

Ecological Modelling 87 (1996) 169-179

ECOLOGICAL

Modelling the suspended matter distribution in an estuarine system Application to the Odiel river in southwest Spain

R. Periáñez ^{b,*}, J.M. Abril ^b, M. García-León ^a

^a Dpto. Física Atómica, Molecular y Nuclear, Universidad de Sevilla, Apdo. 1065, 41080-Sevilla, Spain ^b Dpto. Física Aplicada, E.U. Ingeniería Técnica Agrícola, Universidad de Sevilla, Ctra. Utrera km. 1, 41014-Sevilla, Spain

Received 28 March 1994; accepted 25 January 1995

Abstract

A numerical model which solves the advective-diffusive dispersion equation for suspended matter and includes the deposition and resuspension is presented. The model requires the simultaneous solution of the hydrodynamic equations under tidal dynamics and atmospheric forcing, using time steps of a few seconds. The model has been applied to the Odiel river (southwest Spain). The hydrodynamic module has been widely validated for neap and medium tides, whereas the dispersion was calibrated against the dissolved ²²⁶Ra dispersion pattern. This ²²⁶Ra is discharged to the Odiel river from a phosphate fertilizer factory. The model was able to reproduce the observed behaviour of the suspended matter in the estuary. The sedimentation rates have shown that a net, although slow, sedimentation is being produced. Sensitivity tests were inconclusive with respect to parameters describing settling and resuspension, as internal processes within the estuary are overridden by the high influx and efflux of particulate material from the sea.

Keywords: Estuary ecosystems; Hydrodynamics; Sediments; Tides

1. Introduction

An estuary is a semi-enclosed coastal body of water which has a free connection with the open sea. Estuaries form the transition from river to sea, so they are influenced by conditions in the river as well as in the coastal sea. The suspended matter is supplied from various sources: atmo-

* Corresponding author.

sphere, river inflow, erosion of the estuary bottom, shore erosion, coastal sea, waste disposal, runoff etc. Depending on the local situation one or more sources will dominate the supply. In large rivers, with a high suspended load, the river supply dominates. In tidally mixed estuaries the river inflow will dominate in the inner part and the supply from the coastal sea will dominate in the outer part of the estuary.

The knowledge of suspended matter dynamics is particularly important to study the dispersion of non-conservative substances in aquatic envi-

 $^{0304\}text{-}3800/96/\$15.00$ © 1996 Elsevier Science B.V. All rights reserved SSDI 0304-3800(95)00026-7

ronments (Abril and García-León, 1993a,b), since its distribution and interaction with sediments will influence the behaviour of such substances.

Recently two of the authors developed a model to describe the basic aspects of suspended matter dynamics over large time scales (Abril and García-León, 1994). The model was formulated in terms of residual water circulation and mean settling and resuspension velocities. This produces a mean (annually averaged) suspended matter and sedimentation rate distributions.

Here we are going deeper in the short scale (both spatial and temporal) aspects. Thus, we present a mathematical model which solves the advective-diffusive dispersion equation with the resuspension and deposition terms, both depending on the instantaneous water state. The model requires the simultaneous solution of the hydrodynamic equations under tidal dynamics and atmospheric forcing, using time steps of a few seconds. The formulation of the deposition and resuspension processes is more complex than the one used in models which work with residual circulations. Our description allows studying the influence of tidal oscillations in suspended matter concentrations and in sedimentation rates. This information is essential in modelling the dispersion of non conservative substances in a non equilibrium situation: the transfer of such substances among the dissolved and solid phases depends on the instantaneous suspended matter concentration and instantaneous sedimentation rate.

The model has been applied to the Odiel river (southwest Spain), which is an estuarine system affected by tidal dynamics (M_2 is the main component). The Odiel river is surrounded by a marsh area and an industrial complex, in which two fertilizer processing plants are located (see Fig. 1). These factories release part of their wastes directly to the Odiel river and so important concentrations of U-, Th-(Martínez-Aguirre et al., 1994) and Ra-isotopes (Periáñez and García-León, 1993; Periáñez et al., 1994a) have been measured in its waters and suspended matter.

A 2D, with high temporal resolution, model of the suspended matter dynamics is necessary to study the 226 Ra dispersion in the Odiel river,



Fig. 1. Map of the Odiel river (southwest of Spain). The part of the river which is covered by the grid is also shown.

since ionic exchanges must be taken into account. The model must be 2D because the ²²⁶Ra source is very local and, on the other hand, there is a point source of suspended matter in the Odiel river from a mining factory. This produces a bidimensional structure in suspended matter concentrations and sedimentation rates. The high temporal resolution is needed to study the influence of tidal oscillations in the results.

