

1 This is the accepted version of an article published in Biological Invasions. The final version is  
2 available online at: <https://doi.org/10.1007/s10530-020-02368-5>

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4 **Functional segregation of resource-use strategies of native and invasive plants across Mediterranean**  
5 **Biome communities**

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13 **Abstract**

14 Functional segregation among species in a community depends on their mean trait values (i.e. functional  
15 distinctiveness), and the range of trait attributes exhibited by each species (i.e. functional diversity). Previous  
16 evidence suggests that invasive plants tend to display traits related to a more acquisitive resource-use strategy  
17 than natives. However, the contribution of intraspecific trait variation to functional diversity has received little  
18 attention in community ecology, and might provide interesting information about community processes. In this  
19 study, we used eight plant traits related to carbon and nutrient acquisition of coexisting dominant native and  
20 invasive plants in eight communities across the Mediterranean-climate biome to determine sources of  
21 functional segregation between native and invasive species. We found three major axes of functional variation,  
22 related to leaf economics, resource-use efficiency, and plant height. Invasive species across communities had  
23 leaf traits related to an acquisitive resource-use strategy in contrast to native species, whereas differences in the  
24 second and third principal components were community dependent. Invasive species were more functionally

25 diverse than native species across the dataset and in four out of the eight communities. Intraspecific variance  
26 accounted for 11-27% of total trait variation and was on average greater in invasive species, and especially  
27 important in the axis related to resource use efficiency. These results, although dependent on the trait and  
28 community considered, offer interesting insights to the sources of functional trait diversity of native and  
29 invasive species within communities, indicating that intraspecific variation might not be equally distributed  
30 between native and invasive species.

31 **Key words:** functional diversity; functional strategies; intraspecific variance; invasive plants; leaf economics  
32 spectrum; plant life form;

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35 **Conflict of interest.** The authors declare that they have no conflict of interest.

36 **Credit authorship contribution statement.** **J.G.D.** Conceptualization, Methodology, Formal analysis,  
37 Investigation, Data curation, Writing - original draft, Writing - review & editing. **E.G.** Conceptualization,  
38 Methodology, Investigation, Writing - review & editing. **J.L.F.** Investigation, Writing - review & editing.  
39 **M.V.** Conceptualization, Investigation, Writing - review & editing.

40 **Funding.** J.G.D. has received financial support through the “la Caixa” INPhINIT Fellowship Grant for  
41 Doctoral studies at Spanish Research Centres of Excellence (LCF/BQ/IN17/11620012). J.G.D. and M.V. from  
42 EXARBIN (RTI2018-093504-B-I00) project of the Ministerio de Ciencia e Innovación of Spain.

43 **Introduction**

44 Functional segregation among species in a community depends on the mean difference between their trait  
45 values, which represents their functional distinctiveness; and the range of trait values exhibited by each  
46 species, which contributes to functional diversity (Violle and Jiang 2009; Hulme and Bernard-Verdier 2018).  
47 High functional distinctiveness allows invasive species to establish in the recipient community by minimizing  
48 interspecific competition (MacDougall et al. 2009; Cadotte et al. 2018; Divišek et al. 2018; de la Riva et al.  
49 2019). Previous evidence suggests that invasive species tend to display traits related to fast return on  
50 investments of nutrients, such as higher specific leaf area or lower tissue construction costs, than native  
51 congeners (Pyšek and Richardson 2007), or coexisting natives (Daehler 2003; Ordonez et al. 2010; Funk et al.  
52 2016). This is often the case in Mediterranean habitats, where invasive plants frequently display traits of rapid  
53 resource acquisition in comparison with coexisting native species (D'Antonio and Vitousek 1992; Arianoutsou  
54 et al. 2013).

55 Mediterranean regions are frequently dry and low-resource environments where strong abiotic constraints  
56 structure trait composition (de la Riva et al. 2018; Michelaki et al. 2019). Thus, functional distinctiveness  
57 might be limited within the environmental constraints of recipient Mediterranean communities (environmental  
58 filtering; Cadotte et al. 2018). It has been argued that environmental filtering is important at higher scales, with  
59 species from the same biome being functionally similar (Echeverría-Londoño et al. 2018), whereas functional  
60 distinctiveness operates at a local scale (Loiola et al. 2018). For instance, plant species adapted to dry habitats,  
61 such as the Mediterranean, often show contrasting resource uptake strategies due to opportunistic behaviour  
62 with respect to water and nutrient use efficiency (Querejeta et al. 2018; Carvajal et al. 2019). Furthermore,  
63 studies have observed invaders with very different functional profiles to establish in Mediterranean habitats  
64 (Tecco et al. 2010; de la Riva et al. 2019), suggesting that different mechanisms of invasion might operate in  
65 Mediterranean ecosystems.

66 Functional diversity of a community depends on interspecific and intraspecific trait variance (Violle and Jiang  
67 2009). Intraspecific variance depends on the species and traits under consideration (Albert et al. 2011; Siefert  
68 et al. 2015), environmental constraints (i.e. climate and resource availability), and ecological processes  
69 operating in each community (Grime and Mackey 2002; Messier et al. 2010; Walters and Gerlach 2013).  
70 Because intraspecific trait variance often accounts for a smaller proportion (~25%) of total trait diversity than

71 interspecific variance (Albert et al. 2011; Siefert et al. 2015), it is frequently overlooked in the context of  
72 biological invasions (Hulme and Bernard-Verdier 2018; but see Helsen et al. 2020). However, the sources and  
73 extent of intraspecific trait variance might differ between native and invasive species within the same  
74 community. It has been argued that, whereas native species occupy specific suitable patches within their range  
75 (Gallien et al. 2010), invasive plants are often generalist species (Okimura and Mori 2018), able to thrive under  
76 diverse ecological constraints (Clavel et al. 2011), and may show greater plasticity than phylogenetically  
77 related non-invasive species (Sultan 2001; Funk 2008; Davidson et al. 2011; Martín-Forés et al. 2017). Thus,  
78 we might expect that, in a given community, the contribution of intraspecific variance to total trait diversity  
79 will be greater in invasive than native species.

