

Exploring interactive effects of climate change and exotic pathogens on *Quercus suber* performance: Damage caused by *Phytophthora cinnamomi* varies across contrasting scenarios of soil moisture

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1 **Abstract**

2 Climate change and exotic pests and pathogens are causing alarming forest declines
3 worldwide. However, we still lack a comprehensive understanding of how damage
4 caused by exotic pests and pathogens might vary under the different scenarios of water
5 availability imposed by a changing climate, particularly in water-limited forests as those
6 that occupy Mediterranean areas. In this paper we aimed to experimentally analyse the
7 interactive effects of the aggressive exotic pathogen *Phytophthora cinnamomi* and
8 climate change-related reductions in soil moisture on seedling performance of the
9 Mediterranean host *Quercus suber*. We conducted a full-factorial greenhouse
10 experiment where the physiology and growth of *Q. suber* seedlings was measured in
11 soils with different combinations of *P. cinnamomi* inoculum density (0, 30, 60 and 120
12 colony forming units per gram of dry soil) and soil moisture (15%, 40%, 50% and 100%
13 soil water holding capacity) simulating different invasion and climate change scenarios.
14 We found additive effects of *P. cinnamomi* and drought on *Q. suber* performance
15 aboveground, although these effects were not always negative. In fact, seedlings showed
16 a compensatory physiological response to *P. cinnamomi* infection by increasing their
17 net photosynthetic rates. Our results also supported important interactive effects of
18 pathogens and soil moisture on belowground performance. Thus, the inoculum density
19 in the soil required to cause significant root damage in experimental seedlings decreased
20 as soil moisture increased. From a climate change perspective, these results suggest that
21 an average drier climate might imply sub-optimal conditions for *P. cinnamomi*
22 infections allowing for a slower advance of the disease in invaded areas. However, this
23 effect will be modulated by the also predicted more frequent extreme climatic events. A
24 higher frequency of extreme rain events that saturate the soil might be particularly

25 beneficial for *P. cinnamomi*, boosting its soil density beyond any possible response
26 capacity of susceptible hosts.

27

28 **Keywords:** climate change, drought, invasive pathogens, Mediterranean forests, oak
29 decline, seedling performance.

30

31 **Introduction**

32 Global change drivers such as climate change and the invasion of exotic pests and
33 pathogens are causing alarming forest declines worldwide (Allen et al. 2015, Anderegg
34 et al. 2015). Several studies have related the increase in tree defoliation and mortality
35 rates to extreme drought events and heat waves, as well as to the more gradual and
36 continuous process of temperature increase and rainfall reduction experienced by large
37 parts of the globe (van Mantgem et al. 2009, Carnicer et al. 2011, Sangüesa-Barreda et
38 al. 2015). On the other hand, the number of exotic pests and pathogens and the
39 magnitude of their damages to tree species seem to be increasing at unprecedented rate
40 (Dukes et al. 2009, Santini et al. 2013, Hansen et al. 2015). Among the potential causes
41 behind such increase, the expanding transport of goods and people together with
42 indirect climate change alterations of host-pest and host-pathogen relationships have
43 been proposed to play a fundamental role (Jactel et al. 2012, Trumbore et al. 2015).
44 However, the experimental evidence showing interactive effects of climate change and
45 exotic pests and pathogens on tree species is still very limited, particularly in water-
46 limited forests as those that occupy semi-arid and Mediterranean areas (Pautasso et al.
47 2010, Sturrock et al. 2011).

48 The Mediterranean Basin is considered one of the most susceptible regions to
49 global change around the world (Schröter et al. 2005, Doblas-Miranda et al. 2017). On

50 the one hand, climatic models predict for this region a 2-5°C increase in mean annual
51 temperature by the end of the 21st century, as well as a 30% annual rainfall reduction
52 and a higher frequency of extreme climatic events such as droughts and floodings
53 (Giorgi and Lionello 2008, Lindner et al. 2010, IPCC 2013). On the other hand, exotic
54 pathogenic microbes are causing devastating epidemics throughout the Mediterranean
55 Basin, particularly those belonging to the genus *Phytophthora* (Hansen et al. 2015, Jung
56 et al. 2018). This oomycete genus includes some of the most aggressive pathogens of
57 woody species on earth, as *Phytophthora cinnamomi* Rands. This pathogen has a likely
58 origin in Papua New Guinea (Arentz and Simpson 1986), but it is already introduced in
59 forests worldwide (Sena et al. 2018). *Phytophthora cinnamomi* is a soil-borne pathogen
60 that destroys the fine roots of its hosts, impeding nutrient and water uptake, and leading
61 to defoliation, loss of vigour, and eventual death of infected trees. In the Mediterranean
62 Basin, *P. cinnamomi* is decimating natural populations of dominant evergreen oaks such
63 as *Quercus ilex* and *Quercus suber* (Brasier 1992, Sánchez et al. 2002, Camilo Alves et
64 al. 2013, Ávila et al. 2017). This high mortality rate represents a problem of paramount
65 ecological and social importance, since evergreen oaks are major structural elements in
66 Mediterranean forests and savannah-like ecosystems (“dehesas” and “montados”) that
67 provide extremely valuable biodiversity, economic and cultural services (Marañón et al.
68 2012). Predicting the future viability of Mediterranean oak forests will undoubtedly
69 require a comprehensive understanding of how the severity of the disease caused by *P.*
70 *cinnamomi* might vary under the different scenarios of water availability (the main
71 limiting resource in these forests) imposed by a changing climate.