The mathematical model is presented in see section 2 and its application to the Odiel river is shown in see section 3. Finally, results are presented and discussed in see section 4, where several sensitivity tests are also shown. These tests are used to study the model response to changes in the parameters involved in the equations.

2. The model

The estuarine system in which the suspended matter dynamic is to be studied is represented by a grid containing a certain number of compartments or grid-cells. Each compartment has coordinates (x,y) which define its position in the grid and contains a certain concentration m(x,y) of suspended matter in ppm (parts per million). We assume a homogeneous distribution of m in the water column, that is, vertical discretization is not



Fig. 2. (a) Bathymetry of the Odiel river; depths are shown in m. (b) The grid used in the model. (c) Distribution of small particles in the top sediment layer of the Odiel river. $\rho_m f(x,y)/\rho_w$ is given in %.

considered. This assumption is realistic since the estuary is very shallow (see Fig. 2a) and the stream flows are very low, ranging from 4 to 50 m³ s⁻¹ in usual conditions (Periáñez et al., 1994b). Only in the case of very heavy rainfalls stream flows could reach higher values. Nevertheless, the mixing of salt and fresh water takes place far upstream of the studied area. A fast dispersion of the fresh water into a much larger volume of salt water occurs (Borrego and Pendón, 1988), as usual in tidal mixed estuaries of small rivers (Eisma, 1993).

By convention, particulate matter in suspension is defined as the material that is retained on a 0.4-to 0.5- μ m pore size filter. Smaller material is considered to be dissolved (although it may be colloidal or particulate). The upper size limit of particulate matter in suspension is not fixed; heavy particles sink rapidly to the bottom but very large, low-density structures can remain in suspension for a long time. Nevertheless, in order to simulate the suspended matter dynamics, we have to accept a simplification which is usual in this kind of studies. Two size fractions of particulate matter are defined: one with a diameter < 62.5 μ m and one with a diameter > 62.5 μ m. We will consider that the first class can remain as suspended matter; the second class will sink rapidly so it will not be included in this study since its horizontal movement is negligible (Belderson, 1964; Guburt et al., 1987). On the other hand, particles with $\phi > 62.5 \ \mu m$ will not be resuspended from sediments. The water velocity at which particles begin to be resuspended is called the critical resuspension velocity. Its value depends on the particle diameter and the roughness length factor (see section 2.1). Typical values for the roughness length range from 0.02 cm for muds to 0.3 cm for gravels (Pugh, 1987). The critical resuspension velocity increases with the particle diameter (larger particles need more energy to be resuspended) and decreases with the roughness length factor: if the roughness is small the sediment is more cohesive and then it is more difficult to resuspend. Roughness lengths in the Odiel estuary range from some 0.1 to 0.3 cm, thus, water velocities of about 0.25 m s^{-1} are needed to resuspend sediments with $\phi \sim 100 \ \mu m$ (Pugh, 1987). Water velocities in the Odiel river are larger than this value only during a short time, when the water level is increasing or decreasing in the southern part of the river, but as we move

upstream water velocities decrease. Thus, it seems that to be a good approximation to take the limit diameter as $62.5 \ \mu$ m, since larger particles will be resuspended only in the south of the river during very short times.

Several processes contribute to changes in m(x,y): the deposition of particles on the estuary bottom and the resuspension of particles from it because of the drag provided by the turbulent motion of the water. They provoke a vertical movement of m(x,y) inside the compartment. The exchange of matter with the adjoining compartments contributes to a horizontal movement of the suspended matter. As a boundary condition, suspended matter input from waste disposal and/or runoff can take place in the compartment.

2.1. Vertical transport

The suspended matter in a water column of height h falls down with a mean settling velocity v_s . So the sedimentation process contributes to a variation in time of m which, in the case of a homogeneous distribution of m in depth, is:

$$\frac{\partial m}{\partial t} = -\frac{v_{\rm s}}{h} m \left(1 - \frac{|v|}{v_{\rm cs}} \right) \tag{1}$$

if $|v| < v_{cs}$, where v is the depth-averaged horizontal water velocity and v_{cs} is a critical settling velocity. Thus, deposition of suspended matter occurs only when the water velocity is below a critical value, as observed in nature. If $|v| > v_{cs}$, there is no sedimentation due to the turbulence: particles remain in suspension and, on the other hand, flocs are broken. A value of 0.18 m s⁻¹ has been measured for v_{cs} (Eisma, 1993). Below this velocity, deposition of all particles in suspension occurs.