80 To correctly assess the role of intraspecific variance in the segregation of native and invasive species, it is  
81 necessary to consider two things. First, the relative contribution of intraspecific trait variance to total diversity  
82 might largely depend on the observed interspecific variance (de Bello et al. 2011). Thus, a greater contribution  
83 of intraspecific trait variance to total diversity of invasive species might just reflect that native species are  
84 overall more diverse, i.e. are more different among themselves (have greater interspecific variance). Second,  
85 native and invasive species across Mediterranean Regions tend to show contrasting life forms, with invasive  
86 species being more frequently annual species in contrast to native perennials species (Arianoutsou et al. 2013;  
87 Funk et al. 2016). Therefore, intraspecific trait variance in Mediterranean communities could depend on  
88 species' life forms, rather than origin per se. It has been argued that long-lived plant species might show higher  
89 intraspecific variation in traits related to leaf morphology due to greater ontogenetic variation (Watson et al.  
90 1995; Sultan 2004), but be more physiologically constrained by costly leaf tissues than ruderal and fast-  
91 growing plant species (Maire et al. 2013). It is crucial to assess these considerations to correctly understand  
92 functional segregation in trait-space.

93 Plant performance is rarely determined by a single trait, thus it is necessary to move towards a whole-plant  
94 approach by exploring trait covariation along functional axes (Albert et al. 2011; Díaz et al. 2016; Hulme and  
95 Bernard-Verdier 2018). There are several axes of trait covariation, such as leaf and water economy, light  
96 competition or reproductive effort (Laughlin 2014; Díaz et al. 2016; Prieto et al. 2018). In this regard, the best  
97 known axis of plant trait covariation is the leaf economics spectrum (Wright et al. 2004), which represents a  
98 trade-off between acquisition and conservation of resources, i.e. quick or slow return of investments. Whereas

99 the acquisitive-end of this spectrum encompasses species with short-lived leaves with high maximum  
100 photosynthetic rates and leaf nutrient content, the conservative-end includes species with long-lived leaves and  
101 high construction costs. As there are several known axes of trait covariation, it is necessary to use tools that  
102 allow to estimate functional segregation considering all functional axes simultaneously such as trait  
103 hypervolumes based on kernel density estimation methods (Blonder et al. 2018). Hypervolumes define high-  
104 dimensional, non-continuous shapes and permit an accurate quantification of the amount of trait-space  
105 occupied by a species (i.e. functional diversity) (Blonder et al. 2018). Hence, hypervolumes may be a useful  
106 tool to explore functional segregation between native and invasive species (see Guerin et al. 2019; Helsen et al.  
107 2020).

108 To test this framework, we analysed native and invasive species' traits data from eight communities across the  
109 five Mediterranean-climate regions (Funk et al. 2016). Previous analyses with a focus on single traits showed  
110 that invasive species had traits related to faster resource acquisition and use than native species, and that this  
111 difference was related to life form (Funk et al. 2016). Here, we aim to understand how native and invasive  
112 species segregate in trait-space by exploring the two sources of functional segregation, i.e. functional  
113 distinctiveness and diversity. First, we identified the major axes of trait covariation, and determined whether  
114 native and invasive species occupy different positions along them (i.e. functional distinctiveness). Second, we  
115 compared the functional diversity of native and invasive species, and quantified whether the contribution of  
116 intraspecific trait variance to the total functional diversity differs between invasive and native species.

## 117 **Material and methods**

### 118 Field data collection

119 We analysed plant traits of dominant native and invasive species of eight representative communities from five  
120 Mediterranean regions collected by Funk et al. (2016, 2017) (Table 1). Particularly, we measured eight traits of  
121 leaf morphology, physiology (leaf chemical compounds) and plant size related to plant resource-use and  
122 acquisition strategies (Table 2). These traits have been widely studied in the literature because of their  
123 importance in community assembly (Tecco et al. 2010; Michelaki et al. 2019; Henn et al. 2019; Helsen et al.  
124 2020). The database included a total of 734 observations: 137 species and four to five replicate plants per  
125 species and community (Table A1). Eighteen species were present in more than one community.

126 **Table 1** Studied communities and their soil characteristics, mean annual precipitation, management regimes,  
 127 and number of species grouped by origin and life form.

Region	Community	Soil N (g N/100 g soil)	Soil P (mg P/kg soil)	Soil pH	Mean annual rainfall	Grazing	Years since last fire	Number of native species	Number of invasive species	Annual	Herbaceous perennial	Woody	Total
Australia	banksia woodland	0.05	47.7	5.38	834	Low	>20	10	8	5	9	4	18
	coastal banksia woodland	0.06	150.9	5.74	734	Low	>30	7	6	3	7	3	13
Spain	coastal grassland	0.05	117.8	7.96	550	High	>50	20	9	11	10	8	29
California	serpentine grassland	0.23	30	6.69	760	Low	>50	23	4	17	6	4	27
	coastal sage scrub	0.19	628.3	6.58	330	Low	3	14	9	10	6	7	23
South Africa	acid sands fynbos	0.03	31	4.81	522	Low	10	11	5	3	6	7	16
	renosterveld	0.29	304	5.75	515	Low	>50	11	5	5	3	8	16
Chile	sclerophyll woodland	0.09	1001	7.15	360	Low	>100	13	5	5	5	8	18

128

129 **Table 2** Traits considered in this study, abbreviation, units and functional role.