72 Understanding the interactive effects of exotic soil-borne pathogens and climate
73 change on tree health in water-limited forests is not an easy task, since it will depend on
74 complex responses of host resistance and pathogen growth to decreasing levels of soil

75 moisture. Low values of soil moisture associated with lower precipitation and extreme
76 droughts cause a direct abiotic stress on tree species that strongly reduces their
77 performance (Peñuelas et al. 2013, 2017) and that might make them more susceptible to
78 pathogen attack (Marçais et al. 1993, Corcobado et al. 2014). However, most
79 pathogenic soil-borne fungi and oomycetes require high soil moisture for the
80 germination and dispersal of their spores (Lacey & Harper 1986, Desprez-Loustau et al.
81 2006). Moreover, low water potential of the host cortical tissues has been shown to have
82 a negative direct effect on pathogen growth within the host (Tippett et al. 1987, Marçais
83 et al. 1993). Therefore, it could happen that the severity of the damages caused by
84 exotic soil-borne pathogens to susceptible hosts in a climate change scenario were lower
85 than expected due to direct negative effects of lower soil moisture on pathogen
86 population growth. Since the damage caused by soil-borne pathogens (and by *P.*
87 *cinnamomi* in particular) on tree species is strongly dependent on soil inoculum density
88 (Gómez-Aparicio et al. 2012, Serrano et al. 2015), its maintenance in the soil under the
89 minimum threshold required for disease expression might indirectly favour performance
90 of seedlings and adults of susceptible species.

91 In this paper we aimed to experimentally analyse the interactive effects of the
92 exotic pathogen *P. cinnamomi* and climate change-related reductions in soil moisture on
93 seedling performance of the highly susceptible host *Q. suber*. To achieve this aim, we
94 conducted a full-factorial greenhouse experiment where the growth and physiology of
95 *Q. suber* seedlings was measured in soils with different combinations of *P. cinnamomi*
96 density and soil moisture. We simulated a range of increasing pathogen densities from
97 zero inoculum to large inoculum densities (120 colony forming units per gram of dry
98 soil, CFU/g) known to produce root symptoms in *Q. suber* seedlings under controlled
99 conditions (Serrano et al. 2015). Such large densities, and even much larger, can be

100 found in declining *Q. suber* forests of the Mediterranean Basin (Gómez-Aparicio et al.
101 2012). We then reproduced different scenarios of soil moisture expected under
102 contrasting scenarios of climate change, from very low levels (15% of soil water
103 holding capacity) typical of springs in extremely dry years to saturated soils typical of
104 wet years in Mediterranean forests. We tested three specific hypotheses: 1) pathogen
105 infection and drought will act as multiple stressors of *Q. suber*, both having negative
106 effects on seedling performance; 2) *Q. suber* seedlings will show physiological and
107 morphological adaptive responses to the two stress factors, with the aim of increasing
108 water use efficiency and compensating the root damages caused by *P. cinnamomi* (i.e.
109 lower photosynthetic rate and stomata conductance, higher biomass allocation
110 belowground); and 3) pathogens and drought will have interactive effects on
111 performance of *Q. suber* seedlings, the severity of the damages caused by *P. cinnamomi*
112 increasing with soil moisture. The results of this experiment will greatly contribute to
113 achieving a better understanding of the future of Mediterranean oak forests under
114 interactive effects of global change drivers of yet unclear consequences.

115

116 **Material and Methods**

117 *Experimental design*

118 The experiment was conducted from January to April 2016 in a greenhouse at the
119 University of Córdoba (Córdoba, Spain, 37°51'N, 4°48' W). Average air temperature
120 and moisture during the experiment were $12.3 \pm 3.1^{\circ}\text{C}$ and $77 \pm 13.2\%$ (respectively),
121 varying from $10.6 \pm 2.9^{\circ}\text{C}$ and $89.3 \pm 7.2\%$ in January to $15.7 \pm 2.4^{\circ}\text{C}$ and $66.5 \pm 11.3\%$ at
122 the end of the experiment. On January 13th, one-year old seedlings of *Q. suber* provided
123 by a local nursery (Viveros San Jerónimo, Sevilla, Junta de Andalucía) were
124 individually planted in 3L pots filled with a mixture of sand, silt, clay, peat and soil

125 from a natural *Q. suber* forest free of *P. cinnamomi* (Marismillas de Doñana, Huelva,
126 Spain) in proportions 55:20:10:10:5 v/v. With this substrate we aimed to reproduce the
127 sandy texture and acidic pH typical of *Q. suber* forest soils (Gómez-Aparicio et al.
128 2012), as well as its native microbiota. The absence of *P. cinnamomi* in the soil was
129 tested following Romero et al. (2007). Soil samples (10 g) from three different trees in
130 the area were air dried, sieved and suspended in 100 ml 0.2% sterilized water-agar.
131 Then, 1 ml aliquots were plated on Petri dishes containing NARPH selective medium
132 and incubated. No *P. cinnamomi* colonies were detected under the inverted microscope.