The settling velocity depends on the concentration of suspended matter: clouds of particles settle faster than single particles because a large number of particles in the cloud settle in the wake of others. It has been found (Mentha, 1989) that the settling velocity can be written as:

$$v_{\rm s} = 1.74 \times 10^{-9} \, m^{1.6} \tag{2}$$

if v_s is measured in m s⁻¹ and *m* in ppm. This equation can be used when the concentration of suspended matter is not higher than ~ 1000 ppm because at such concentrations the falling particle is hindered by the other particles in suspension. Indeed, at concentrations of 10⁵ ppm the settling is negligible (Eisma, 1993). Anyway, the actual concentrations remain below ~ 100 ppm.

The resuspension effect, produced by the shear stress, will contribute to the variation of m over time with an amount of (in ppm) (Abril and García-León, 1994):

$$\frac{\partial m}{\partial t} = \frac{v_{\rm r} \,\rho_{\rm m} \,f}{h \,\rho_{\rm w}} 10^6 \tag{3}$$

where $v_{\rm r}$ is the resuspension velocity (m s⁻¹), $\rho_{\rm m}$ is the dry matter density of the sediment, f is the weight fraction of small particulates ($\phi < 62.5$ μ m) in the top layer of the sediment in compartment (x,y) and ρ_w is the water density. Laboratory measurements have shown that as the current speed is gradually increased from zero, there is a speed at which the sediment begins to move. This is called the critical resuspension velocity $v_{\rm cr}$. Its value depends on the roughness length factor, z_0 , which depends on f. For instance, $z_0 = 0.02$ cm for muds and $z_0 = 0.3$ cm for gravels (Pugh, 1987). From data shown in Pugh (1987) we have constructed an analytical function which relates v_{cr} with z_0 for a particle diameter of about 100 μ m (since there is no information for a diameter closer to 62.5 μ m). Thus,

$$v_{\rm cr} = 0.441 \,\mathrm{e}^{-1.117 \,z_0^{0.231}} \tag{4}$$

if v_{cr} is measured in m s⁻¹ and z_0 in cm.

Taking into account the critical resuspension velocity, the resuspension term has been rewritten as:

$$\frac{\partial m}{\partial t} = \frac{v_{\rm r} \,\rho_{\rm m} \,f}{h \,\rho_{\rm w}} 10^6 \left(\frac{|v|}{v_{\rm cr}} - 1\right) \tag{5}$$

This formulation is valid if $|v| > v_{cr}$. If the current speed is not higher than the critical resuspension velocity, the resuspension term will be zero.

As shown in (see section 4), there are no important changes in the model results if v_r is slightly varied, so we have used the value found in the current literature: $v_r = 0.03$ m year⁻¹ (Abril and García-León, 1994).

2.2. Horizontal transport

The suspended matter can remain in the water column for a long time, so it will participate in the water movements and will behave as a conservative substance. The exchange of suspended matter with the neighbouring compartments is expressed by the advective-diffusive dispersion equation (Periáñez et al., 1994b):

$$\frac{\partial m}{\partial t} + u \frac{\partial m}{\partial x} + v \frac{\partial m}{\partial y} = \frac{1}{h} \left[\frac{\partial}{\partial x} \left(h K_u \frac{\partial m}{\partial x} \right) + \frac{\partial}{\partial y} \left(h K_v \frac{\partial m}{\partial y} \right) \right]$$
(6)

where u and v are the components of the water velocity in the direction of the x-and y-axis respectively and K_u and K_v are the diffusion coefficients in the corresponding directions of the x-and y-axis, which can be written as (Periáñez et al., 1994b):

$$K_{u} = \beta_{1} |u| \sqrt{u^{2} + v^{2}}$$

$$K_{v} = \beta_{2} |v| \sqrt{u^{2} + v^{2}}$$

$$(7)$$

where β_1 and β_2 (dimension [T]) have to be calibrated for each specific site.