trait	abb.	units	significance
Leaf mass per area	LMA	$\text{g} \times \text{m}^{-2}$	Plant investment in structural leaf tissue. High LMA indicates a conservative resource-use strategy.
Mass-based photosynthetic rate	A <sub>mass</sub>	$\text{nmol CO}_2 \times \text{g}^{-1} \times \text{s}^{-1}$	Photosynthetic rate per leaf mass at saturating light levels.
Instantaneous water-use efficiency	WUE	$\mu\text{mol CO}_2 \times \text{mmol H}_2\text{O}^{-1}$	Ratio between CO <sub>2</sub> assimilation and transpiration.
Mass-based leaf nitrogen content	N <sub>mass</sub>	$\text{mg N} \times \text{g}^{-1}$	Amount of nitrogen per leaf mass. Related to plant growth and economics.
Photosynthetic nitrogen-use efficiency	PNUE	$\text{mmol CO}_2 \times \text{mol}^{-1} \text{N} \times \text{s}^{-1}$	Photosynthetic capacity per unit leaf nitrogen.
Mass-based leaf phosphorus content	P <sub>mass</sub>	$\text{mg P} \times \text{g}^{-1}$	Amount of phosphorus per leaf mass. Related to plant growth and economics.
Photosynthetic phosphorus-use efficiency	PPUE	$\text{mmol CO}_2 \times \text{mol}^{-1} \text{P} \times \text{s}^{-1}$	Photosynthetic capacity per unit leaf phosphorus.
Plant height	Height	cm	Indicates the position of the plant in the vertical light gradient of the community.

130 Species were classified according to their origin as native or invasive; and according to their life form as  
131 annual, herbaceous perennial, or woody. Here “invasive” follows the definition of Richardson et al. (2000), i.e.  
132 non-native species with great reproductive potential which become very abundant locally and are able to  
133 quickly spread from the area of introduction. The category annual were therophytes, i.e. species that spend the  
134 summer in the seed bank such as grasses and forbs. Herbaceous perennials were plants without lignified stems  
135 but with dormant organs below or near the ground. Woody species included shrubs and trees.

#### 136 Data analyses

137 First, we ran a principal component analysis (PCA) to identify major axes of trait covariation, and reduce the  
138 dataset into fewer functional dimensions (Laughlin, 2014). We used the first three principal components for  
139 posterior analyses, those with eigenvalues greater than one. Then, we explored the functional distinctiveness  
140 between native and invasive species, i.e. if native and invasive species differ in their mean trait values along  
141 each principal component. For each principal component, we fitted a linear mixed model for the total dataset  
142 with community and species, nested within community, as random effect, and a linear mixed model for each  
143 community with species as random effects. We used the Satterthwaite method to approximate degrees of  
144 freedom. The residual versus fitted plots revealed that errors were normally distributed and homoscedastic.

145 Second, to characterize the functional diversity of native and invasive species, we built hypervolumes using a  
146 Box kernel density estimation method and Silverman bandwidth estimator (hypervolume package, Blonder et  
147 al. 2018). For the trait-space defined by the three principal components, we built several hypervolumes: one for  
148 all native species in the dataset, one for all invasive species in the dataset, and one for each group of native and  
149 invasive species within each community. We also built hypervolumes to estimate the diversity of the total pool  
150 of native and invasive species in each principal component. The units of the hypervolumes are reported as the  
151 standard deviations of PCA scores in the first three axes, raised to the power of the number of trait dimensions  
152 ( $SD^3$ ). As hypervolumes depend on species richness, and all communities have more native than invasive  
153 species, we created 99 randomized communities where the number of native species was adjusted to the  
154 number of invasive species in the community (see invasive species column in Table 2). To compare the  
155 functional diversity of native and invasive species across randomized communities, we calculated the mean  
156 effect size (Hedges’*d*) and bias-corrected 95%-bootstrap confidence intervals (effsize package, Torchiano

157 2018). A mean effect size was considered significantly different from zero when its confidence interval did not  
158 bracket zero.

159 Third, we assessed if the relative contribution of intraspecific variance to total trait diversity (i.e. total variation  
160 of a pool of native or invasive species in a given principal component) is greater in invasive than native  
161 species. We partitioned the total community variance into interspecific and intraspecific variance as formulated  
162 in equation 1 of de Bello et al. (2011), i.e. the extent of trait variation in a community which results from  
163 variation between coexisting species and variation among individuals of a species. In this formulation, the  
164 contribution of each species to the variance partitioning is identical (de Bello et al. 2011). Then, we divided the  
165 absolute intraspecific variance by the total community variance to obtain the relative contribution of  
166 intraspecific trait variance to total trait diversity (Siefert et al. 2015). We did this for the total pool of native  
167 and invasive species, and native and invasive species within each community.

168 Finally, we explored whether the functional diversity and the contribution of intraspecific trait variance to total  
169 diversity of native and invasive species is related to their life forms. For all three dimensions, and each  
170 principal component, we estimated the functional diversity of the total pool of annual, herbaceous perennial  
171 and woody species with hypervolumes (Blonder et al. 2018), and compared them by computing Hedges' d and  
172 bias-corrected 95%-bootstrap confidence intervals between all groups (effsize package, Torchiano 2018).  
173 Then, we partitioned the total trait diversity of each group into interspecific and intraspecific variance (de Bello  
174 et al. 2011), and calculated their contributions to total diversity (Siefert et al. 2015). If the relative contribution  
175 of intraspecific trait variance of native and invasive species is related to their life forms, then we might expect  
176 to observe a similar or greater effect size between annual and perennial species compared to between native  
177 and invasive. All statistical analyses were performed, and all figures produced, in the R-3.6.1 statistical  
178 platform (R Development Core Team, 2019).

