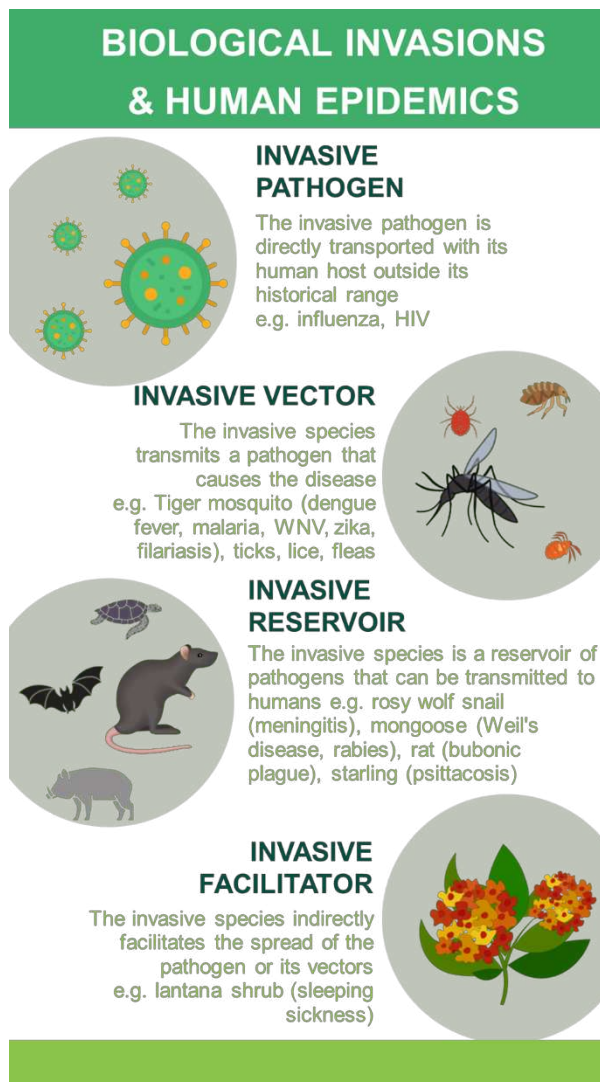


Figure 2

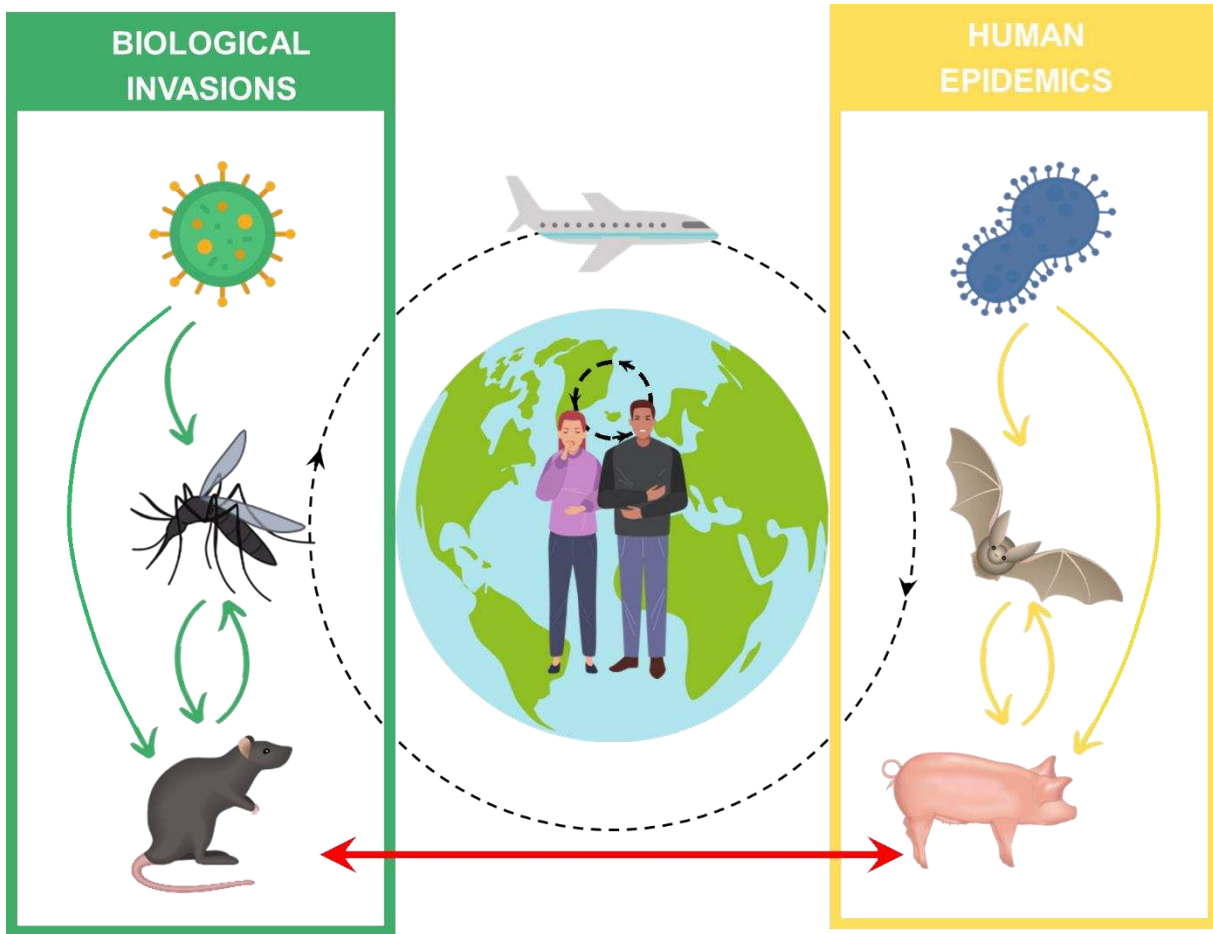


Figure 3