To solve the advective-diffusive dispersion equation, the water velocities and the height of the water column must be known for each compartment and for each time step. They are obtained by solving simultaneously the hydrodynamic equations (conservation of mass and momentum), which include the Coriolis term, bed friction, response to wind stress and response to changes in atmospheric pressure. Spatial gradients in atmospheric pressure have not been considered because of the small dimensions of our estuarine site (about 4 km length). A complete description of the hydrodynamic equations can be seen in Periáñez et al. (1994b). The equation which describes the dynamic of suspended matter m in compartment (x,y) is obtained by summation of all the above-mentioned terms: advective, diffusive, deposition, resuspension and a source term, which takes into account external sources of suspended matter.

The solution of that equation provides information on the sedimentation processes that take place in the area under study. The sedimentation rate w is obtained as the net balance between sedimentation and resuspension (or erosion). Expressing w in kg m⁻² s⁻¹, we have:

$$w = v_{\rm s} \, m \, \rho_{\rm w} \left(1 - \frac{|v|}{v_{\rm cs}} \right) 10^{-6} - v_{\rm r} \, \rho_{\rm m} \, f \left(\frac{|v|}{v_{\rm cr}} - 1 \right)$$
(8)

3. Application of the model to the Odiel river

3.1. Geological structure of the Odiel river

The Odiel river mouth forms a sedimentary environment. It is an estuarine system affected by tidal dynamics. The outer part of the river was invaded by the sea approximately 5000 years ago, when a big bay was formed over neoquaternary detritus (Borrego and Pendón, 1988). Sediments are introduced in the estuary from several sources:

River inflow: The Odiel river introduces, in the inner part of the estuary, thick-grained sands. This material contains important amounts of iron oxides, which originate from erosion of Palaeo-zoic rocks. Muds are also introduced during tor-rential rain episodes.

Sea supply: In the outer part of the estuary. Siliceous medium and thick-grained sands are observed, as well as carbonated material (principally shells).

Muds: Sea origin, from the flocculation of dissolved particles. This is the greatest amount of material introduced into the estuary.

A map of the part of the Odiel river under study is shown in Fig. 2, where the grid used in our model is also represented. It includes 456 compartments, each of which can be water (white) or land (black). Compartments are described by the spatial coordinates (x, y) and depths were introduced as input data from marine charts. The dimensions of the compartments are $\Delta x = \Delta y =$ 100 m.

The available information for f(x,y) (Universidad de Sevilla, 1991) is given in Fig. 2c, where $\rho_m f(x,y)/\rho_w$ is represented in %.

3.2. Hydrodynamics of the Odiel river

The hydrodynamic equations have been calibrated for the Odiel river for neap and medium tides, as well as the advective-diffusive dispersion equation, which has been used to study the ²²⁶Ra dispersion in the Odiel river treating it, as a first approach, as a conservative substance (Periáñez et al., 1994b,c,d). In these references the values obtained by calibration can be found for parameters which appear in the hydrodynamic and dispersion equations. These include the bed friction coefficient and β_1 and β_2 in the diffusion terms. Boundary conditions are also described in detail.



Fig. 3. Time evolution of the water elevations and velocities along the y-axis in a compartment near the southern border (--) and a compartment in the middle of the grid (--). The differences in elevations are a few mm, so elevations for both compartments are represented by the same curve.



— Low water ● Low water ⁻⁻ High water △ High water

Fig. 4. Model results for ²²⁶Ra dispersion. Points are experimental data (Periáñez and García-León, 1993) and lines are the computed ²²⁶Ra concentrations for high (---) and low (—) water. Values in mBq/l. The x-axis corresponds to the localization in the grid.

For details on the method used to solve these equations we also refer to the references. A centred finite differences scheme was applied. The time step was $\Delta t = 6$ s, so the Courant-Friederich-Lewy criterion (Prandle, 1984) is met and the numerical dispersion is negligible (Periáñez et al., 1994b).

In Fig. 3 the course of the water elevations and velocities along several tidal cycles for medium tides are shown. Water velocities are similar to those experimentally obtained. The calculated maximum water velocities are 0.45 and 0.61 m s⁻¹ when the water level is increasing and decreasing respectively, while the measured ones were 0.48 and 0.66 m s⁻¹. In the case of neap tides, the maximum calculated velocities are 0.38 and 0.29 m s⁻¹, while the measured maximum velocities were 0.40 and 0.28 m s⁻¹ when the water level is increasing respectively.