## 179 **Results**

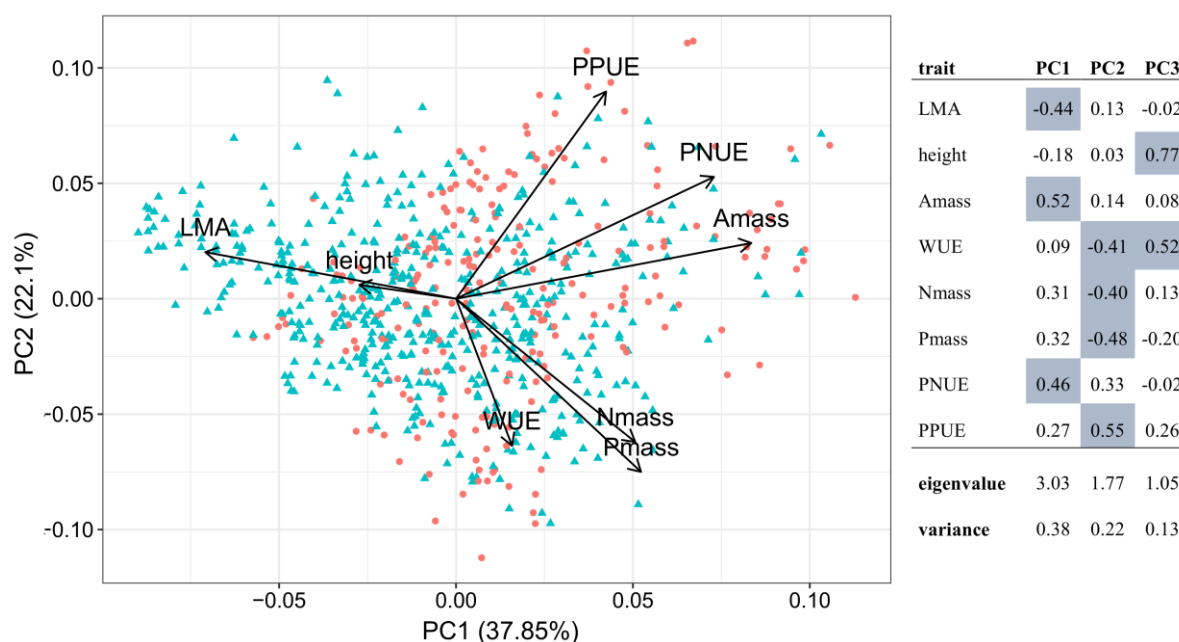
### 180 Functional strategies of native and invasive species

181 The first three components of the PCA accumulated 73.11% of the total variance (Fig. 1). The first principal  
182 component explained 37.85% of the variance and reflected a coordination between Amass (mass-based  
183 photosynthetic rate), LMA (leaf mass per area), and PNUE (photosynthetic nitrogen-use efficiency) (Fig. 1),



184 which is representative of the leaf economics spectrum (Wright et al. 2004). The second principal component  
 185 explained 22.10% of the variance and reflected a covariation between leaf nutrient concentration, WUE  
 186 (instantaneous water-use efficiency) and PPUE (photosynthetic phosphorus-use efficiency) and represents the  
 187 trade-off between water- and phosphorous-use efficiency. The third principal component explained 13.15% of  
 188 the variance and was linked to plant height and, to a lesser extent, WUE.

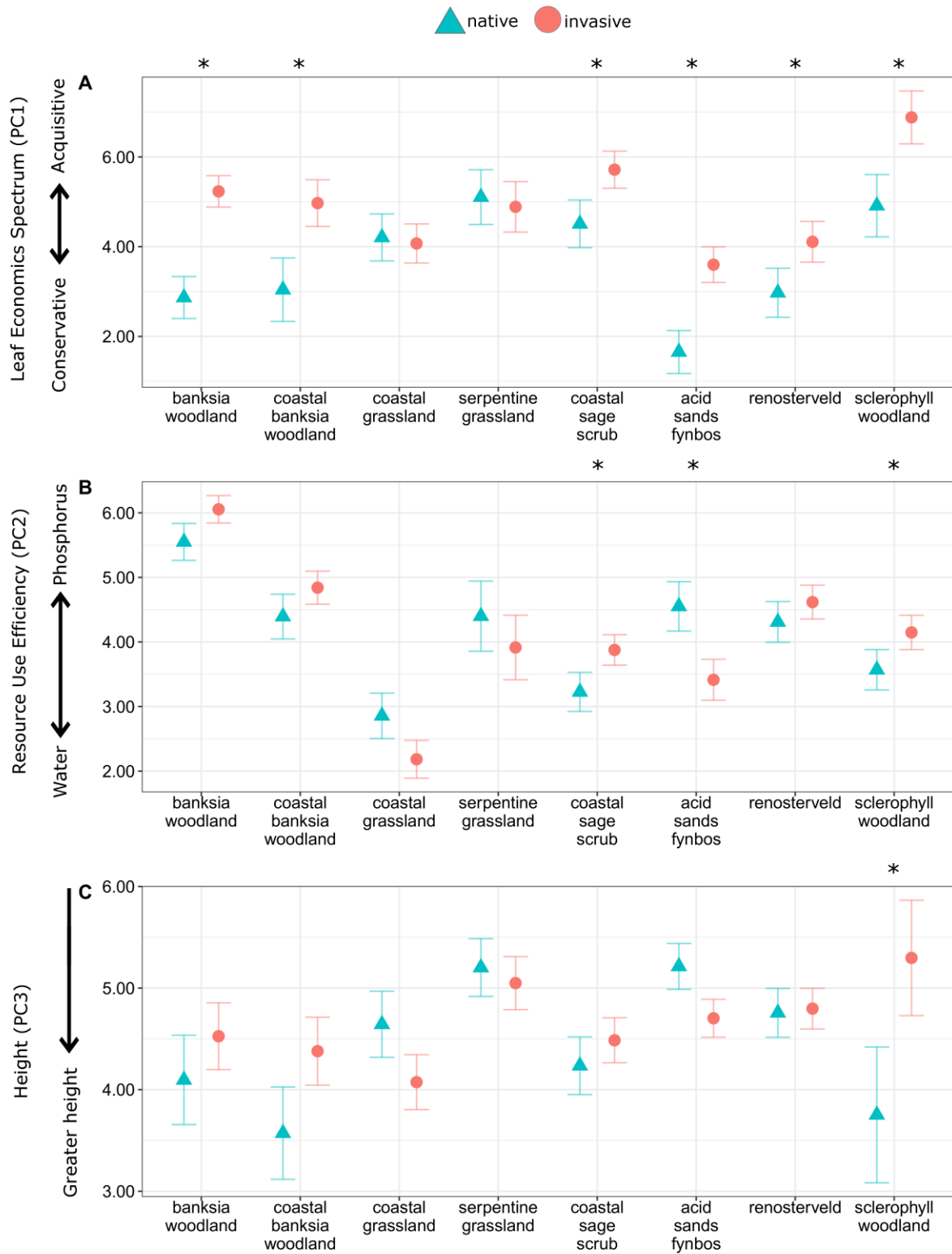
189 **Fig. 1** Principal Component Analysis (PCA) of eight plant traits from 137 natives (blue triangles) and invasive  
 190 (red dots) plant species in Mediterranean communities (4 to 5 replicates per species). The table below shows  
 191 the loadings and variance associated with each principal component with eigenvalues over 1. The most  
 192 relevant traits of each principal component have been shaded. Traits: LMA: leaf mass per area, Amass: mass-  
 193 based photosynthetic rate, WUE: instantaneous water use efficiency, Nmass: mass-based leaf nitrogen  
 194 concentration, Pmass: mass-based leaf phosphorus concentration, PNUE: photosynthetic nitrogen-use  
 195 efficiency, PPUE: photosynthetic phosphorus-use efficiency, and height: vegetative plant height.