133 Seedlings were assigned to different experimental groups following a full-
134 factorial design with two factors: *P. cinnamomi* inoculum density and soil moisture.
135 Each factor had four levels, resulting in 16 experimental treatments and 160 seedlings (4
136 inoculum densities \times 4 moisture levels \times 10 replicates). Inoculum consisted of *P.*
137 *cinnamomi* chlamydospores in sterile water suspension (isolate PE90) prepared
138 following Sánchez et al. (2002). The four levels of inoculum density were: 1) Non-
139 inoculated seedlings, used as control treatment. These seedlings received 100 ml of
140 inoculum free water; 2) Low inoculum density, corresponding to a soil inoculated with
141 100 ml of water suspension with 2.3×10^3 chlamydospores/ml equivalent to 30 colony
142 forming units per gram of dry soil (CFU/g) (Serrano et al. 2015). This density
143 represents half of the minimum experimental threshold identified in a previous study for
144 root disease expression in *Q. suber* seedlings (Serrano et al. 2015); 3) Medium
145 inoculum density, where each pot received 100 ml of inoculum with 4.6×10^3
146 chlamydospores/ml equivalent to 60 CFU/g, near the minimum inoculum for root
147 disease expression in *Q. suber* seedlings (Serrano et al. 2015); and 4) High *P.*
148 *cinnamomi* density, where each pot received 100 ml of inoculum with 10^4
149 chlamydospores/ml equivalent to 120 CFU/g, enough to cause severe root damage in *Q.*

150 *suber* seedlings (Serrano et al. 2015).

151 The four different soil moisture levels were chosen to simulate soil water
152 availability under contrasting climate change scenarios: 1) Saturated soil at 100% water
153 holding capacity (WHC), simulating a wet spring where soils remained saturated most
154 of the time; 2) 50% WHC, simulating the average soil moisture in spring in *Q. suber*
155 forests of southern Spain (Gómez-Aparicio et al., unpublished data); 3) 40% WHC,
156 simulating a 20% reduction over the previous treatment predicted for 2050 using an
157 ACGCM for the scenario SRES IS92a (Manabe et al. 2004). We used this scenario
158 because, as far as we know, it is the only one that has been used to make predictions in
159 soil moisture; and 4) 15% WHC, simulating an extremely dry spring in *Q. suber* forests
160 of southern Spain (Ávila et al. 2019). Soil moisture was controlled twice per week for
161 every pot, weighting them and watering those which needed it to maintain constant
162 moisture levels. Average volumetric soil water content along the study period was
163 $22.8\pm 0.4\%$, $11.3\pm 0.1\%$, $9.1\pm 0.1\%$ and $3.4\pm 0.1\%$ for the 100% WHC, 50% WHC, 40%
164 WHC and 15% WHC treatments, respectively. Pots were randomly distributed within
165 the greenhouse and repositioned monthly to avoid the effect of possible small
166 differences in environmental conditions. The height of all seedlings was measured
167 before the application of the experimental treatments. Mean initial height of
168 experimental seedlings was 39.46 ± 0.41 cm ($n=160$) and did not differ among levels of
169 pathogen density or soil moisture ($p>0.05$ in both cases).

170

171 *Physiological and morphological seedling measurements*

172 To evaluate the effect of the different treatments on *Q. suber* seedlings, physiological
173 traits were measured during the length of the experiment. Net photosynthetic rate (A_{\max})
174 and stomatal conductance (g_s) were measured three times along the experiment (after 4,

175 9 and 14 weeks of the experimental treatments) using a portable infrared CO₂ gas
176 analyzer (Li6400XT, Li-Cor, Inc., Lincoln, NE, USA) fitted with a 6-cm² cuvette.
177 Environmental conditions were fixed during measurements at 400 ppm of ambient CO₂,
178 air flow of 400 cm³/min, saturating light conditions of 1000 μmol/m²s, and leaf
179 temperature of 20 °C. All measurements were conducted between 11:00 am and 13:30
180 am (GMT). Physiological measurements were made in a subset of four seedlings per
181 treatment (n = 4 seedlings × 16 treatments = 64 seedlings).

182 All seedlings were harvested on April 20th. No seedling died along the course of
183 the experiment. Severity of leaf symptoms was assessed for each seedling using a 0-4
184 scale, according to the percentage of yellow or wilted foliage (0 = 0% necrotic tissue,
185 1 = 1–33%, 2 = 34–66%, 3 = > 67%, 4 = dead tissue) (Sánchez et al. 2002). Root damage
186 was assessed by using the same scale referred to the percentage of necrotic roots
187 (Sánchez et al. 2002). A subsample of necrotic fine roots of seedlings growing in the
188 different treatments was plated on NARPH agar medium for re-isolation of the
189 pathogen. Then each plant was divided into stems, leaves and roots for quantification of
190 morphological traits. All plant material was dried at 70°C for a minimum of 48 h to
191 estimate shoot biomass (sum of stem and leaf biomass), total root biomass, fine root
192 biomass (roots < 0.2 mm of diameter), and root mass fraction (RMF, root dry mass per
193 unit of total plant dry mass). Total biomass (i.e., shoot + root biomass) was strongly
194 correlated (r > 0.95) with shoot biomass, and therefore it was not included in the
195 statistical analyses.

196

197 *Data analysis*

198 Physiological traits (A_{\max} and g_s) were analyzed using repeated measures ANOVA,
199 including pathogen density, soil moisture and its interaction as between-subject factors.

200 Severity of damages and morphological traits were analyzed using Generalized Linear
201 Models (GLMs), also including pathogen density, soil moisture and its interaction as
202 fixed factors. The severity of root damage was not normally distributed, so it was
203 modeled using a Gamma distribution with identity as the link function. The remaining
204 physiological and morphological variables were modeled using a Gaussian distribution
205 with identity as the link function. Initial seedling height was introduced as covariable in
206 all the analyses. When a factor or interaction was significant, differences among levels
207 were tested using post-hoc Tukey tests. All statistical analyses were performed using R
208 version.3.3.2 software (R Core Team 2017), using the package “car” for the analyses
209 (Fox and Weisberg 2011) and the package “ggplot2” for the graphs (Wickham 2009).