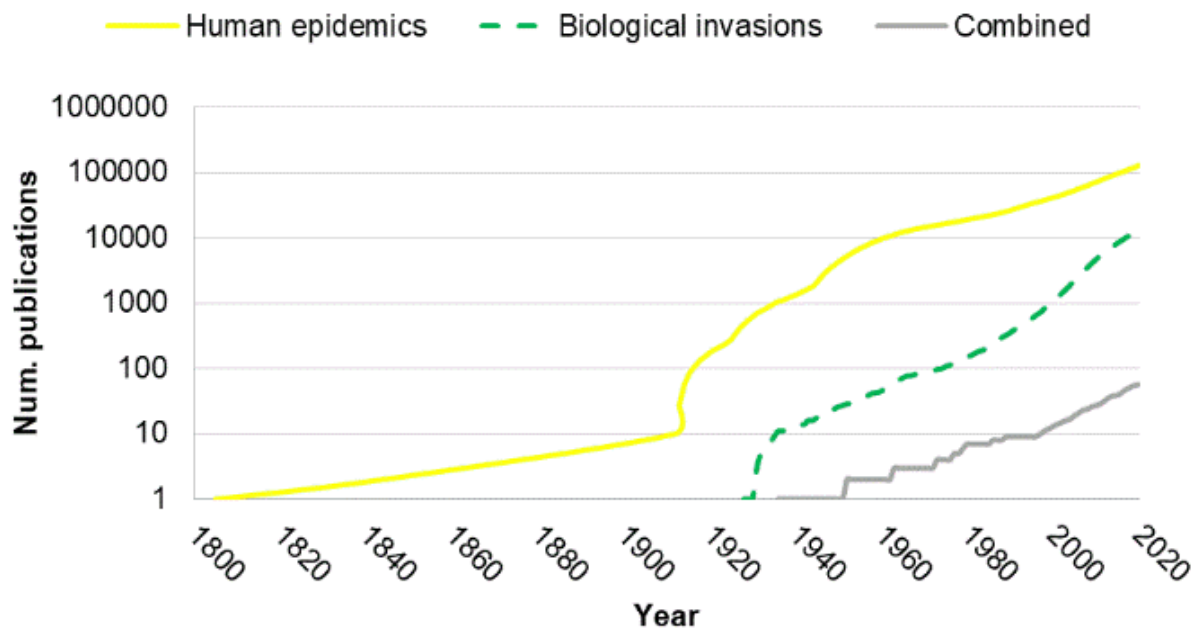


Figure 4

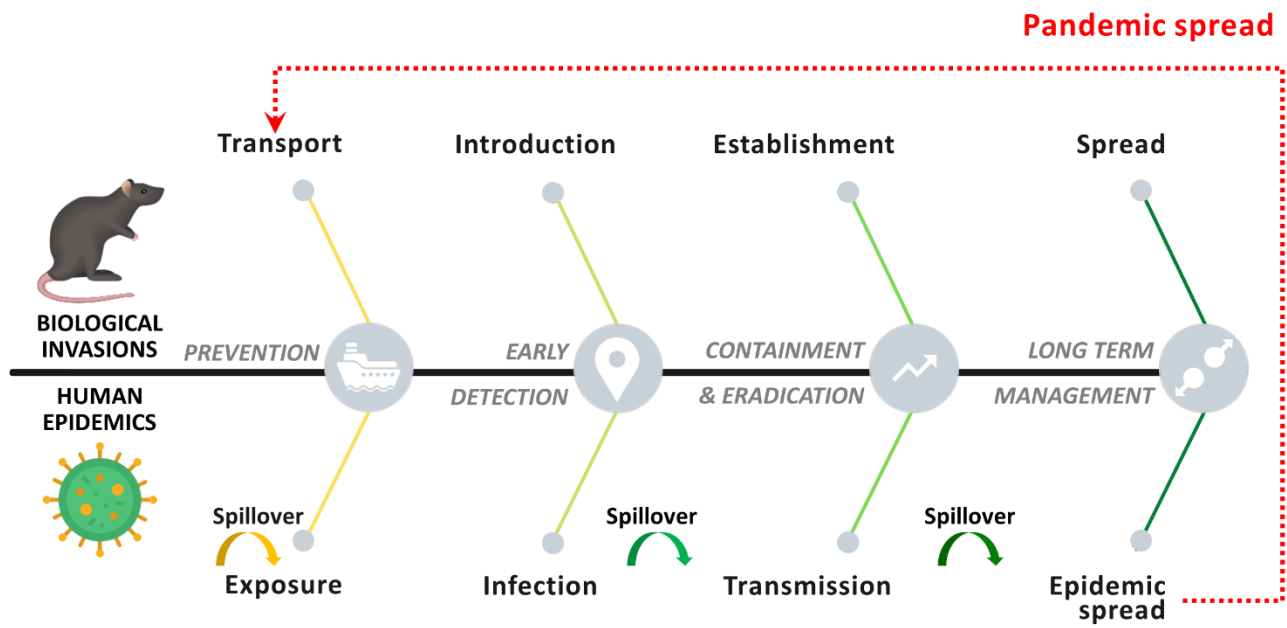


Figure 5