196  
 197 Overall, native species occupied the high LMA end of PC1, whereas invasive species occupied the high Amass  
 198 and PNUE end ( $F_{1,157}=27.04$ ,  $p<0.001$ ). Native and invasive species occupied similar positions in PC2  
 199 ( $F_{1,157}=0.03$ ,  $p=0.87$ ) and PC3 ( $F_{1,158}=0.94$ ,  $p=0.33$ ). Within communities, invasive species were significantly  
 200 displaced towards the high Amass and PNUE end of PC1 in six communities (Fig. 2A). Invasive species in  
 201 sclerophyll woodland and coastal sage scrub occupied a position of greater PPUE end in PC2 than native

202 species, whereas in acid sands fynbos invasive species occupied a position of greater N<sub>mass</sub> (mass-based leaf  
203 nitrogen content), P<sub>mass</sub> (mass-based leaf phosphorus content) and WUE (p<0.05 in all cases) (Fig. 2B).  
204 Native species in sclerophyll woodland were on average taller than invasive species (F<sub>1,18</sub>=5.34, p=0.03).

205 **Fig. 2 a** Distribution of native (blue triangles) and invasive (red dots) species along PC1 (leaf economics  
206 spectrum) within communities. Greater values of PC1 correspond to the high A<sub>mass</sub> (mass-based  
207 photosynthetic rate) and PNUE (photosynthetic nitrogen-use efficiency) end. **b** Distribution of native and  
208 invasive species along PC2 (resource use efficiency). Greater values of PC2 correspond to high PPUE  
209 (photosynthetic phosphorus-use efficiency). **c** Distribution of native and invasive species along PC3 (plant  
210 height). Greater values of PC3 correspond to the high height end. Values indicate estimates ± standard error of  
211 linear mixed models with species as random effect. Asterisks denote significant differences between native and  
212 invasive species for a given community and principal component (p-value < 0.05).