210

211 **Results**

212 *Effects of pathogen density and soil moisture on physiological traits*

213 We found a significant effect of pathogen density on net photosynthetic rates, as well as
214 a marginal significant effect on stomatal conductance (Table 1). Seedlings inoculated
215 with *P. cinnamomi* had in general higher net photosynthetic rates (A_{max}) and stomatal
216 conductance (g_s) than non-inoculated seedlings (Figure 1a,b). The increase in the net
217 photosynthetic rate was higher at intermediate inoculum level (60 CFU/g) than at the
218 other two pathogen densities (30 and 120 CFU/g). We did not find any effect of soil
219 moisture on the physiological traits of the experimental seedlings (Table 1).

220

221 *Effects of pathogen density and soil moisture on aboveground morphological traits*

222 Shoot biomass was not affected by either pathogen density or soil moisture (Table 2).
223 However, the two experimental factors have additive effects on the severity of leaf
224 damage (Table 2). Leaf damage increased with the density of *P. cinnamomi* in the soil,

225 being significantly higher at 60 CFU/g and 120 CFU/g than in control seedlings (Figure
226 2a). Leaf damage was higher at the two extreme soil moisture levels (15% and 100%
227 WHC) than at the two intermediate levels (40% and 50%, Figure 2b).

228

229 *Effects of pathogen density and soil moisture on belowground morphological traits*

230 Soil pathogen density and soil moisture had significant interactive effects on the four
231 belowground variables measured (total root biomass, fine root biomass, root mass
232 fraction and root damage), although for fine root biomass the interaction was only
233 marginally significant (Table 2). Total and fine root biomass were affected by soil
234 pathogens only at the highest level of soil moisture, being lower at the highest pathogen
235 density (120 CFU/g) than in control seedlings (Figure 3a,b). Root mass fraction (RMF)
236 was affected by soil pathogens only at the two highest levels of soil moisture (50% and
237 100% WHC). It was highest in control seedlings than in those inoculated with low
238 pathogen density (30 CFU/g), seedlings inoculated with medium (60 CFU/g) and high
239 pathogen density (120 CFU/g) showing intermediate RMF values (Figure 3c). Root
240 damage was influenced by pathogen density at the four soil moisture levels explored
241 (Figure 3d). However, the pathogen density required to cause significant higher damage
242 in experimental than control seedlings decreased as soil moisture increased. Thus, at the
243 lowest soil moisture level (15% WHC), only seedlings inoculated with the highest *P.*
244 *cinnamomi* density (120 CFU/g) showed significantly more root damage than control
245 seedlings. At 40% WHC, significant root damage was detected for seedlings inoculated
246 with 60 CFU/g and 120 CFU/g. Finally, at the highest soil moisture levels (50% and
247 100% WHC), all inoculated seedlings showed more damage than control seedlings
248 independently of inoculum density. *P. cinnamomi* was re-isolated from the roots of
249 seedlings growing in infested soil but never from control seedlings.

250

251 **Discussion**

252 Global change drivers are known to act simultaneously, although many of their
253 combined effects on natural ecosystems still remain unknown (Didham et al. 2007,
254 Tylianakis et al. 2008). The current increase of tree decline and mortality around the
255 world and the increasing evidence of climate-change related drought and exotic
256 pathogens as primary drivers of these processes makes it particularly important
257 understanding the existence of independent vs. interactive effects of these drivers on
258 tree health. In order to further this knowledge, in this study we used a complex
259 experimental approach where physiological and morphological traits of *Q. suber*
260 seedlings were measured under 16 different contrasted combinations of *P. cinnamomi*
261 density and soil moisture simulating different invasion and climate change scenarios.
262 Our results support the existence of additive negative effects of both exotic pathogens
263 and drought on seedling performance aboveground, but also showed the existence of
264 important interactive effects on belowground root performance that might strongly
265 determine the rate of decline of *Q. suber* populations in Mediterranean forests and
266 dehesas invaded by *P. cinnamomi*.

267

268 *Additive negative effects of Phytophthora cinnamomi and soil moisture on physiological*
269 *and aboveground morphological traits*

270 We found additive effects of both *P. cinnamomi* density and soil moisture on
271 ecophysiological and aboveground morphological traits of *Q. suber* seedlings.
272 However, contrary to our first hypothesis, such effects were not always negative. Thus,
273 seedlings infected by *P. cinnamomi* showed higher net photosynthetic rates and
274 stomatal conductance than control seedlings. This result was quite unexpected, since the

275 few studies that have explored to date the physiological response of *Quercus* species to
276 *P. cinnamomi* infection had generally found a reduction in stomatal conductance (Luque
277 et al. 1999, Robin et al. 2001, Maurel et al. 2001) and net photosynthetic rate (Sghaier-
278 Hammami et al. 2013), a response similar to that induced by drought and aimed to
279 decrease water loss. Here, on the contrary, seedlings showed a compensatory
280 physiological response to *P. cinnamomi* infection by increasing their net photosynthetic
281 rates and to a lower extent their stomatal conductance, contrary to our second
282 hypothesis. This physiological response could reflect an attempt of *Q. suber* seedlings
283 to compensate the loss of root functionality by increasing carbon uptake in the presence
284 of *P. cinnamomi*, as has been reported for other host-pathogen interactions (Walters
285 2015). However, the net photosynthetic rate of inoculated seedlings did not increase
286 linearly with the density of *P. cinnamomi* soil inoculum. On the contrary, such
287 compensatory response was maximum at intermediate pathogen density, seedlings
288 inoculated with 60 CFU/g showing almost twice higher net photosynthetic rates than
289 control seedlings, and decreased at higher pathogen densities (Figure 1a). Overall, these
290 results suggest that susceptible hosts as *Q. suber* can have certain physiological
291 plasticity to counteract the negative effects of an aggressive pathogen as *P. cinnamomi*,
292 but that once overpassed certain thresholds of pathogen density the damage imposed to
293 the root system might be too high for the seedlings to compensate it through adaptive
294 physiological responses.