In Fig. 4 results of the application of the dispersion module to ²²⁶Ra dispersion are compared to field data. As can be seen in Fig. 4 the agreement is rather good; the real peak is more intense than the calculated one because that sample was collected just in the effluent of one of the fertilizer plants, and we are assuming instantaneous homogenization inside each compartment.



Fig. 5. Suspended matter concentrations (ppm) in the northern (a) and southern (b) borders of the grid over a tidal cycle. Elevations are given with respect to an arbitrary reference.

3.3. Boundary conditions and sources of m(x,y)

The boundary conditions for m(x,y) in the northern border of the grid are the ones which were used when the advective-diffusive equation was calibrated. They consists of taking the suspended matter concentration in the last row equal to that of the previous row:

m(x,38) = m(x,37)

Along the southern border, suspended matter concentrations were specified for each time step. These concentrations were experimentally measured in the Odiel river estuary at about 0.5 m below the water surface. To do this, a sampling campaign was performed: water samples were collected every 20 min during a complete tidal cycle in both the southern and northern borders of our grid. The results are shown in Fig. 5. Suspended matter concentrations were measured using a nephelometer, previously calibrated to convert ntu (nephelometric turbidity units) in ppm following a method similar to that of Durrieu de Madron et al. (1992). As said before, suspended matter concentrations in the southern border are used as boundary conditions, while those of the northern border will be used to calibrate the model.

On the left shore of the Odiel river there is an industrial complex. One of the factories located here discharges important amounts of material into the river, so the source term S must be considered. The compartment in which the source term is located is (7,9). In order to evaluate the magnitude of this term, we must take into account that the ²²⁶Ra and U-isotopes activity concentrations measured in sediments collected at the point of discharge are about a factor 10 smaller than the activity concentrations measured in sediments of the surroundings (Martínez-Aguirre et al., 1994). Thus, the sedimentation rate when the discharges of material are performed must be about 10 times larger than when there are no discharges. This is due to the fact that the material discharged is basically Fe, and consequently there is a reduction in the U and Ra content (Respaldiza et al., 1993). These considerations have allowed us to find the source term in a calibration exercise, which is 0.78 kg/s.

4. Results and discussion

4.1. Suspended matter distribution and sedimentation rates

To study the suspended matter distribution, a simulation over several tidal cycles was performed in a medium tide situation.

The course of the suspended matter concentration in three compartments of the grid can be seen in Fig. 6a. As we move towards the northern border, the oscillations in suspended matter concentrations, which are due to the tidal oscillations, decrease. Indeed, the suspended matter concentration in the compartment near the northern border is almost constant. This behaviour has been observed in the Odiel river: Fig. 5 shows that concentrations at the northern border are quite constant, the mean value is 28 ± 4

ppm. The computed mean value is 24.4 ppm, so the agreement is rather good. This behaviour of the suspended matter concentrations along the



Fig. 6. (a) Time evolution of the suspended matter concentrations in three compartments of the grid: near the southern border (—), in the middle of the grid (---) and near the northern border (- · -). (b) Time evolution of the sedimentation rate (g cm⁻² year⁻¹) in a point located in the middle of the grid (—) and in a point near the northern border (---). The moments at which high and low water occur are also shown.



Fig. 7. Suspended matter concentration (ppm) maps when water level is increasing (a) and decreasing (b).

Odiel river is due to the fact that the water velocities decrease as we move towards the northern boundary of the grid (see Periáñez et al., 1994b and Fig. 3), so the resuspension becomes negligible and the deposition must be almost constant. Thus, oscillations in suspended matter concentrations vanish. On the other hand, it can be seen in Fig. 6a that there are two peaks in the compartment near the southern border for each tidal cycle: when approaching high and low water levels, the velocities are higher and there is an important resuspension of matter from the river bed and concentrations reach the maximum values. But during high and low water only deposition occurs and the concentrations decrease, reaching minimum values.

Suspended matter concentration maps when water level is increasing and decreasing can be seen in Fig. 7a and b respectively. The decrease in concentrations as we move towards the north is again clear, as well as the effect of the source of suspended matter, which produces a local concentration increase.

The development of the sedimentation rate in



Fig. 8. Sedimentation rates (g cm^{-2} year⁻¹) averaged over several tidal cycles.

two compartments of the grid is shown in Fig. 6b. The sedimentation rate ranges from some 0.02 to $0.07 \text{ g cm}^{-2} \text{ year}^{-1}$. It remains positive, so there is a net sedimentation in the area although the process is slow, as confirmed by the low values of the sedimentation rates. As can be seen in Fig. 6b there are two peaks for each tidal cycle in the compartment located in the middle of the grid: one during low water and one during high water. In these situations the water velocities are very low, being the deposition the dominant process. In the compartment near the northern border the pattern is the same but the peaks are lower. This is again related to the fact that, due to the low water velocities, the deposition process is almost constant.