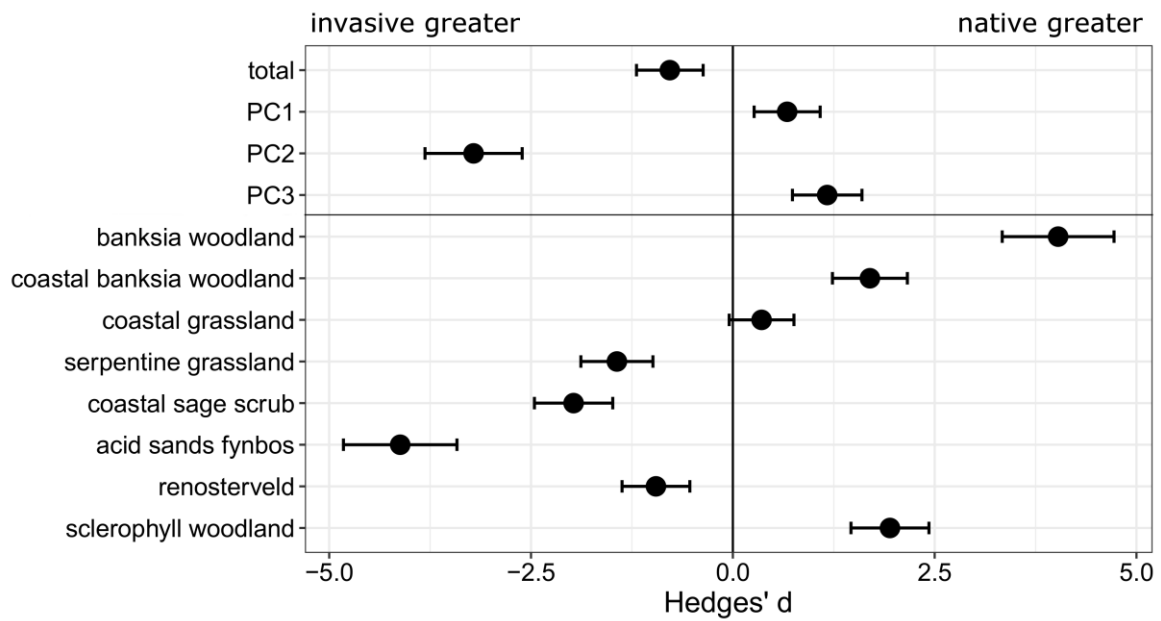


213

214 Functional diversity of native and invasive species

215 Overall, invasive species showed 9.65% greater functional diversity in trait-space respect to natives (natives:  
216 231.28 SD<sup>3</sup> ±5.8, invasives: 253.6 SD<sup>3</sup> ±2.3, mean ± standard error). Particularly, invasive species were  
217 26.02% more diverse than native species in PC2 (water and phosphorous use efficiency; nat: 7.84 SD ±0.09,  
218 inv: 9.98 SD ±0.10), whereas native species showed 4.80% greater functional diversity than invasive species in  
219 PC1 (leaf economics spectrum; nat: 10.49 SD ±0.14, inv: 9.99 SD ±0.04) and 8.51% in PC3 (height; nat: 7.27  
220 SD ±0.09, inv: 6.70 SD ±0.03) (Fig. 3). At the community level, invasive species showed greater functional  
221 diversity than natives in serpentine grassland (nat: 50.41 SD<sup>3</sup> ±5.46, inv: 90.19 SD<sup>3</sup> ±0.99), coastal sage scrub  
222 (nat: 83.16 SD<sup>3</sup> ±2.33, inv: 108.02 SD<sup>3</sup> ±0.98), acid sand fynbos (nat: 35.58 SD<sup>3</sup> ±1.58, inv: 78.16 SD<sup>3</sup> ±1.34)  
223 and renosterveld (nat: 22.41 SD<sup>3</sup> ±1.23, inv: 28.69 SD<sup>3</sup> ±0.48) (Fig. 3). Native species showed greater  
224 functional diversity than invasive species in banksia woodland (nat: 103.44 SD<sup>3</sup> ±2.55, inv: 50.58 SD<sup>3</sup> ±0.66),  
225 coastal banksia woodland (nat: 104.94 SD<sup>3</sup> ±2.75, inv: 78.17 SD<sup>3</sup> ±1.56) and sclerophyll woodland (nat:  
226 170.20 SD<sup>3</sup> ±9.25, inv: 79.42 SD<sup>3</sup> ±1.33). Native and invasive species within coastal grassland were equally  
227 diverse (nat: 115.53 SD<sup>3</sup> ±3.57, inv: 109.14 SD<sup>3</sup> ±0.45).

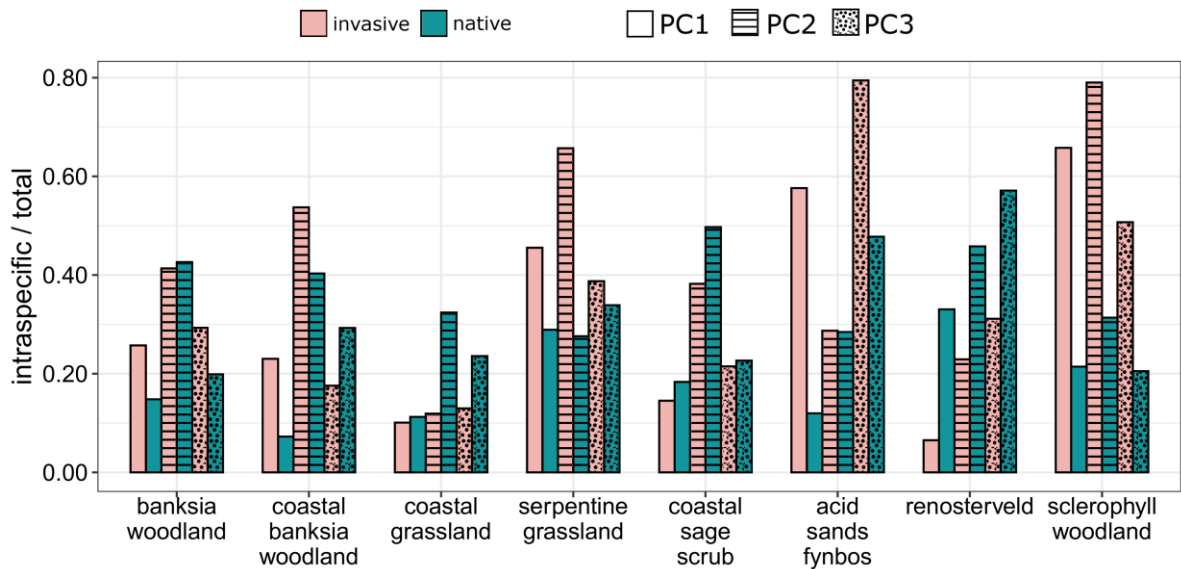
228 **Fig. 3** Mean effect size (Hedges' d) and bias-corrected 95%-bootstrap confidence intervals for differences in  
229 the native and invasive species hypervolumes for the total pool of native and invasive species in trait-space, the  
230 total pool of native and invasive species in each dimension, and native and invasive species in trait-space  
231 within each community. Hypervolume sizes are included in Table A2. Negative mean effect sizes indicate that  
232 invasive species had on average greater hypervolume size than natives. A mean effect size is significantly  
233 different from zero when its confidence interval does not bracket zero.



234

235 The mean relative contribution of intraspecific diversity across the dataset was on average greater in invasive  
 236 than native species for all principal components (PC1: 25.4% and 10.7%, PC2: 23.9% and 18.5%, PC3: 26.6%  
 237 and 18.9%). At the community level, the relative contribution of intraspecific variation to PC1 trait diversity  
 238 was greater for invasive than native species in five communities, and only greater for native species in  
 239 renosterveld (Fig. 4). The relative contribution of intraspecific variation to PC2 trait diversity was greater for  
 240 invasive than native species in three communities, and greater for native species in three communities. The  
 241 relative contribution of intraspecific variation to PC3 trait diversity was greater for invasive than native species  
 242 in four communities, and greater for native species in two communities.