295 We did not find significant effects of soil moisture on the physiological
296 performance of *Q. suber* seedlings, despite the fact that we reproduced four contrasted
297 soil moisture scenarios ranging from a very low soil water content (15% WHC) to a
298 saturated soil (100% WHC). This result likely reflects that *Q. suber* is a drought-tolerant
299 Mediterranean species evolutionary adapted to cope with low levels of soil water

300 availability under long periods of time (Ramírez-Valiente et al. 2009, Gil-Pelegrín et al.
301 2017). Soil moisture did however affect the severity of leaf damage shown by the
302 seedlings at the end of the experiment, although the severity of the damage did not show
303 a linear response to the decrease in soil moisture. On the contrary, leaf damage was
304 maximal at the two extremes of the soil moisture gradient, showing how the lack of
305 water can be as detrimental as its excess (Gómez-Aparicio et al. 2008, Pérez-Ramos and
306 Marañón 2009). Soil saturation can generate hypoxia conditions that translate into a
307 poor root functioning and consequent wilting of some part of the foliage (Sairam et al.
308 2008). Overall, our results demonstrated that *P. cinnamomi* and soil moisture had
309 additive and non-linear effects on aboveground seedling performance that varied in sign
310 and magnitude depending on the trait considered. However, these effects were in
311 general of much lower magnitude than those detected belowground, likely due to a
312 delay between pathogen infection and aboveground symptoms (Robin et al. 1998,
313 Moreira et al. 2000, Ruiz-Gómez et al. 2018), which suggests that understanding the
314 interactive effects of global change drivers on plant performance requires a close look
315 belowground.

316

317 *Interactive effects of Phytophthora cinnamomi and soil moisture on belowground*
318 *morphological traits*

319 A main result of our study is that pathogen density and soil moisture had interactive
320 effects on the belowground performance of *Q. suber* seedlings, in agreement with our
321 third hypothesis. Thus, the severity of the root damage caused by the aggressive soil-
322 borne pathogen *P. cinnamomi* to *Q. suber* seedlings differed substantially among the
323 four soil moisture levels used as surrogates of different climate change scenarios. As
324 predicted, the largest levels of root damage occurred in general at the highest levels of

325 soil moisture (50% and 100% WHC), translating into a somewhat lower RMF in
326 infected seedlings compared to control seedlings. Pathogen effects on total and fine root
327 biomass were detected only at the highest soil moisture level (100% WHC), where
328 highly infected seedlings had on average 22% and 61% lower total and fine root
329 biomass (respectively) than control seedlings (Figure 3a,b). Because of the poor ability
330 of *P. cinnamomi* to degrade the lignocellulose complex of mature roots (Cahill and
331 McComb 1992, Nicoski 1996), it is likely that our estimation of root biomass included
332 damaged root tissue not yet totally decomposed, therefore representing a more
333 conservative estimate of the belowground effects of *P. cinnamomi* than the direct visual
334 estimation of necrotic root tissue.

335 The fact that the largest negative effects of *P. cinnamomi* on root biomass
336 occurred under soil water saturation conditions is consistent with previous knowledge
337 that shows high soil water levels to favor the infective inoculum build-up of *P.*
338 *cinnamomi*. High soil moisture and free water favors sporangial production and
339 zoospore release, and increases zoospore mobility through the soil to infect roots,
340 allowing for rapid secondary cycles and multiple infections on the host roots (see
341 Camilo-Alves et al. 2013 and references therein). However, our results also suggest that
342 the pathogen capacity to cause root damage can be sensitive to moderate reductions in
343 soil moisture, in agreement with previous studies that have shown water restriction to
344 reduce root damage induced by *P. cinnamomi* (Tippett et al. 1987, 1989, Maurel et al.
345 2001). Thus, whereas in the 100% and 50% WHC scenarios a low density of *P.*
346 *cinnamomi* (30 CFU/g) was enough to cause significant higher root damage in infected
347 than control *Q. suber* seedlings, in the 40% WHC scenario only seedlings infected with
348 moderate and high densities (>60 CFU/g) differed from control seedlings. This implies
349 that even a slight decrease of soil moisture simulating a 30% reduction in annual

350 precipitation was enough to trigger differences in root damage caused by *P. cinnamomi*.
351 Further decreasing the soil water content to 15% WHC increased even further (>120
352 CFU/g) the minimum threshold required to cause significant damage to the root system.
353 This negative relationship between soil moisture and the inoculum threshold required to
354 cause disease is consistent with the more acute decline of *Quercus* species observed in
355 fine textured soils with high water retention compared to well-drained sandy soils
356 (Gómez-Aparicio et al. 2012, Corcobado et al. 2013). From a global change perspective,
357 these results imply that disease trajectories of forests already invaded by *P. cinnamomi*
358 will strongly depend on rainfall regimes and associated variations in soil moisture that
359 might exert a significant control on the growth rates of *P. cinnamomi* populations.