The sedimentation rates have been averaged over several tidal cycles all along the Odiel river. The results are shown in Fig. 8, where a map of the Odiel river with the sedimentation rates expressed in g cm⁻² year⁻¹ is shown. The highest value of the sedimentation rate, 0.3 g cm⁻² year⁻¹, corresponds to the point of discharge. In the rest of the river the sedimentation rates are smaller, showing a net deposition of material all over the river bed.

All these results correspond to a medium tide situation. Extreme conditions, such as torrential rains and equinoctial spring tides, can alter the results. These extreme conditions are presently under investigation.

In the case of neap tides, the time evolution of suspended matter concentrations and sedimenta-



Fig. 9. Suspended matter concentration for low and high water (a and b) and time evolution of the sedimentation rate (c) when the source term is doubled $(- \cdot -)$ and when there is not a source term $(- \cdot -)$. The continuous lines (-) are the model results.

tion rates are very similar to those of medium tides, as well as the concentration profiles along the river. The averaged sedimentation rates are a bit higher as a result of the lower water velocities. Thus, at the point of discharge, the averaged sedimentation rate reaches $0.47 \text{ g cm}^{-2} \text{ year}^{-1}$. Anyway, the sedimentology in the Odiel river is not greatly altered.

4.2. Sensitivity tests

The response of the model to changes in parameters of the hydrodynamic and dispersion model has already been studied (Periáñez et al., 1994b). These are the bed friction coefficient and β_1 and β_2 in the diffusion coefficients. The effect of modifying boundary conditions in both the northern and southern borders of the grid is also discussed in that paper. We have now explored the response of the model to changes in the suspended matter source term and in settling and resuspension velocities (and in the critical values).

The response of the model when the source term is doubled and when there is no source term (S(7,9) = 0) is shown in Fig. 9. In the first case, concentrations which are far too high occur in the source, although there are no important changes in the rest of the river. The (averaged) sedimentation rate in the point of the discharges is about 27 times the (averaged) sedimentation rate in that point when the discharges are not being performed. As said before, this factor should be about 10 (see section 3.3). When the source term is zero there is no peak at all, although the concentrations in the rest of the river remains the same again. On the other hand, there are no important changes in the sedimentation rates in the middle of the grid. Thus, we can conclude that the source term affects only the area around the point of discharge, as expected.

The suspended matter profiles along the river when settling velocities are doubled or halved (modifying Eq. 2) are not changed. Only sedimentation fluxes suffer a little variation. Thus, when halving v_s , the maximum values of the sedimentation rate in the middle of the grid is about 0.05 g cm⁻² year⁻¹ and when doubling v_s it is about 0.18 g cm⁻² year⁻¹. There are no valuable changes in the results when $v_{\rm cs}$ is slightly varied. Thus we have selected the value found in the current literature. This must be due to the fact that water velocities are larger than $v_{\rm cs}$ only during short periods in the southern part of the system Elsewhere the sedimentation rate is almost constant.

There are no changes in the concentration profiles or in the sedimentation rates if the resuspension velocity is doubled and halved. This is due to the fact that the suspended matter which enters the estuary from the sea is much greater than the suspended matter which arises from the river bed erosion. Thus, the resuspension velocity used in the model is the one that we have found in current literature. Again, modifying the critical resuspension velocity only produces slight variations in sedimentation rates.

5. Conclusions

The hydrodynamic and the dispersion models based on the advective-diffusive dispersion equation, have been validated in the Odiel estuary. These models have been applied to study the suspended matter distribution as well as the sedimentation processes in the same area. To do this, a sedimentation term and a resuspension term, including sources, had to be included in the equation. The source term was estimated from experimental results on U-isotopes and ²²⁶Ra activity concentrations in sediments collected in the Odiel river. The boundary conditions were obtained from field observations.

The model, with the new formulation of the deposition and resuspension processes, has a resolution of 6 s, which allows a detailed study of the influence of tidal oscillations on suspended matter concentrations and sedimentation rates. This was not possible with usual models based on residual circulations.