243 **Fig. 4** Relative contribution of intraspecific variance of native and invasive species to total diversity for  
 244 principal components with eigenvalues over one.



245

246 Overall, woody species ( $228.3 \text{ SD}^3 \pm 2.9$ , mean  $\pm$  standard error) showed 18.05% greater diversity than  
 247 herbaceous perennial species ( $193.4 \text{ SD}^3 \pm 3.2$ ) in trait-space, and herbaceous perennial species were 14.85%  
 248 more diverse than annual species ( $168.4 \text{ SD}^3 \pm 1.8$ ) (effect sizes in Table A3). Herbaceous perennial species  
 249 ( $10.8 \text{ SD} \pm 0.1$ ) were significantly more diverse than the other life forms in PC1, whereas annual ( $8.9 \text{ SD} \pm 0.1$ )  
 250 and woody species ( $8.8 \text{ SD} \pm 0.1$ ) were equally diverse. In PC2, annual species ( $8.9 \text{ SD} \pm 0.1$ ) were significantly  
 251 more diverse than the other life forms, whereas woody ( $8.3 \text{ SD} \pm 0.1$ ) and herbaceous perennial species ( $8.5 \text{ SD}$   
 252  $\pm 0.1$ ) were equally diverse. In PC3, woody species ( $7.5 \text{ SD} \pm 0.1$ ) were significantly more diverse than  
 253 herbaceous perennials ( $5.8 \text{ SD} \pm 0.1$ ), and herbaceous perennials were significantly more diverse than annual  
 254 species ( $5.5 \text{ SD} \pm 0.1$ ). In PC1, intraspecific variation accounted for 31.3% of total diversity of annual species,  
 255 11.2% for herbaceous perennial species, and 10.5% for woody species. In PC2, intraspecific variation  
 256 accounted for 27.7% of annual species trait diversity, 29.9% for herbaceous perennial species, and 15.6% for  
 257 woody species. In PC3, intraspecific variation accounted for 37.0% for annual species trait diversity, 23.2% for  
 258 herbaceous perennial species, and 19.7% for woody species.

259 **Discussion**

260 Functional strategies of native and invasive species

261 We found that trait-space was mainly dominated by three axes of trait covariation: leaf economics, water- and  
 262 phosphorus-use efficiency, and plant height. This supports the idea that certain suites of traits tend to vary

263 together (Laughlin 2014), and plant functioning is constrained to a determined range of viable combinations  
264 (Díaz et al. 2016; Lloret et al. 2016). In line with other studies, we found great functional distinctiveness in  
265 resource use strategies between invasive species and native species of the recipient community (Pyšek and  
266 Richardson 2007; Ordonez et al. 2010; Tecco et al. 2010). These results suggest that contrasting resource use  
267 strategies are important in driving the establishment of invasive plants (e.g. Funk et al. 2016; Helsen et al.  
268 2020). In fact, our results confirm previous evidence that the leaf economics spectrum is a main axis of  
269 functional differentiation between native and invasive species across Mediterranean communities (Tordoni et  
270 al. 2019). That is, the range of trait values displayed by the invasive species is consistent with the ‘fast return  
271 on investments’ end of the leaf economics spectrum not only as a general trend, but also in most of the  
272 communities studied separately (significantly different in six of them). These results indicate that higher  
273 capacity for the extraction of resources could be an advantageous strategy for invasive species (Daehler 2003;  
274 Ordonez et al. 2010; Funk et al. 2016; Henn et al. 2019); especially in habitats with strong abiotic constraints  
275 such as those in Mediterranean regions, where native species display conservative resource uptake adaptations  
276 (e.g. Lloret et al. 2016; de la Riva et al. 2017).

277 It is worth noting that we detected water and phosphorous availability as another trait dimension related to  
278 resource constraints. Thus, the particular segregation of native and invasive species across these communities  
279 might depend also on specific resource limitations of each community. We found functional differences  
280 between native and invasive species with respect to this dimension in sclerophyll woodland (Chile), coastal  
281 sage scrub (California), and acid sands fynbos (South Africa). Sclerophyll woodland and coastal sage scrub  
282 have the lowest mean annual precipitation and the highest soil phosphorus concentration compared to the other  
283 Mediterranean communities. In these communities, native species showed higher water-use efficiency and leaf  
284 nutrient concentration than invasive species. It is likely that native species, more frequently perennial species,  
285 benefit from having a higher stomatal control, and root systems to enhance nutrient uptake than annuals (Pérez-  
286 Ramos et al. 2013; Prieto et al. 2018; Tordoni et al. 2019). By contrast, acid sands fynbos is the community  
287 with the lowest phosphorus and nitrogen concentration in the soil, and invasive species occupy the high leaf  
288 nutrient concentration and WUE end of the spectrum. These results concur with those from studies of  
289 Mediterranean grassland species that found higher WUE in invasive species compared to natives (Vaughn et  
290 al. 2011). Higher leaf nutrient concentrations in invasive species in our dataset might reflect their fast resource  
291 uptake strategy, particularly during periods of high water availability (Pérez-Ramos et al. 2013). Collectively,

292 these results support the idea that trait variation depends on the specific combination of environmental factors  
293 and highlights the utility of such studies for predicting plant and community responses in a changing world  
294 (Funk et al. 2016).