360

361 *Conclusions*

362 Understanding how climate change might affect plant disease epidemics is a current
363 challenge, since the different components of climate change (i.e. changes in average
364 temperature and precipitation, increase of extreme events) can have contrasting effects
365 on the dynamics of pathogen populations. To date, the most frequently recognized
366 consequence of climate change on plant diseases has been the positive effect of a
367 warmer climate due to enhance reproductive capacity and survival of pathogenic species
368 and their expansion towards cold areas (Trumbore et al. 2015, Burgess et al. 2017).
369 Therefore, it is often assumed that plant disease epidemics might become more frequent
370 as climate changes (Sturrock et al. 2011). However, in systems like southern European
371 forests where the most severe abiotic stress is undoubtedly drought, changes in
372 precipitation and associated levels of soil moisture might be the most relevant factors
373 driving alterations in the rate and magnitude of tree diseases, with yet unforeseen
374 consequences. To our knowledge, this study addressed for the first time the

375 consequences of different realistic scenarios of soil moisture on the damages caused by
376 the aggressive exotic pathogen *P. cinnamomi* at different levels of soil inoculum that
377 might represent different stages of invasion (from non-invaded soils to highly invaded
378 areas). Although results involving experimental seedlings under controlled conditions
379 need to be carefully extrapolated to adult trees under natural conditions, our results
380 strongly suggest that the capacity of *P. cinnamomi* to cause disease is modulated by
381 even small variations in soil moisture, and that a drier climate might imply sub-optimal
382 conditions for root infections, lengthening the time required for disease expression in
383 susceptible hosts. This finding, together with the physiological plasticity shown by *Q.*
384 *suber* to counteract pathogen effects under moderate inoculum abundance, might allow
385 for a slower advance of the root rot caused by *P. cinnamomi* in a drier future.

386 However, it is important to take into account that the consequences of an
387 average drier climate on disease dynamics will be strongly modulated by the effects of
388 the also predicted most frequent extreme climatic events (droughts and floodings).
389 Together with previous studies (Robin et al. 2001, Corcobado et al. 2014), our results
390 suggest that a higher frequency of extreme rain events that saturate the soil and cause
391 temporal waterlogging will be particularly beneficial for *P. cinnamomi* infections,
392 boosting its density beyond any possible response capacity of susceptible hosts.
393 Because of this, there is an urgent need to adopt management strategies that contribute
394 to maintain soil inoculum densities in already invaded forests as low as possible for the
395 longest period of time. Once a high *P. cinnamomi* density is achieved in the soil, severe
396 damage to susceptible hosts like the evergreen oaks that dominate Mediterranean forests
397 and dehesas will be caused independently of the soil moisture level, with catastrophic
398 consequences for the long term conservation of these valuable systems.

399

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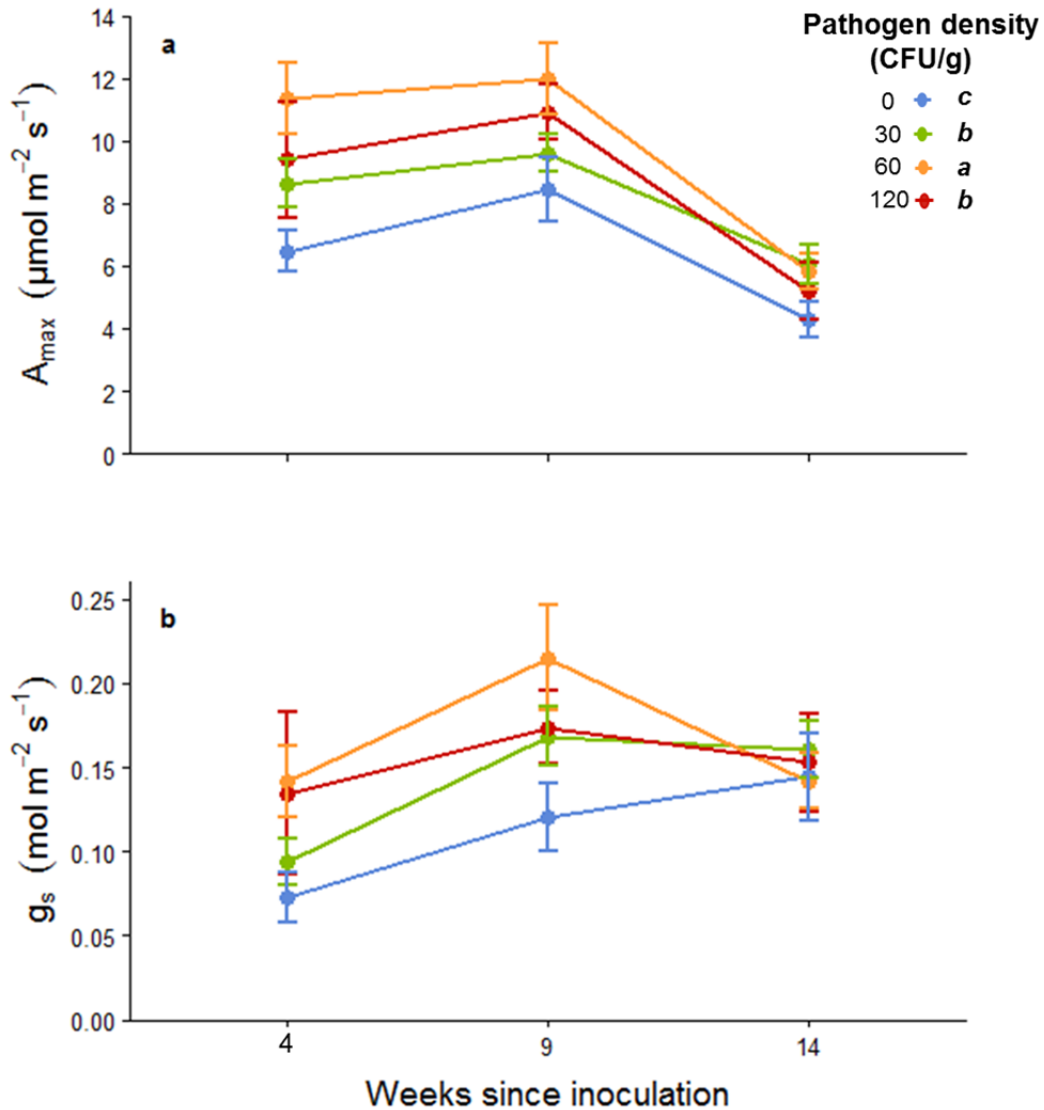