The computed suspended matter distributions agree with the field observations, e.g. the oscillations in suspended matter concentrations, due to tidal oscillations, decrease when moving towards the northern border of the grid. Moreover, the computed suspended matter concentration in the northern border is very similar to the measured value. The sedimentation rates have shown that there is a net sedimentation in the river, although the process is slow.

Acknowledgements

This work was partially supported by EN-RESA and the Spanish DGICYT (contract PB89-0621).

References

- Abril, J.M. and García-León, M., 1993a. A 2D 4-phase dispersion model for non-conservative radionuclides. 1. Conceptual and computational model. J. Environ. Radioact., 20: 71–88.
- Abril, J.M. and García-León, M., 1993b. A 2D 4-phase dispersion model for non-conservative radionuclides. 2. Two applications. J. Environ. Radioact., 20: 89–115.
- Abril, J.M. and García-León, M., 1994. Modelling the distribution of suspended matter and the sedimentation process in a marine environment. Ecol. Model., 71: 197–219.
- Belderson, R.H., 1964. Holocene sedimentation in the western half of the Irish sea. Mar. Geol., 2: 147-163.
- Borrego, J. and Pendón, J.G., 1988. Algunos ejemplos de influencia de los procesos antrópicos en el medio sedimentario: la ría de Huelva. Henares Rev. Geol., 2: 299–305 (in Spanish).
- Durrieu de Madron, X., Nyffeler, F., Balopoulos, E.T. and Chronis, G., 1992. Circulation and distribution of suspended matter in the Sporades Basin (northwester Aegean Sea). J. Mar. Syst., 3: 237–248.
- Eisma, D., 1993. Suspended Matter in the Aquatic Environment. Springer-Verlag, Berlin.
- Guburt, P.A., Kershaw, P.J. and Durance, J.A., 1987. Modelling the distribution of soluble and particle-absorbed

radionuclides in the Irish sea. In: J.C. Guary, P. Guegueniat and R.J. Penthreath (Editors), Radionuclides. A Tool for Oceanography. Elsevier, London, pp. 395–407.

- Martínez-Aguirre, A., García-León, M. and Ivanovich, M., 1994. U and Th distribution in solution and suspended matter from rivers affected by the phosphate rock processing in southwestern Spain. Nucl. Instr. Meth. Phys. Res. A, 339: 287-293.
- Mehta, A.J., 1989. On estuarine cohesive sediment suspension behaviour. J. Geophys. Res., 94 (C10): 14303-14314.
- Periáñez, R. and García-León, M., 1993. Ra-isotopes around a phosphate fertlizer complex in an estuarine system at the southwest of Spain. J. Radioanal. Nucl. Chem., 172: 71-79.
- Periáñez, R., García-León, M. and Abril, J.M., 1994a. Radium isotopes in suspended matter in an estuary system in the southwest of Spain. J. Radioanal. Nucl. Chem., 183: 395-407.
- Periáñez, R., Abril, J.M. and García-León, M., 1994b. A modelling study of ²²⁶Ra dispersion in an estuarine system in southwest Spain. J. Environ. Radioact., 24: 159–179.
- Periáñez, R., Abril, J.M. and García-León, M., 1994c. ²²⁶Ra concentrations in the Odiel river: experimental results and a 2-D modelling study. In: M. García-León and R. García-Tenorio (Editors), Low Level Measurements of Radioactivity in the Environment: Techniques and Applications. World Scientific, Singapore.
- Periáñez, R., Abril, J.M. and García-León, M., 1994d. Formulación y desarrollo de un modelo matemático de un sistema estuario. Aplicaciones. In: A. Valle and C. Parés (Editors), Modelado de Sistemas en Climatología, Oceanografía y Ciencias Medioambientales: Aspectos Matemáticos y Numéricos. Imagraf, Málaga (in Spanish).
- Prandle, D., 1984. A modelling study of the mixing of ¹³⁷Cs in the seas of the European continental shelf. Phil. Trans. R. Soc. Lond. A, 310: 407–436.
- Pugh, D.T., 1987. Tides, Surges and Mean Sea Level. John Wiley and Sons, Chichester.
- Respaldiza, M.A., López-Tarrida, A.J. and Gómez-Camacho, J., 1993. Environmental control of Tinto and Odiel river basins by PIXE. Nucl. Instr. Meth. Phys. Res. B, 75: 334–337.
- Universidad de Sevilla, 1991. Contract with ENRESA, Final Report (in Spanish).