295 Disentangling the functional trait diversity of native and invasive species

296 Overall, our results indicate that dominant invasive species are more diverse than native species across  
297 communities and in four out of the eight communities, which indicates that many resource-use profiles might  
298 allow species to invade communities (Tecco et al. 2010; de la Riva et al. 2019). That is, contrasting with  
299 previous findings (Okimura and Mori 2018), we show that invasive species constitute a functionally diverse  
300 pool which contributes to a great proportion of the community trait diversity. As discussed by Loiola et al.  
301 (2018), this may reflect that invasive species are creating new functional spaces outside the extant native pool,  
302 i.e. excluding functionally similar natives or occupying empty space, or filling empty gaps within the existing  
303 space. Yet, our approach does not allow us to identify these underlying mechanisms. Interestingly, although the  
304 total pool of perennial species was more diverse in trait-space than annuals, the pool of invasive species (most  
305 frequently annuals) was more diverse than natives. The sources of functional diversity for native and invasive  
306 species differ: greater diversity in water- and phosphorus-use efficiency for invasives and greater diversity in  
307 leaf economics and plant height for natives. In this regard, the higher hypervolumes of invasive species in trait-  
308 space support the importance of trait distinctiveness between invasive species and the native community  
309 (Helsen et al. 2020).

310 Overall, in agreement with previous findings (Albert et al. 2011; Siefert et al. 2015), our results indicate  
311 intraspecific variance contributed 10.7 to 26.6% of the total functional diversity. On average, the contribution  
312 of intraspecific variance was greater in invasive compared to native species. Regarding the dimensions related  
313 to leaf economics and plant height, where native species were more diverse than invasive species, the greater  
314 contribution of intraspecific variance in invasive species might reflect smaller interspecific differences.  
315 Interestingly, we show that annual and woody species were equally diverse in leaf economics traits, but the  
316 relative contribution of intraspecific variation was three times greater in annual species, which might reflect  
317 that annual species have leaves less structurally and physiologically constrained than woody species (Maire et  
318 al. 2013).



319 With regard to the dimension related to water- and phosphorous-use efficiency, invasive species were overall  
320 more diverse than native species and the contribution of intraspecific trait variance was 5.4% greater. This  
321 supports the theory that invasive species might be physiologically less constrained in trait-space than their  
322 native counterparts (Funk 2008; Valliere 2019). In contrast to native species, invasive species might tend to be  
323 generalists that perform well along ecological gradients (Gallien et al. 2010; Clavel et al. 2011). While this  
324 might reflect to some extent the dependency between origin and life form, studies of co-occurring annual  
325 species have found that invasives do not adhere to the same trade-off between growth and water conservation  
326 displayed in natives (Angert et al. 2009). The contribution of intraspecific trait variation to total diversity was  
327 twice as much in annual and herbaceous perennial species than woody species. It is likely due to the fact that  
328 herbaceous species are more responsive to microhabitat heterogeneity in water and soil nutrient availability  
329 (Chapin et al. 1990; Fernández-Alés et al. 1993). Thus, contrasting structures of inter- and intraspecific trait  
330 variance might provide interesting insights about functional responses of native and invasive species operating  
331 in different communities (Hulme and Bernard-Verdier 2018), supporting the notion that overlooking  
332 intraspecific variability might compromise the ability to correctly infer trait-driven ecological processes  
333 (Helsen et al. 2020).

334 One aspect of our study that warrants further comment is that the above-mentioned patterns were strongly  
335 context dependent. In fact, across most communities intraspecific trait variance was especially important in the  
336 second principal component, related to nutrient availability, which is in line with previous evidence highlights  
337 how microhabitat heterogeneity might shape leaf attributes (Albert et al. 2010; Jung et al. 2010; Siefert et al.  
338 2015). Invasive species were functionally more diverse than native species in acid sands fynbos and serpentine  
339 grassland, in some extent due to greater intraspecific variability. The great diversity of invasive species within  
340 these communities, and the substantial contribution of intraspecific variance to this diversity, points to invasive  
341 species being generalists with great phenotypic variation (Sultan 2001; Funk 2008; Clavel et al. 2011;  
342 Davidson et al. 2011; Martín-Forés et al. 2017), and more phenotypic variability associated with annual and  
343 herbaceous perennial species (Maire et al. 2013). In renosterveld and coastal sage scrub, the contribution of  
344 intraspecific variance to total diversity was comparable between invasive and native species, and reflected to  
345 some extent that invasive species were overall more diverse. By contrast, native species in banksia and  
346 sclerophyll woodlands were more functionally diverse than invasive species. Native species in these  
347 communities are more frequently woody perennials, much more diverse in plant height than invasive species

348 (more often herbaceous). Hence, functional trait diversity, and the contribution of intra- and interspecific  
349 variation to this diversity, varies among communities. The regional species pools are generally the result of the  
350 environmental and historical filtering, while some specific adaptations of invasive species allow them to  
351 establish successfully at a broad spatial scale with a completely different morpho-physiological profile to the  
352 native pool (Loiola et al. 2018).

### 353 Conclusions

354 We applied a multidimensional analysis of plant traits and trait variation to understand the segregation between  
355 native and invasive species in trait-space across Mediterranean communities. We found great functional  
356 distinctiveness in leaf economic traits and that invasive species can be more functionally diverse than native  
357 species. Therefore, our results suggest that invasive species can employ different functional profiles to thrive in  
358 recipient communities. In some communities, intraspecific trait variation can contribute a great proportion of  
359 functional trait diversity in invasive species, which might reflect evolutionary and life form differences. These  
360 results highlight that intraspecific variation might not be equally distributed between native and invasive  
361 species, offering interesting insights to species functional trait diversity within communities.

362 **Acknowledgments.** We thank Rachel Standish, William Stock, and Fernando Valladares for collecting the  
363 data.

364 **Supplementary data.** Table A1: Communities included in this study and their species grouped by origin.  
365 Table A2: Hedges' d and confidence intervals of Figure 3. Table A3: Hedges' d and confidence intervals of the  
366 comparison of the hypervolumes of annual, herbaceous perennial and woody species.

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