Figure 1. Net photosynthetic rate (A_{max}) and stomata conductance (g_s) of *Quercus suber* seedlings under the four different treatments of pathogen density (measured as number of colony forming units per gram of dry soil, CFU/g) along the course of the experiment. Different letters in the legend indicate significant differences among treatments after Tukey tests. The effect of pathogen density on stomata conductance was only marginally significant (Table 1). Bars represent means \pm SE (n=6).

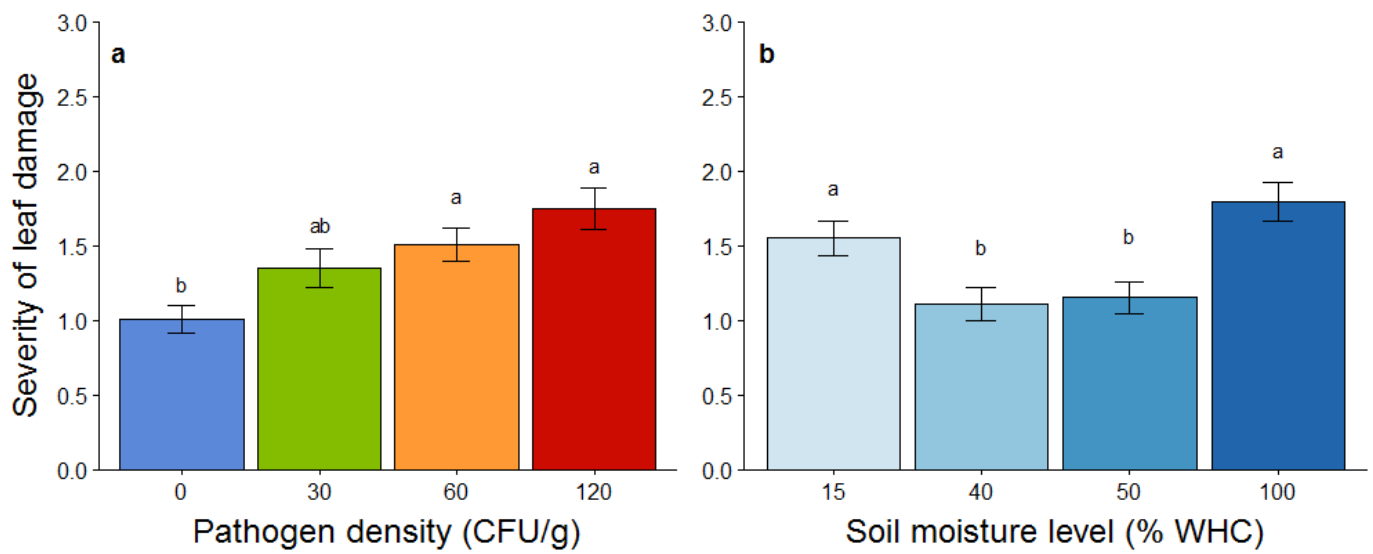


Figure 2. Effects of pathogen density (measured as number of colony forming units per gram of dry soil, CFU/g) and soil moisture (measured as % of water holding capacity, WHC) on the severity of leaf damage of *Quercus suber* seedlings. Different letters show significant differences after Tukey tests. Bars represent means±SE (n=10).

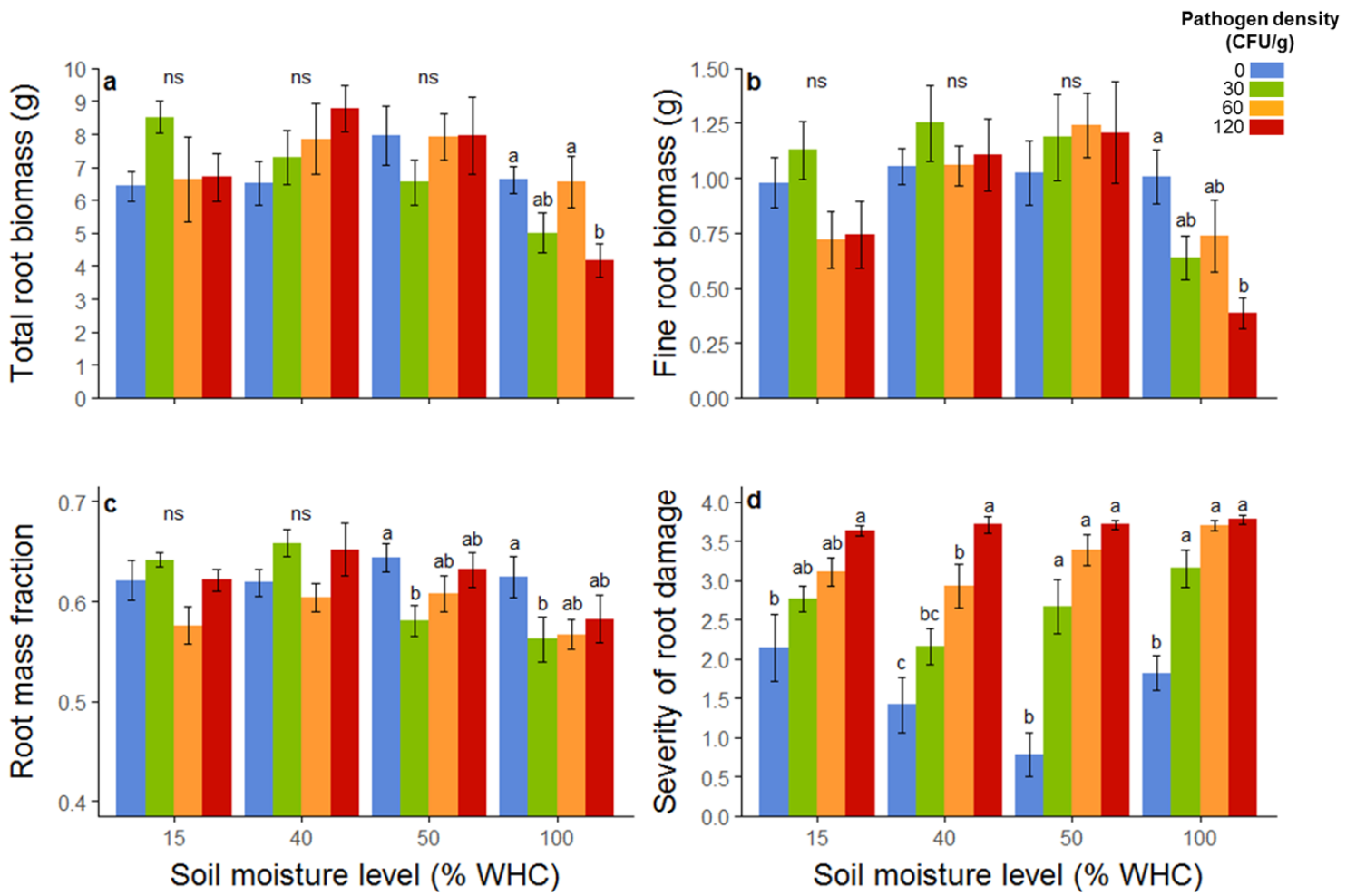


Figure 3. Interactive effects of pathogen density (measured as number of colony forming units per gram of dry soil, CFU/g) and soil moisture (measured as % of water holding capacity, WHC) on belowground morphological traits of *Quercus suber* seedlings. Different letters show significant differences after Tukey tests. Bars represent means ± SE (n=10).

Table 1. Results of the repeated measures ANOVA (rmANOVA) analysis for the influence of pathogen density, soil moisture, and its interaction on physiological traits of *Q. suber* seedlings over time. Variables analyzed were net photosynthetic rate (A_{\max}) and stomatal conductance (g_s).

Variable	Factors	df Num	df Den	F	P-value	
A_{\max}	<i>Between-subjects source</i>					
	Pathogen density (PD)	3	47	4.59	0.007	
	Soil moisture (SM)	3	47	1.81	0.16	
	PD x SM	9	47	0.67	0.73	
	Error	15	47	1.65	0.10	
	<i>Within-subjects source</i>					
	Time (T)	2	46	49.22	<0.0001	
	T × PD	6	92	1.39	0.22	
	T × SM	6	92	1.1	0.37	
	T × PD × SM	6	92	0.6	0.89	
	Error	30	92	0.87	0.65	
	g_s	<i>Between-subjects source</i>				
		PD	3	47	2.25	0.09
SM		3	47	1.54	0.22	
PD × SM		9	47	0.6	0.79	
Error		15	47	1.10	0.38	
<i>Within-subjects source</i>						
T		2	46	21.36	<0.0001	
T × PD		6	92	1.17	0.42	
T × SM		6	92	1.05	0.33	
T × PD × SM		6	92	0.64	0.85	
Error		30	92	0.86	0.68	

Table 2. Results of the statistical analyses (GLMs) performed to test the effect of the experimental treatments and its interaction on the morphological traits of *Q. suber* seedlings. Significant differences among treatments ($p < 0.05$) are highlighted in bold, and marginal differences ($p < 0.10$) in italics.

Treatment		Morphological traits					
		Aboveground traits		Belowground traits			
		Shoot biomass	Leaf damage	Total root biomass	Fine root biomass	Root mass fraction	Root damage
Pathogen density (PD)	P-value	0.07	<0.0001	0.93	<i>0.08</i>	0.008	<0.0001
	df	3	3	3	3	3	3
	Deviance	15.34	1.99	0.09	0.39	0.03	16.36
Soil moisture (SM)	P-value	0.33	<0.0001	0.0002	<0.0001	0.001	0.16
	df	3	3	3	3	3	3
	Deviance	7.35	2.19	4.14	1.89	0.04	0.95
PD × SM	P-value	0.37	0.27	0.012	<i>0.07</i>	0.03	0.0005
	df	9	9	9	9	9	9
	Deviance	21.26	0.94	4.52	0.89	0.06	5.4
Initial Height	P-value	0.01	0.52	<i>0.07</i>	0.90	0.03	0.93
	df	1	1	1	1	1	1
	Deviance	12.46	0.03	0.71	0.17	0.01	0.01
	Explained deviance	0.15	0.30	0.24	0.29	0.27	0.23